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Additional Information

1 Microalgae population dynamics growth with AnMBR effluent:
2 effect of light and phosphorous concentration.

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5

6 **Abstract**

7 The aim of this study was to evaluate the effect of light intensity and phosphorus
8 concentration on biomass growth and nutrient removal in a microalgae culture and
9 their effect on their competition. The photobioreactor was continuously fed with the
10 effluent from an AnMBR pilot plant treating real wastewater. Four experimental
11 periods were carried out at different light intensities (36 and 52 $\mu\text{mol s}^{-1} \text{m}^{-2}$) and
12 phosphorous concentrations (around 6 and 15 mgP L^{-1}). Four green algae:
13 *Scenedesmus*, *Chlorella*, *Monoraphidium* and *Chlamydomonas* and cyanobacterium
14 were detected and quantified along whole experimental period. *Chlorella* was the
15 dominant specie when light intensity was at the lower level tested, being competitively
16 displaced by a mixed culture of *Scenedesmus* and *Monoraphidium* when light was
17 increased. When phosphorous concentration in the photobioreactor was raised up to
18 15 $\text{mgP}\cdot\text{L}^{-1}$, a growth of cyanobacterium became the dominant specie in the culture.
19 The highest nutrient removal efficiency (around $58.4 \pm 15.8 \%$ and $96.1 \pm 16.5 \%$ of
20 nitrogen and phosphorous respectively) was achieved at 52 $\mu\text{mol s}^{-1} \text{m}^{-2}$ of light
21 intensity and 6.02 $\text{mgP}\cdot\text{L}^{-1}$ of phosphorous concentration, reaching about $674 \pm 86 \text{ mg}$
22 L^{-1} of VSS. The results obtained reveal how the light intensity supplied and the
23 phosphorous concentration available are relevant operational factors that determinate
24 the microalgae specie that is able to predominate in a culture. Moreover, changes in

25 microalgae predominance can be induced by changes in the growth medium produced
26 by the own predominant species.

27 **Keywords**

28 Anaerobic membrane bioreactor; continuous photobioreactor; microalgae
29 competition; nutrient removal; wastewater.

30

31 **INTRODUCTION**

32 The cultivation of microalgae has recently attracted growing interest as a solution for
33 tertiary wastewater treatment. This interest is based on several potential benefits as: i)
34 the simultaneous removal of nitrogen and phosphorus compounds in wastewater with
35 a lower sludge generation than in conventional treatments; ii) the use of CO₂ as carbon
36 source, contributing to reduce greenhouse gas emissions and iii) the generation of a
37 valuable biomass which may be used to produce biogas (Collet & Hélias, 2011),
38 manufacture biofuels (Chisti, 2007) or improve the energetic balance by direct
39 combustion of algae biomass (Sturm & Lamer, 2011).

40 In addition, Anaerobic Membrane Bioreactors (AnMBRs) technology has been
41 presented as a treatment able to reach high removal of Total Suspended Solids (TSS)
42 and Chemical Oxygen Demand (COD), producing a high quality effluent while improve
43 the energetic balance through a generation of biomethane and a lower production of
44 sludge (Bornare *et al.*, 2015; Giménez *et al.*, 2011). However, this technology is not
45 able to remove inorganic nutrients efficiently. Therefore, when the effluent is
46 discharged into aquatic environments could cause important eutrophication problems.

47 Coupling the AnMBR technology with microalgae cultivation can benefit from all the
48 above mentioned advantages, thus, being an interesting technology for wastewater
49 treatment.

50 Up to now, very few studies have demonstrated the feasibility of a microalgae post-
51 treatment for the effluent produced by AnMBRs technology (Ruiz-Martinez *et al.*,

52 2012). The main challenge still persists and is to obtain a stable microalgae culture able
53 to reduce nitrogen and phosphorus concentration to values below the discharge limits
54 established in Council Directive 91/271/EEC.

55 Nutrient removal by microalgae is influenced by many factors: physical, such as light,
56 nutrient concentration, pH or temperature (Richmond, 2004), as well as biological,
57 such as competition between bacteria and microalgae or between different species of
58 microalgae.

59 Light is the most relevant parameter in microalgae growth (Jonker & Faaij, 2013). It has
60 to be supplied at the optimum intensity, duration and wavelength to reach the
61 maximum algal growth and nutrient removal efficiency (Termini *et al.*, 2011).
62 Moreover, light can also determine which phytoplankton can proliferate in the culture.
63 Hence, the predominant microalgae species determine nutrient removal.

64 At one extreme: no nutrient limitation culture, microalgae usually compete for light.
65 Light that has not been absorbed by microalgae reaches the bottom of the water
66 column with intensity I_{out} (Huisman *et al.*, 1999). Hence, I_{out} is variable as a function of
67 the microalgae growth. Therefore, the critical light intensity (I_{critic}) of a specie is defined
68 as the light intensity registered at the bottom of a well-mixed water column at which
69 this species can just survive (Passarge *et al.*, 2006). In a constant and well-mixed
70 environment, theory predicts that the species with the lowest I_{critic} will be the superior
71 competitor for light (Huisman & Weissing, 1994). Experiments reported by Huisman *et al.*
72 (1999) and Litchman (2003) with phytoplankton in light limited conditions support
73 this prediction.

74 Nutrient concentration can also determine the phytoplankton which can survive in the
75 culture. Thus, at the other extreme: no light limitation culture, in a constant and well-
76 mixed environment, the species with lowest nutrient requirements will be the superior
77 nutrient competitor (Passarge *et al.*, 2006). This prediction has been upheld by
78 numerous nutrient competition studies (e.g., Van Donk & Kilham, 1990; Ducobu *et al.*,
79 1998; Passarge *et al.*, 2006).

80 Nevertheless, the abovementioned studies have been focused on the competition of
81 species in batch conditions. So the studies focused in the effect of this competitions in

82 a continuous culture are very scarce (e.g., Pisman, 2002), feeding the culture with
83 synthetic water, without the inherent variability associate to the real influents. For this
84 reason, in this work different experimental conditions are tested in order to assess the
85 possibility to remove nutrients (meeting legal requirements) from the effluent of a
86 pilot plant AnMBR (processing real wastewater) with microalgae. For this purpose, is
87 essential to analyze the microalgae population dynamics using real AnMBR effluent to
88 ensure the accomplishment of discharge limits established.

89 Therefore, the aim of this study is to analyze the effect of light intensity and nutrient
90 concentration on growth, nutrient removal efficiency and species competition in an
91 indigenous microalgae culture fed by AnMBR effluent which treated real urban
92 wastewater.

93

94 **MATERIAL AND METHODS**

95 *Inoculum*

96 The microalgae used as inoculum in this study came from the photobioreactors pilot
97 plants located in Carraixet WWTP (Valencia, Spain) and owned by the CALAGUA
98 research team. This inoculum was initially composed by *Monoraphidium* and
99 *Scenedesmus* with a relative abundance of 73 % and 27 % respectively.

