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3 **Habitat evaluation for the endangered fish species *Lefua echigonia* in**
4 **the Yagawa River, Japan**
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Habitat evaluation for the endangered fish species *Lefua echigonia* in the Yagawa River, Japan

Abstract: Spring-fed streams in Tokyo are important habitats for various aquatic species, whereas urbanisation as well as introduction of invasive species are threatening the sustainability of such aquatic ecosystems. This study applies the System for Environmental Flow Analysis (SEFA) in a small urban river in Tokyo to assess the dynamics of the suitable habitats for the endangered freshwater fish *Lefua echigonia* (Jordan & Richardson 1907). A set of Habitat Suitability Curves (HSCs) for water depth, velocity and substrate was developed to evaluate the suitable habitats. The habitat assessment indicated that the Area Weighted Suitability (AWS) reached the maximum at 0.02 m³/s, which is close to the base flow of the target river; a gradual decrease in AWS was observed for higher flows. The temporal distribution of AWS, during forty-one consecutive months, showed that, on average, the best habitat conditions for adult *L. echigonia* occur during the period between January and July, whereas the worst situation occur during the period between August and December. This work presents information and tools for instream habitat analysis that should help managers to conserve this aquatic species and prioritize actions to further rehabilitate urban rivers, using *L. echigonia* as a case study.

Keywords: Environmental flow assessment, instream habitat simulation, habitat suitability curve, hydraulic modelling, small-bodied fish, spring-fed urban stream

Introduction

Urbanisation has led to a significant degradation and biodiversity loss of nestled river catchments due to increases in pollutants concentrations, modifications of river banks and hydrological alteration (Coelho et al. 2014). An acknowledged side-effect is the process by which cities phagocytise endemic-species distribution areas (e.g. McDonald et al. 2018); the extreme case is that of the *Ambystoma mexicanum* (Shaw and Nodder, 1798) whose natural distribution area is restricted to a few ponds of Mexico City

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3 (Recuero et al. 2010). Despite the relevance of urban rivers from an anthropocentric
4 perspective due to the number of ecosystem services they provide (*i.e.* nonmaterial
5 benefits that people obtain from nature such as aesthetic values, cultural heritage or
6 recreation) (Vollmer et al. 2015), the expected rates of metropolis development suggest
7 these processes are unlikely to discontinue over the 21st century (Sardak et al. 2018).
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15 Tokyo has undergone a wide and rapid development and urbanisation process
16 after WWII (Taniguchi et al. 1999), which alike other large conurbations had a
17 significant impact on the enclosed fluvial ecosystems (Vollmer et al. 2015). Tokyo river
18 catchments originally suffered the impact of several channel modification works,
19 especially concrete-lining that turned them into irrigation channels, which were
20 extensively implemented to increase agricultural productivity resulting into
21 deterioration and fragmentation of river habitats (Fukuda et al. 2015). As it has been
22 observed for other urban streams (*e.g.* Lee et al. 2010), Tokyo streams have been
23 subsequently managed and exploited with an emphasis on both flood control and
24 instream water uses.
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38 The urban area of Tokyo city has already engulfed partially the restricted
39 distribution area of the native fish species *Lefua echigonia* (Jordan & Richardson 1907)
40 (fam. *Balitoridae*), which may become a key species for the aquatic biodiversity
41 conservation in spring-fed streams in the eastern part of Japan. Nonetheless, this species
42 was listed as an endangered species by the Environmental Ministry of Japan (category:
43 EN) (Hosoya, 2003). The genus *Lefua* has a very small distribution area in the world; it
44 can be found in the Amur drainage (Korea), north-eastern China and Japan, where only
45 7 species have been described globally (Kottelat 2012). Despite the ecological
46 importance of these species, there is no available information about their habitat
47 requirements. Thus, it is only known that they usually inhabit small headwater streams
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3 in lowland areas (Mitsuo et al. 2009). To fill this knowledge gap, a microhabitat
4 suitability model was developed for adult *Lefua echigonia* in this study.
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8 The Yagawa River (Tokyo, Japan) is a representative of general problems for a
9 significant number of small spring-fed urban rivers and streams in Tokyo, where the
10 loss of natural habitats and habitat diversity due to urbanization and river modification
11 are jeopardizing the sustainability of the aquatic communities (Mitsuo et al., 2009). In
12 this situation, the conservation of endemic species, and more specifically *L. echigonia*
13 and accompanying species, should become a key in biodiversity sustainability. An
14 additional indication of its value is that this fish is thought to be a key host fish for
15 glochidia of Unionidae (Toshishige and Maruyama, 2005). In this regard, the study of
16 species-environment relationships, to later evaluate the impact of different flow
17 regimes, can offer a way to understand the impact of further habitat manipulations such
18 as restoration of nature-like substrates in a concrete-lined section, together with an
19 improved watershed management including rainfall infiltration for spring water
20 restoration in urban and residential areas, which may favour biodiversity conservation.
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37 The natural flow regime (*sensu* Poff et al. 1997) has been stated to be the *master*
38 *variable* that shaped fish assemblages and triggered species adaptation and evolution
39 (Poff 2010). This regime is profoundly altered by river regulation and water diversion,
40 which becomes deleterious to native fish communities. Therefore, the establishment of
41 environmental flows (*i.e.* the quantities, quality, and patterns of water flows) to balance
42 protection of natural environment with out-of-stream uses is a worldwide enterprise
43 (Rosenfeld 2017).
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54 Despite the debate, the physical habitat simulation approach (Bovee et al. 1998)
55 is still considered the most defensible approach, from a legal perspective, to evaluate
56 the impact of different flow regimes on the habitat suitable for the target freshwater taxa
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(Reiser & Hilgert 2018). It combines habitat suitability models with hydraulic simulation and hydrological data to determine the most appropriate environmental flow regime. In accordance it has become not only useful for regulated streams, but also to know the environmental requirements before any development plans are made (O’Keeffe & Quesne 2009).