100

101 *Culture medium*

102 The fresh culture medium fed into the lab-scale photobioreactor (LabPBR) was
103 obtained from the effluent of the Submerged anaerobic membrane bioreactor pilot
104 plant (AnMBR) located in Carraixet WWTP (Valencia, Spain) and owned by the
105 CALAGUA research team. This pilot plant is feed with the effluent of the pre-treatments
106 units of the Carraixet WWTP (a full-scale urban wastewater treatment plant that treats
107 131050 PE). Further details about AnMBR process can be found in previous studies
108 (Robles *et al.*, 2015; Giménez *et al.*, 2011).

109 To feed the LabPBR, the AnMBR effluent was collected in opaque glass bottles and
110 taken to laboratory every three days. In order to prevent the proliferation of
111 microorganisms in the collected effluent, it was kept in the dark at a temperature of 5
112 °C. The average main composition of AnMBR effluent is shown in Table 1.

113 **Table 1.** Average AnMBR effluent composition.

Parameter	Mean ± SD
pH	7.29 ± 0.10
COD (mg COD L ⁻¹)	58.6 ± 10.2
BOD _L (mg BOD L ⁻¹)	26 ± 9
VFA (mg COD L ⁻¹)	2.0 ± 0.3
Alk (mg CaCO ₃ L ⁻¹)	817.24 ± 22.56
NO ₂ -N (mg N L ⁻¹)	0.37 ± 0.18
NO ₃ -N (mg N L ⁻¹)	1.42 ± 0.67

114

115 *Lab-scale Photobioreactor operation*

116 The LabPBR consisted of a cylindrical clear tank with 19 cm of internal diameter (9 L
117 working volume) (See Fig. 1a), installed in an incubator chamber with temperature
118 control.

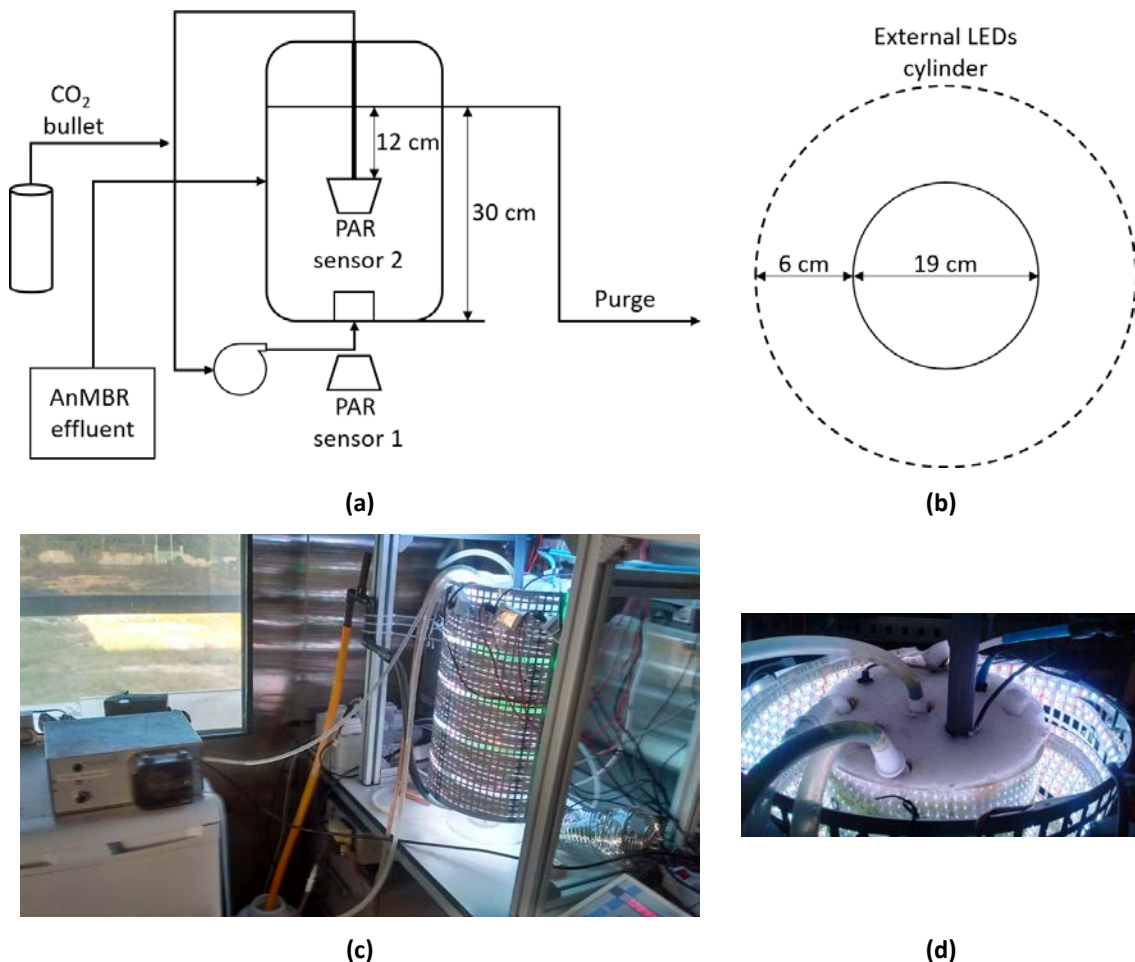
119 In order to achieve appropriate homogenization and maintain the pH fixed at 7.5, the
120 culture was agitated with air across four fine bubble diffusers positioned crosswise on
121 the bottom and pure CO₂ (99.9%) was injected into the gas flow from a gas bullet.

122 The temperature and pH were monitored online and logged on a PC through a self-
123 made data acquisition software, processing the signal by a multiparametric analyzer
124 (CONSORT C832, Belgium).

125 During the start-up, the LabPBR was operated in batch mode for 3 days, and then, was
126 fed with the nutrient-loaded effluent from AnMBR system, and was operated in a
127 semi-continuous mode, without retention of biomass. A peristaltic pump controlled by
128 a PC was used to feed every three hours (8 feed cycles a day) a flow of 280 mL in order

129 to keep constant the SRT in 4 days, maintaining the total volume with an overflow at
130 the top of the reactor. Moreover, allylthiourea was feed to inhibit nitrification bacteria
131 growth, thus assuring that nitrogen removal was due to the microalgae activity.

132 The illumination was provided by eight led strips (Efecto LED, SMD5050 60LED/M 5M
133 RGB IP65) fixed in an external cylinder (31 cm internal diameter) around the LabPBR
134 (see Fig. 1b). Lighting was supplied 24 hours a day, and two photo sensors (Sensor PAR
135 Apogee SQ-222) were disposed under and inside of the LabPBR in order to measure
136 light intensity supplied to the culture (see Fig. 1a). PAR sensor 1 was used to
137 determinate and control the light intensity supplied to the photobioreactor while PAR
138 sensor 2 was used to determinate the value of I_{out} as a function of the microalgae
139 concentration.



140 **Figure 1.** Experimental set-up: Scheme of the (a) laboratory scale photobioreactor and (b) Illumination
141 system. Photo of the (c) laboratory scale photobioreactor and (d) Illumination system.

142

143 *Experimental design*

144 The study aimed at assessing the influence on the microalgae culture of two variables:
145 light supplied on the reactor surface (at two different intensities: 36.3 and 52.2 $\mu\text{mol s}^{-1}$
146 m^{-2}) and phosphorous concentration feed (around 6.02 and 15.23 mgP L^{-1}), keeping
147 constant all others operational conditions. Four experimental periods were conducted.
148 Table 2 displays the operational conditions imposed at each experimental period.

149 Since the concentration of nitrite and nitrate were negligible in the AnMBR effluent
150 and the continuous addition of CO_2 to the lab-PBR, only light, ammonium and
151 phosphate concentrations were considered as limitations to the microalgae growth.

152

153 **Table 2.** Operational conditions of the lab-scale photobioreactor during each experimental period.

Period	Duration (d)	Light intensity ($\mu\text{mol s}^{-1} \text{m}^{-2}$)	Ammonium ($\text{mg NH}_4\text{-N L}^{-1}$)	Phosphorus ($\text{mg PO}_4^{3-}\text{-P L}^{-1}$)	Temperature ($^{\circ}\text{C}$)
Exp. 1	50	36.3 \pm 5.3	60.62 \pm 2.81	6.05 \pm 0.73	27.4 \pm 0.8
Exp. 2	54	52.2 \pm 4.8	59.31 \pm 6.64	5.95 \pm 1.22	27.7 \pm 1.0
Exp. 3	9	52.1 \pm 0.9	59.32 \pm 0.10	15.23 \pm 0.03	28.1 \pm 0.4
Exp. 4	16	52.2 \pm 1.7	66.06 \pm 4.02	7.73 \pm 0.23	27.5 \pm 0.8

154

155 *Analytical methods*

156 Nutrient recovery by microalgae was assessed by recording thrice a week nitrogen and
157 phosphate concentration in both, the influent and the soluble fraction collected from
158 the lab-scale PBR purge. This soluble fraction was obtained by membrane filtration.
159 There were used 0.45 mm pore size filters of polycarbonate glass fiber.

160 Total and volatile suspended solids (TSS and VSS) were determined thrice a week to
161 evaluate biomass growth under each experimental period.

162 The nitrogen and phosphorous content of the dry biomass were measured in triplicate
163 once every fifteen days. For this determination an acid-digestion of the dry biomass
164 was performed.

165 Solids, phosphorous biomass content and all nutrients (ammonium, nitrate and
166 phosphate) were obtained according to *Standard Methods for the Examination of*
167 *Water and Wastewater* (APHA, AWWA, WEF, 2012). These methods were
168 implemented in a multiparametric analyzer (*Smartchem200 de AMS/Alliance*
169 *Instruments*). Nitrogen biomass content was determined via espectrofotometric
170 method using commercial kit (MERCK, 100613) (Spectroquant® Pharo 300 MERCK).

171

172 *Microbiological method*

173 To assess the microalgae community evolution, twice a week a cell count was
174 performed. A sample of 50 µL was filtered through 0.2 µm membranes. In order to
175 eliminate the retained salt, the filters were washed using distilled water and then,
176 dehydrated through successive washes with ethanol (50%, 80%, 90% and 99%). Cell
177 counts were accomplished by the 100x oil immersion lens of an epifluorescence
178 microscopy on a Leica DM2500. In the cell counts, a minimum of 300 cells were
179 counted, assuring that were counted at last 100 cells of the most abundant genera
180 with an error of less than 15% (Pachés *et al.*, 2012). All the measurements were
181 obtained in triplicate.

182

183 *Calculations*

184 Nutrient removal efficiency was calculated considering influent and effluent terms in a
185 daily balance basis.

186 In the nitrogen balance, only NH₄ was considered to be available for biomass growth.
187 This assumption was made based on the concentration of the other soluble species
188 (NO₃ and NO₂) were negligible (below 2.20 mg N L⁻¹). Likewise, nitrification was not
189 considered since allylthiourea was used to inhibit the nitrifying bacteria growth.
190 Nitrogen gas loss (N₂ or NH₃) was not considered since the pH was kept always around
191 7.5 (at this pH value, the predominant form of ammonia nitrogen is by far NH₄).

192 In the phosphorous balance, phosphorous precipitation was assumed to be negligible
193 due to the low solubility of the possible precipitating compounds (as struvite) in water
194 at neutrality (Laliberte *et al.*, 1997).

195 Therefore, nutrient removal efficiency (NRE) was calculated as follows:

$$196 \quad NRE (\%) = \left(1 - \frac{E}{I}\right) \cdot 100 \quad (eq. 1)$$

197 Where I and E are the ammonium or phosphate concentration in the influent and
198 effluent respectively (mg L^{-1}).

199 Likewise, intracellular nutrients concentration (INC) were calculated as follows:

$$200 \quad INC (\%) = \left(\frac{T - E}{VSS}\right) \cdot 100 \quad (eq. 2)$$

201 Where T is the total nitrogen or phosphorous concentration in the purge.

202 Moreover, the N/P elimination and intracellular ratios (N/P_E and N/P_I respectively)
203 were calculated in order to assess the different nutrients needs of each microalgae
204 covered.

$$205 \quad N/P_E = \frac{I_{NH_4} - E_{NH_4}}{I_{PO_4} - E_{PO_4}} \quad (eq. 3)$$

$$206 \quad N/P_I = \frac{T_{NH_4} - E_{NH_4}}{T_{PO_4} - E_{PO_4}} \quad (eq. 4)$$