The first objective of this study was to develop a habitat suitability model for the fish species *L. echigonia* in the Yagawa River. The second one was to implement the physical habitat simulation method in this river. And the third one, to recommend key rehabilitation measures for this river, based on the spatial and temporal analysis of habitat availability in terms of the Area Weighted Suitability (AWS); i.e. the habitat indicator usually termed Weighted Suitable Area (WUA) divided by river length.

In summary, this research presents fundamental ecological information concerning microhabitat use and physical habitat simulation in order to help managers conserve the endemic fish species and prioritize actions of urban river rehabilitation, not only for *L. echigonia* but also for other benthic fish species in Japan and East Asia.

Methods

Study area and target species

The Yagawa River is a 1500-meter-long free-flowing stream located in Kunitachi (i.e. outskirts of Tokyo, Japan) (Matsuzawa et al. 2017b). The river runoff is groundwater-fed (the spring of the river provides the highest flow around October), although it responds rapidly to episodic events of intense rainfalls caused by typhoons (usually in August and September). The riparian habitats present a notable longitudinal variability

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3 with well forested areas intersected with channelized sections. The river is connected to
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5 the Fuchu Yosui irrigation system and contributes to the Tama River (Figure 1).
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15 To obtain dedicated data on the actual flow regime, the flow was gauged
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17 monthly at station 2 (Figure 1) from June 2015 to October 2018 (Figure 2). In this
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19 period, the maximum monthly runoff was observed in October/November, whereas the
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21 minimum runoff occurred in February/March, followed by a period of stable base flow
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23 —only interrupted by storms— that lasted approximately 6 months.
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28 [Here Figure 2]
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33 *Lefua echigonia* is a small benthic fish (total length < 10 cm) with relatively low
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35 swimming ability (Matsuzawa et al. 2017a). The adults burrowed into sand or under
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37 stones and stay in slow currents or static waters (Mitsuo et al. 2007, Aoyama et al.
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39 2007), feeding principally from benthic or planktonic invertebrates (Mitsuo et al. 2009).
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41 The individuals mature between the first and second year and the breeding season takes
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43 place from late March to early June. The females lay adhesive eggs on litter, leaves,
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45 aquatic vegetation or under stones and substrate interstices to avoid scouring (Mitsuo et
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47 al. 2009; Aoyama et al. 2005; Hosoya 2003); the latter is considered an adaptation to
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49 mountainous conditions (Aoyama and Doi 2005). This fish species can be considered as
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51 very sensitive to channelization affecting natural substrate, as well as habitat alterations
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53 related with flow velocity and with any type of barriers that affect longitudinal and
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55 lateral connectivity.
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Data collection