207 Finally, nutrients normalized uptake (NNU) was determined on a daily basis through
208 the following equation:

$$209 \quad NNU = \frac{I - E}{CC} \quad (ec. 5)$$

210 Where CC is the archived value of cell counts.

211

212

213

214

215

216 RESULTS AND DISCUSSION

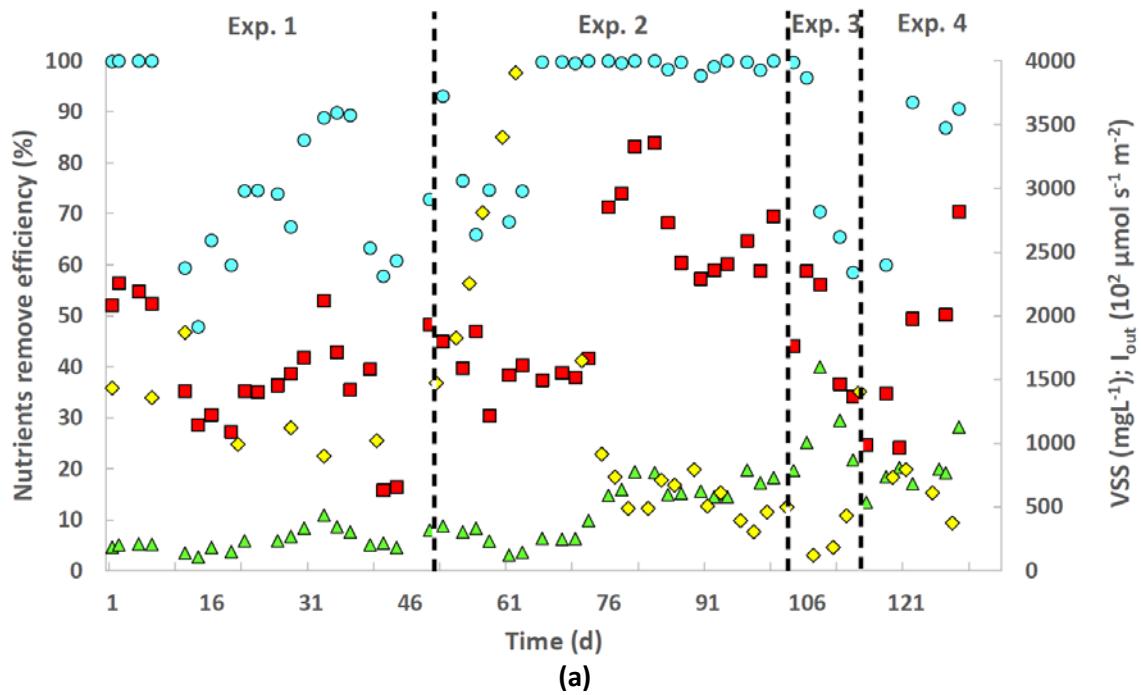
217 *Effect of light*

218 The first two experimental periods (Exp. 1 and Exp. 2) were aimed at assess the
219 growth, nutrient removal efficiency and species competition under two different light
220 intensities (36.3 ± 5.3 and $52.2 \pm 6.1 \mu\text{mol s}^{-1} \text{m}^{-2}$), maintaining constant all others
221 working conditions. Figure 2 shows the time profile evolution of nutrient removal, light
222 intensity, volatile suspended solids and relative microalgae species abundance
223 obtained along the experimental period.

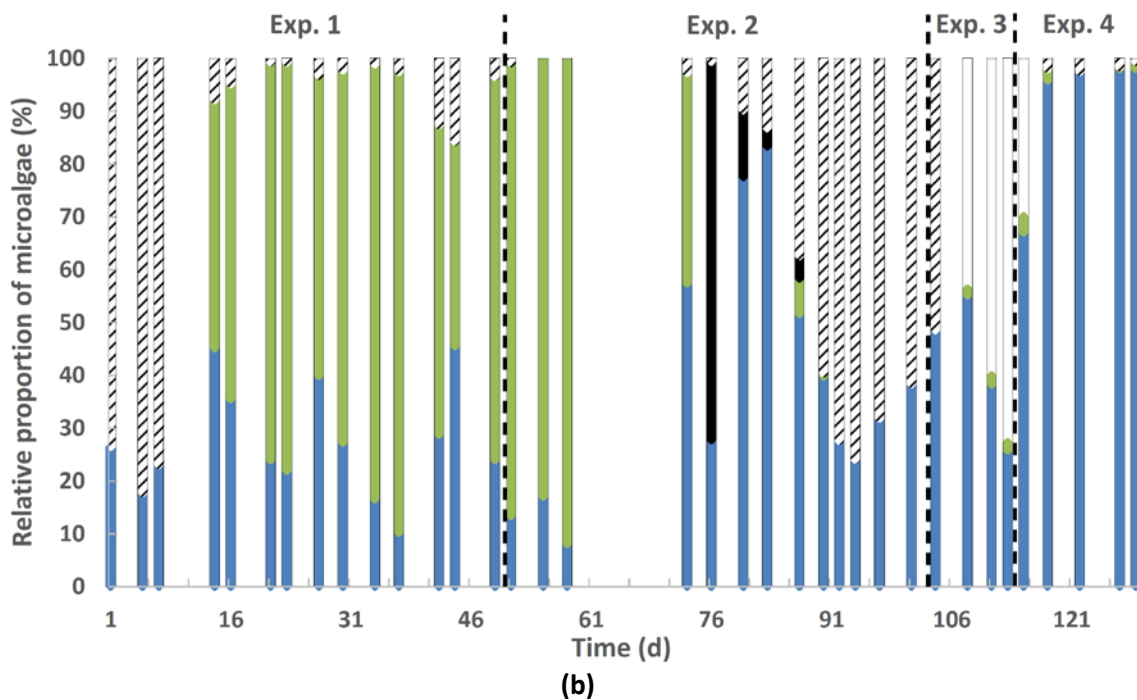
224 The first experimental period lasted 50 days and the volatile suspended solid
225 concentration reached about $244 \pm 83 \text{ mg L}^{-1}$ of VSS. After a week, the genres included
226 in the original inoculum (*Scenedesmus* and mostly *Monoraphidium*) were competitively
227 displaced by genre *Chlorella*, becoming the predominant microalgae in the culture. As
228 can be seen in Figure 2b this genre not reached a stable value of relative abundance,
229 despite being the predominant genus. It can also be observed a decrease in the daily
230 ammonium and phosphorus removal efficiency with this specie change, achieving
231 values about $52.8 \pm 5.7 \%$ and $99.9 \pm 0.1 \%$ of nitrogen and phosphorous respectively
232 when *Monoraphidium* was the dominant specie versus the $35.6 \pm 9.9 \%$ and $71.8 \pm$
233 13.3% of nitrogen and phosphorous removal when *Chlorella* was dominant.

234 However, when light intensity was raised from 36 to $52 \mu\text{mol s}^{-1} \text{m}^{-2}$, green algae
235 *Chlorella* was competitively displaced by both, *Scenedesmus* and *Monoraphidium*. In
236 this second experimental period (54 days of duration), VSS increased until $674 \pm 86 \text{ mg}$
237 L^{-1} . Specie abundance stability was not reached in this experimental period. During this
238 second period, the daily ammonium and phosphorus removal efficiency increased until
239 reaching values about $58.4 \pm 15.8 \%$ and $96.1 \pm 16.5 \%$, respectively.

240 The increase in the volatile suspended solids could indicate that in Exp. 1, the most
241 important biomass growth limiting factor was light. When no other factor is limiting
242 the microalgae growth, enhance light intensity speed up the microalgae metabolism as
243 long as it stays under the optimum value (Martín and Marzal, 1999).



244
245



246
247

248 **Figure 2.** Evolution of (a) nutrient removal efficiency (nitrogen ■ and phosphorous ●), volatile
249 suspended solids (▲), $I_{out} \times 10^2$ (◆) and (b) relative abundance of microalgae (*Chlorella* ■, *Scenedesmus*
250 ■, *Monoraphidium* ▨, *Chlamydomonas* ■ and *Cyanobacterium* □) in the LabPBR during each
251 experimental period. Vertical black dotted lines indicate an experimental period change.

252 Moreover, this increase in the light supplied to the reactor surface allowed
253 *Scenedesmus* and *Monoraphidium* genres to be more competitive than *Chlorella* genre.
254 This result suggests that *Scenedesmus* and *Monoraphidium* genres growth requires
255 noticeably higher I_{out} than the *Chlorella* genre. This conclusion is in agreement with

256 that reported by Huisman *et al.* (1999), who reported that *Scenedesmus* had a much
257 higher critical light intensity than *Chlorella*, being competitively excluded under
258 deficient light conditions. In the same way, Passarge *et al.* (2006) reported that in two
259 pure microalgae culture growths, *Monoraphidium* showed higher I_{critic} than *Chlorella*.

260 Therefore, the lesser competitiveness of *Scenedesmus* and *Monoraphidium* genres in the
261 first experimental period could be explained by the low light intensity supplied.
262 Although a given microalgae specie can proliferate whenever the actual value of I_{out} is
263 above their critical light intensity (Huisman and Weissing, 1994; Weissing and
264 Huisman, 1994), in competition, the species with the lowest I_{critic} displace all others
265 species. This observation is due to the fact that, during its growth, the species with the
266 lowest I_{critic} is able to reduce the light penetration to the bottom of the reactor below
267 the critical light intensities of all others species (Huisman and Weissing, 1994; Weissing
268 and Huisman, 1994).

269 Consequently, since light usually represents the limiting factor in the cultures of
270 photosynthetic microalgae (Cuaresma *et al.*, 2011), and influences their competence
271 (Passarge *et al.*, 2006; Huisman *et al.*, 1999), it is imperative to be able to estimate the
272 value of light reaching the center of a photobioreactor in relation with the VSS in order
273 to supply the accurate light intensity.

274 Commonly this relation is estimated by the Lambert-Beer expression:

$$275 \quad I_{out} = I_0 \cdot \exp(-k_e \cdot c_b \cdot z) \quad (eq. 6)$$

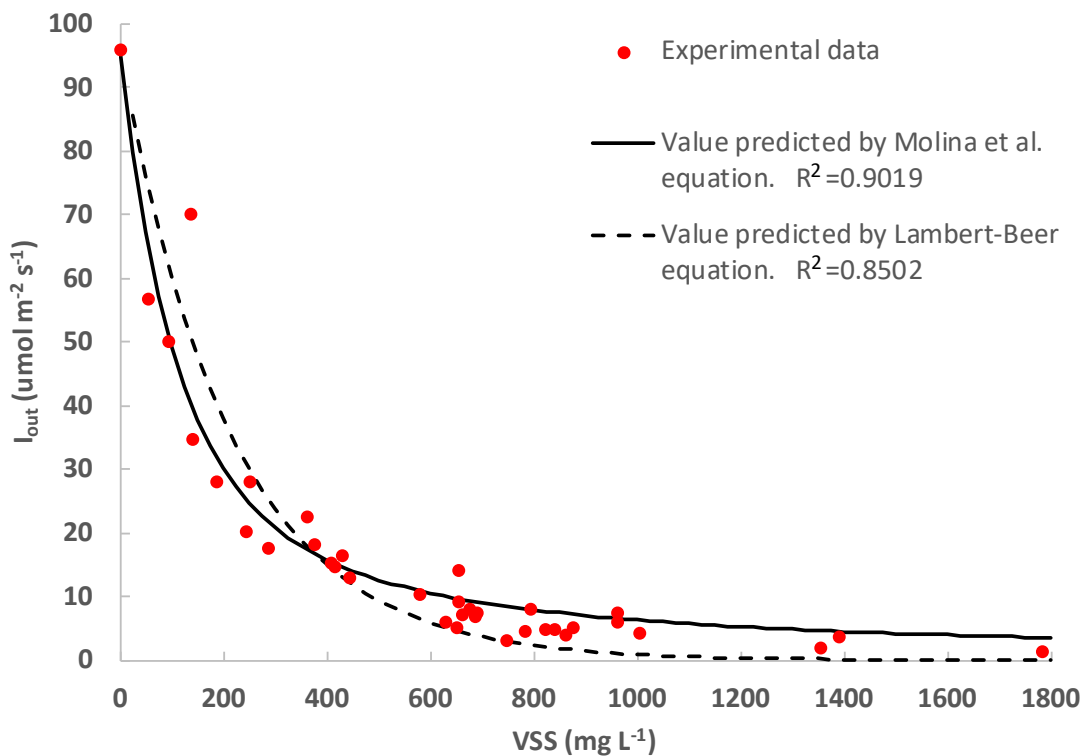
276 Where I_0 is the intensity measured at the surface of the water column ($\mu\text{mol s}^{-1} \text{m}^{-2}$), k_e
277 is the light attenuation coefficient ($\text{m}^2 \text{Kg}^{-1}$), c_b is the solid concentration (Kg m^{-3}) and z
278 is the thickness of the water column (m).

279 Since equation 6 was developed in plane coordinates, it does not adjust well enough to
280 the results obtained in a cylindrical reactor illuminated from all its perimeter. Then, the
281 relation between VSS and I_{out} in a cylindrical reactor can be adjusted in a better way by
282 the equation proposed by Molina *et al.* (1997):

283
$$I_{out} = \frac{I_0}{k_e \cdot c_b \cdot L_{eq}} (1 - \exp(-k_e \cdot c_b \cdot L_{eq})) \text{ (eq.7)}$$

284 Where L_{eq} is the equivalent optical length of the system, which is a function of the
 285 radius and in the most usual conditions takes the value of $L_{eq} = 1.60 \cdot r$, where r is the
 286 radius of the cylinder object of study (9.5 cm in the conditions of this study).

287 Figure 3 shows the experimental I_{out} values measured for each volatile suspended solid
 288 concentration and the fitting provided by both, Lambert-Beer and Molina *et al.*
 289 equations. As can be seen in this Figure, the Molina *et al.* equation provided a much
 290 better fit.



291 **Figure 3.** Fitting the experimental I_{out} values registered with an external light intensity on the surface's
 292 photobioreactor of $52 \mu\text{mol s}^{-1}\text{m}^{-2}$ using the equations of Molina *et al.* (1997) and Lambert-Beer.
 293

294 Moreover, from this equation fitted it can be deduced the k_e coefficient, which can
 295 provide an interesting information by the fact that represents the efficiency with which
 296 light can be harnessed. The higher the k_e coefficient, the higher the amount of supplied
 297 light will be needed to reach the center of the photobioreactor.

298 Usually, k_e depends mostly on the genre and conditions of the algal culture, due that
299 the mainly light that is supplied to the photobioreactor is absorbed by microalgae.
300 However, the thickness, geometry and material of the photobioreactor must be taken
301 in account due that it represents an additional resistance to the light passage.

302 In this study, it has been deduced the k_e from the Molina *et al.* equation, achieving a
303 value of $0.0859 \text{ m}^2 \text{ gTSS}^{-1}$. This value has been calculated assuming a VSS/TSS relation
304 of 86 %, which has been deduced from the values obtained in this study.

305 The k_e obtained is similar to that reported by other authors (Ruiz-Martínez *et al.*, 2016;
306 Molina-Grima *et al.*, 1994), specially by that reported by Ruiz-Martínez *et al.* (2016),
307 whom assumed a value of $0.0758 \text{ m}^2 \text{ gTSS}^{-1}$ operating a flat-plate photobioreactor in
308 outdoors conditions at similar TSS concentrations. Consequently, it can be deduced
309 that the effects of photobioreactor resistance to the light passage on the k_e coefficient
310 can be often considered negligible, being in accord with Molina-Grima *et al.* (1994)
311 whom reported that k_e depends mainly on the algal light absorption.

312 It must be highlighted that, despite microalgae genre was changed in each
313 experimental period (with inherent variability of size and shape), no significant
314 difference in the provided auto-shadow was observed. Thus, these equations could be
315 applicate to estimate the I_{out} of any microalgae culture, pure or in consortium.

316 On the other hand, during the second experimental period (Exp. 2) the value of I_{out}
317 decreased due to the increase in the VSS, reaching values lower than those registered
318 in Exp. 1 (see Fig. 2a), but no significant presence of *Chlorella* was observed.
319 Consequently, it can be concluded that, although the increase in the supplied light to
320 the reactor surface improved the *Scenedesmus* and *Monoraphidium* competition, the
321 dominance of the culture by these genres was not only due to light intensity supplied.

322

323 *Effect of nutrients concentration*

324 Another important operational factor with strong influence on the competition
325 between microalgae species is the nutrient concentration (Yang *et al.*, 2016). In Exp. 1,

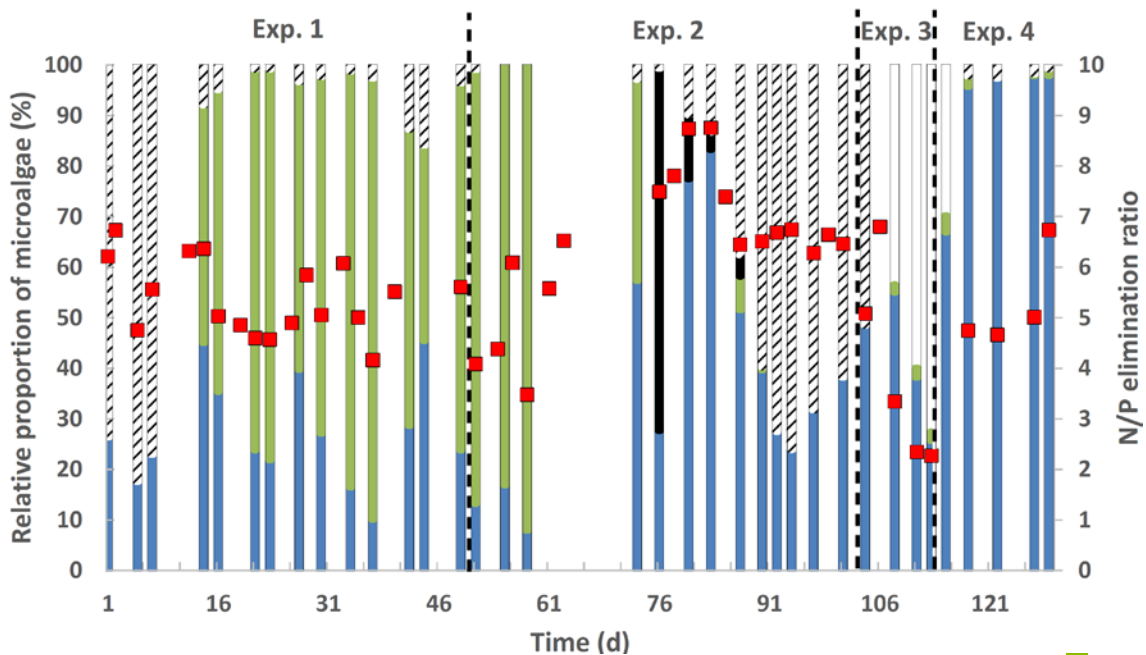
326 nutrient removal efficiency was low enough (see Fig. 2a) to maintain a nutrient
327 concentration in the reactor high enough to not limit the microalgae growth. However,
328 in Exp. 2, the nutrient removal efficiency increased reaching high values, especially for
329 the phosphorus concentration (up to 96% of P-removal). Therefore, nutrient
330 concentration was low enough to be considered also a growth limiting factor.

331 Then, it can be assumed that the increase in nutrient removal was produced by the
332 increment in the external light on the photobioreactor's surface. Nevertheless, this low
333 phosphorus concentration in the reactor could have favoured the *Scenedesmus* and
334 *Monoraphidium* genres competition against *Chlorella* since according to Wu *et al.*
335 (2014), *Chlorella vulgaris* can live at the lower nitrogen concentration, but it is very
336 difficult to survive in the absence of phosphorus. Therefore, phosphorus is the limiting
337 factor for its growth.

338 The N/P elimination ratio showed while *Chlorella* was the predominant genus in the
339 culture, was notably lower than that displayed by the consortium formed by
340 *Scenedesmus* and *Monoraphidium* (5.0 ± 0.8 and 6.8 ± 0.3 respectively, see Fig. 4). The
341 intracellular nitrogen and phosphorous content achieved (see Table 3), seems to
342 indicate that *Chlorella* need higher phosphorus concentration for its growth than
343 *Scenedesmus* and *Monoraphidium*. Consequently, this result suggests that *Chlorella*
344 has more dependence on phosphorous concentration than other green algae like
345 *Scenedesmus* or *Monoraphidium*.

346 Considering the strong influence that the concentration of phosphorus has on the
347 competence between microalgae species as evidenced in this study, Exp. 3 and Exp. 4
348 were planned to confirm how the variations in the phosphorous concentration are able
349 to change the microalgae dynamics in the LabPBR. Phosphorous concentration was
350 raised to 15 mg L^{-1} in Exp. 3 and then were reduced to normal concentrations in Exp. 4
351 (see Table 2). The increase in P-concentration led to an important cyanobacterium
352 growth (until a 72.6 % of relative abundance) and decreasing until its vanishment when
353 the phosphorous concentration was lowered to the typical concentration level
354 recorded in the AnMBR effluent (see Fig. 2). This phenomenon can be explained in the
355 same way that *Chlorella* competition, having cyanobacterium higher phosphorous

356 dependence. This result is in accordance too with the N/P elimination and intracellular
 357 contents ratios obtained when cyanobacterium was the predominant microalgae in
 358 the culture (see Fig. 4 and Table 3), achieving the lower N/P elimination and
 359 intracellular content ratios among all others genres that predominated along each
 360 experimental period.



361 **Figure 4.** Evolution of N/P elimination ratio (■) and relative microalgae abundance (*Chlorella* ■,
 362 *Scenedesmus* ■, *Monoraphidium* ▨, *Chlamydomonas* ■ and *Cyanobacterium* □) in the LabPBR along
 363 the experimental period.
 364

365 During the second period a rapid growth of *Chlamydomonas* genre was observed (day
 366 76 in Figure 4), which vanished after a few days (day 87 in Figure 4). The
 367 *Chlamydomonas* growth seems to be lightly related with the available ammonium
 368 concentration in the PBR, just like *Chlorella* and cyanobacterium with the phosphate
 369 concentration, showing a significant improvement in the ammonium removal
 370 efficiency ($73.5 \pm 7.0 \%$), and registering the higher N/P elimination ratio (7.8 ± 0.9) of
 371 the whole experimental period when this genus was present in the PBR.

372 This result suggests that the biomass growth itself could be able to produce changes in
 373 the medium that influence microalgae competition since in their growth, light
 374 availability and nutrient concentration will decrease until microalgae culture reaches
 375 the equilibrium (i.e., the pseudo-steady state). Then, although the ideal conditions for
 376 one specie were achieved in a photobioreactor, the culture evolves until reaching its

377 own equilibrium. Further research is being carried to provide additional data
378 confirming this interesting finding.

379 Moreover, in regard to the effluent quality produced, it can be seen in Figure 2 that in
380 the second period (52 $\mu\text{mol s}^{-1} \text{m}^{-2}$ of light supplied), the microalgae culture was able
381 to remove practically all phosphorous concentration. Nevertheless, it could not
382 remove enough the nitrogen concentration to meeting legal discharge limits, reaching
383 an ammonium concentration in the effluent about 25 $\text{mgNH}_4\text{-N L}^{-1}$. Thus, since that
384 microalgae stop their activity in the absence of any of required nutrients (Hoff & Snell,
385 2001), the microalgae cultivated in this study would not be able to treat AnMBR
386 effluent properly, requiring an additional process to reduce ammonium in the effluent.
387 However, AnMBR effluent treatment by microalgae have been studied by many
388 different authors (i.e. Ruiz-Martínez *et al.*, 2016; Viruela *et al.*, 2016), whom have
389 reported promising results operating with outdoors pilot-plants, showing the potential
390 of the microalgae as a feasible tertiary treatment for urbane wastewaters.

391

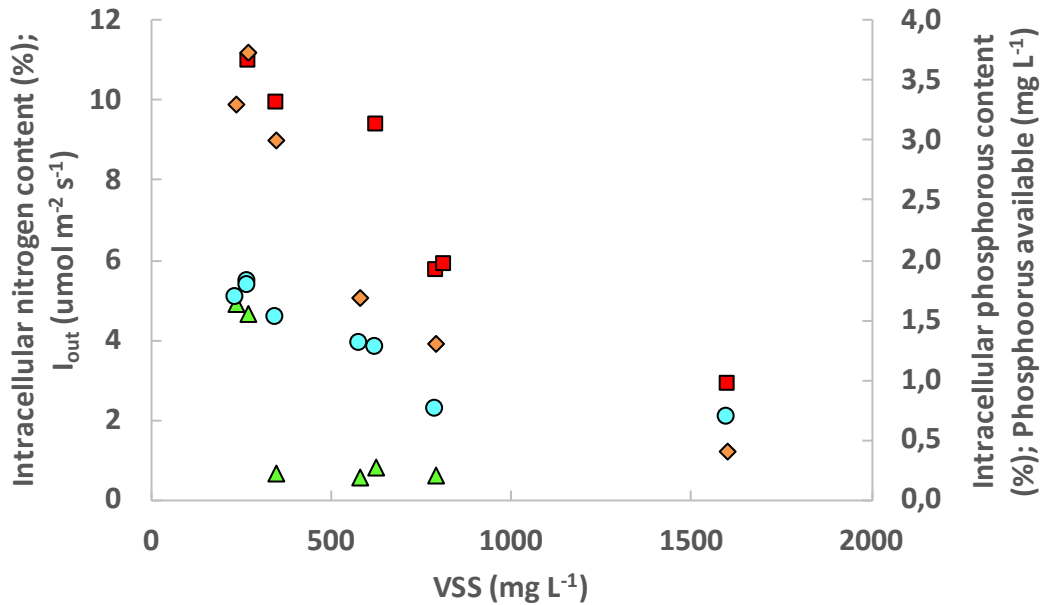
392 *Biomass composition and nutrient uptake efficiency*

393 Table 3 shows the nitrogen and phosphorous content in the biomass obtained by each
394 consortium of microalgae in each experimental period.

395 **Table 3.** Intracellular nutrient content obtained by each consortium of microalgae.

Day	Algae Population (%)				%N	%P	N/P ratio	VSS (mg L ⁻¹)	I ($\mu\text{mol s}^{-1} \text{m}^{-2}$)	I _{out} ($\mu\text{mol s}^{-1} \text{m}^{-2}$)
	<i>Scenedesmus sp.</i>	<i>Chlorella sp.</i>	<i>Monoraphidium sp.</i>	<i>Cyanobacterium</i>						
21	24	75	1	-	-	1.70	-	236	36	9.93
29	27	70	2	-	-	1.83	-	270	36	11.21
34	17	82	1	-	9.99	1.53	6.52	344	36	9.00
42	29	59	13	-	11.02	1.80	6.12	268	52	10.22
90	40	-	60	-	9.40	1.27	7.41	623	52	7.94
92	28	-	72	-	-	1.31	-	580	52	5.07
97	32	-	68	-	5.79	0.76	7.61	790	52	3.94
108	55	2	-	43	2.93	0.70	4.21	1600	52	1.22
120	96	2	2	-	5.93	-	-	810	52	7.31

396 Figure 5 shows a scatter plot of the intracellular nutrient content, I_{out} and PBR available
 397 phosphorus versus the Volatile Suspended Solids (VSS). As can be seen in this Figure,
 398 the nutrient content decreases with the VSS increase.



399

400 **Figure 5.** Intracellular nutrient content (nitrogen ■ and phosphorous ●), I_{out} (◆) and available
 401 phosphorous (▲) versus VSS concentration.

402 According to different authors, a reduction in the culture available light cause an
 403 increase on the biomass phosphorus content (Hessen *et al.*, 2002; Powell *et al.*, 2008;
 404 Ruiz *et al.*, 2014). It has been construed as a reduction in ATP accumulation when is
 405 available enough light energy. Additionally, Hessen *et al.* (2002) reported that high
 406 light intensity caused reductions in the biomass nitrogen content. Consequently, it can
 407 be assumed that supplying low light intensity to the culture, the nitrogen and
 408 phosphorous biomass content must be higher than that obtained at elevated light
 409 intensities.

410 However, from approximately 350 mg VSS L⁻¹ onwards (see Figure 5), the available
 411 phosphorous concentration began to be low enough to be considered a growth
 412 limiting factor, reaching values under 0.2 mg P L⁻¹. Then, it can be assumed that the
 413 decreasing in the biomass phosphorous content is due to the competition among
 414 microalgae species for the scarce available phosphorous concentration. In addition,
 415 nitrogen content in the microalgae biomass decreased also with the low available

416 phosphorous concentration although the available nitrogen concentration was high
 417 enough for not limiting the biomass growth (data no shown). This could be explained
 418 by the fact that microalgae require both nutrients from the environment, stopping
 419 their activity in the absence of any of them (Hoff & Snell, 2001). According Marcilhac *et*
 420 *al.* (2014), when phosphorous concentration was below 0.1 ppm, nitrogen uptake was
 421 limited.

422 The nitrogen and phosphorus content obtained in this study is similar to the contents
 423 reported by other authors in different species (see Table 4).

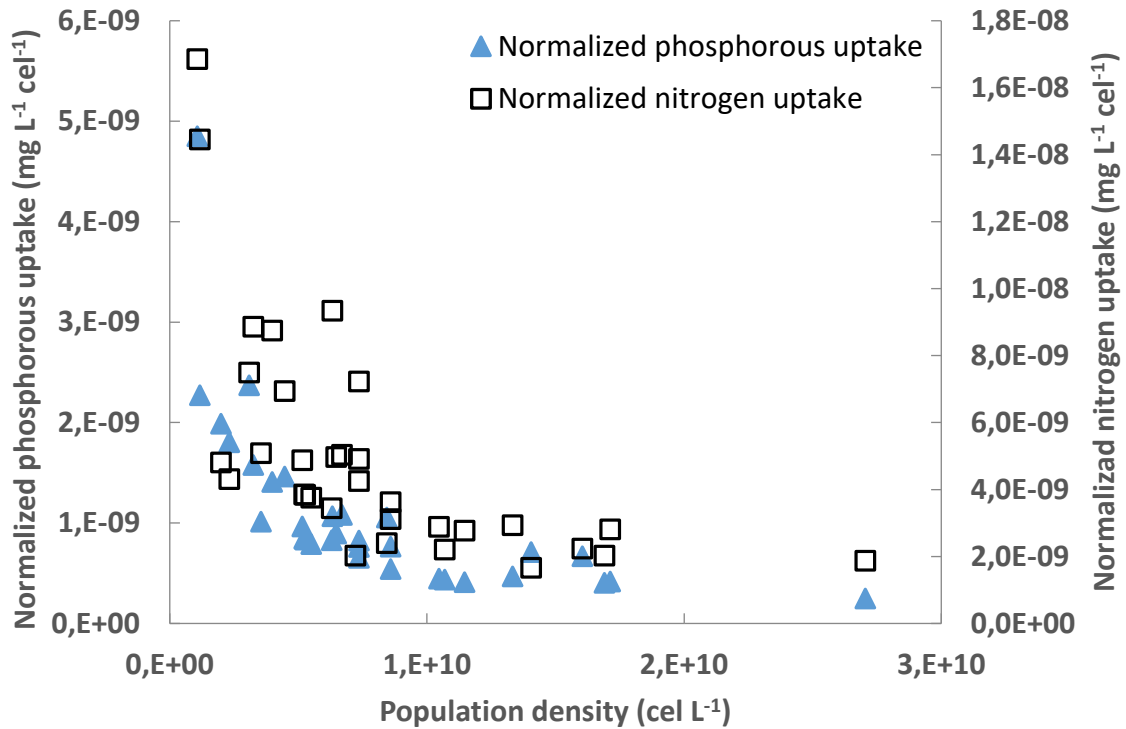
424 **Table 4.** Intracellular nutrient content reported by different authors.

Microalgae specie	% N	% P	N/P	Authors
<i>Chlorella sorokiniana</i>	10.10	-	-	Richardson <i>et al.</i> (1969)
<i>Scenedesmus Obliquus</i>	-	0.99	-	Martínez <i>et al.</i> (2000)
Consortium (<i>Scenedesmus</i> dominant)	-	3.16	-	Powell <i>et al.</i> (2008)
Consortium fed at low loading rate	6.55	0.94	6.97	Mulbry <i>et al.</i> (2008)
Consortium fed at high loading rate	5.45	0.82	6.65	Mulbry <i>et al.</i> (2008)
<i>Nanoochloropsis oculata</i>	8.30	-	-	Hsueh <i>et al.</i> (2009)
Consortium	9.27	0.87	10.66	Chinnasamy <i>et al.</i> (2010)
<i>Scenedesmus sp.</i>	-	3.50	-	Yin-Hu <i>et al.</i> (2012)
<i>Scenedesmus Obliquus</i> inoculated at low biomass	4.85	0.68	7.13	Ruiz <i>et al.</i> (2014)
<i>Scenedesmus Obliquus</i> inoculated at high biomass	5.89	0.78	7.55	Ruiz <i>et al.</i> (2014)
Consortium at low biomass (<i>Chlorella sp.</i> dominant)	9.99	1.53	6.52	This study
Consortium at medium biomass (<i>Scenedesmus sp./Monoraphidium</i> <i>sp.</i>)	5.79	0.76	7.61	This study
Consortium at high biomass (<i>Scenedesmus sp./cyanobacterium</i>)	2.93	0.70	4.21	This study

425

426 Figure 6 is shows the normalized nutrient uptake as a function of microalgae
 427 population. The normalization has been carried out taking into account the number of
 428 cell counts. In this Figure, it can be observed that higher nutrient uptake was achieved
 429 at the lower biomass concentrations, decreasing with the increase in microalgae
 430 population until reach a minimum of uptake efficiency. This fact can be explained in
 431 the same way that the intracellular nutrient content diminution, being the competence

432 among microalgae in the culture for the scarce available phosphorous the limiting
 433 factor. Also, the available light that is able to reach the microalgae culture decrease
 434 due the auto-shadow effect, reducing the nutrient uptake too. Consequently, although
 435 the total nutrient removal can increase due to the biomass growth, the nutrient
 436 removal efficiency per biomass unity in low biomass concentrations is significantly
 437 higher than that displayed at high biomass concentrations.



438
 439

Figure 6. Nutrient uptake normalized by the cell count versus the population density.

440 Regarding the normalized nitrogen uptake, the experimental data acquired displays a
 441 higher dispersion than that registered by the normalized phosphorous uptake. This
 442 phenomenon can be explained by the phosphorous limitation effect too. Since the
 443 removal of nitrogen in the reactor was not only a function of light or available nitrogen
 444 concentration, but also to the available phosphate concentration.

445 The results obtained suggest that light and phosphorous concentrations seem to be
 446 the most relevant variables for the microalgae growth in this study, supporting the
 447 previous conclusions.

448

449 CONCLUSIONS

450 The light intensity supplied, the available phosphorous concentration, nutrient
451 removal and the competition among microalgae species in a continuous fed
452 photobioreactor has been studied. In the experimental period, *Chlorella sp.* was the
453 dominant specie when light intensity was low ($36 \mu\text{mol s}^{-1}\text{m}^{-2}$), reaching about $244 \pm$
454 83 mg L^{-1} of VSS in the photobiorreactor, with a nutrients removal efficiency of $35.6 \pm$
455 9.9% and $71.8 \pm 13.3 \%$ of nitrogen and phosphorous, respectively. Conversely, when
456 the light intensity supplied was increased to $52 \mu\text{mol s}^{-1}\text{m}^{-2}$, the culture was dominated
457 by a consortium of *Scenedesmus sp.* and *Monoraphidium sp.* increasing the VSS until
458 $674 \pm 86 \text{ mg L}^{-1}$ and reaching a nutrients remove efficiency around $58.4 \pm 15.8 \%$ and
459 $96.1 \pm 16.5 \%$ of nitrogen and phosphorous respectively. The results obtained suggests
460 that *Chlorella sp.* shows a lower I_{critic} that *Scenedesmus sp.* and *Monoraphidium sp.*
461 species, such as previous studies reported before.

462 Nutrient removal ratio analysis reflects that *Chlorella sp.* presents higher dependence
463 for phosphorous concentration than *Scenedesmus sp.* and *Monoraphidium sp.*,
464 showing lower N/P remove ratios. Moreover, when phosphorous concentration was
465 raised from 6 to 15 mgP L^{-1} , the culture was dominated by cyanobacterium, decreasing
466 its abundance until vanish their relative abundance when phosphorous was reduced to
467 6 mgP L^{-1} again. These results clearly indicate that phosphorous concentration has an
468 important influence in the competition among microalgae.

469 The analysis of the biomass intracellular nutrients was coherent with previous
470 conclusions, achieving lower N/P ratios in the composition of microalgae which was
471 attributed more dependence for phosphorous concentration (4.21 ± 0.05 , 6.32 ± 0.28
472 and 7.51 ± 0.14 for a consortium of cyanobacterium and *Scenedesmus sp.* and cultures
473 mostly dominated by *Chlorella sp.* and *Monoraphidium sp.* respectively).

474 This study highlights the importance of light and nutrient concentration in the
475 competence among microalgae, showing the dramatic impact that changes in this two
476 variables can have in the microalgae species that can survive in a culture, and
477 consequently, in the nutrient remove efficiency. Besides, it must be highlighted that

478 these changes can be induced by the microalgae themselves since with their growth,
479 shift the available nutrient concentration and/or the available light.

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