The fish sampling (presence or absence) in the entire reach of the Yagawa River was repeated every month from June 2015 to May 2017 by a hand net covering all the present mesohabitat types (*i.e.* pool, run, riffle) in order to capture the largest spatial and temporal variability (Martínez-Capel et al. 2009; Muñoz-Mas et al. 2017). Concomitantly, mean water column velocity (hereafter velocity; m/s) and depth (m) at the sampling spot were collected at the microhabitat scale. Velocity was measured with a current meter (VP30, KENEK, Tokyo, Japan) and depth was measured with a wading rod to the nearest millimetre. In addition, the percentage of substrate classes was visually estimated following previous studies (Martínez-Capel et al. 2009; Muñoz-Mas et al. 2017); 1, vegetation; 2, sand ($\emptyset \leq 2$ mm); 3, small-sized gravel ($2 < \emptyset \leq 16$ mm); 4, medium-sized gravel ($16 < \emptyset \leq 64$ mm); 5, large-sized gravel ($\emptyset > 64$ mm); 6, concrete and continuous rock. Physical habitat measurement was additionally performed where fish was absent for better illustrating species specific habitat suitability. During this 2-year survey, flow discharge ranged from 0.002 m³/s to 0.3 m³/s which covers almost all flow dynamics in the river. Immature fish presented an even distribution across the river specifically after their emergence (*i.e.* with evident lack of habitat selection), thus we focused exclusively on the habitat selection shown by adults.

Habitat suitability model

The relationship between the hydraulic variables and their suitability (suitability index, SI) was mathematised with univariate habitat suitability curves (HSCs) (Waters 1976) based exclusively on presence observations (*i.e.* Use or category II HSCs) (Bovee 1986). For each of the three variables (velocity, depth and substrate), frequency histograms were built considering exclusively the data collected in the sampling spots where at least one individual was observed (presence data). The histograms were

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3 normalised between 0 and 1, and the right tail indicating very low suitability (suitability
4 index ≤ 0.1) were set to 0.1 in order to fill intervals without data. Given the large
5 quantity of microhabitat data (presence/absence observations covering the entire river),
6 this procedure meant no subjective intervention on the resulting curves. The
7 microhabitat selection (non-random use) was statistically tested —use or presence
8 *versus* availability or absence data (Hayes and Jowett, 1994; Martínez Capel 2000)—
9 with a Mann-Whitney U two-sample test ($p < 0.05$) using STATGRAPHICS Centurion
10 XVII software ver. 17.2.04 (2014 Statpoint Technologies, Inc. USA).
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23 ***Hydraulic modelling***

24 The software tool SEFA (System for Environmental Flow Analysis; Payne and Jowett
25 2012; Jowett et al. 2014), which follows the scheme of the Instream flow Incremental
26 Methodology (IFIM, Bovee et al. 1998), was used to perform the physical habitat
27 simulation. With SEFA, a one-dimensional hydraulic model was generated and
28 calibrated, based on the cross-sections —108 sections covering the 1338-m-long
29 segment— and water surface elevation data from field surveys (collected with an optical
30 levelling instrument). The calibration flow was calculated as the trimmed mean flow
31 rate considering the gauges obtained in every transect (10% out). The module for Water
32 Surface Profile (WSP) modelling, based on the common standard-step calculation
33 method, was used to simulate representative river flow rates [0.01, 0.20 m³/s] based on
34 the gauged flows described above (June 2015 - October 2018).
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51 ***Habitat evaluation***

52 The reach Area Weighted Suitability (AWS) (m²/m) —usually termed Weighted Usable
53 Area – WUA (Bovee et al., 1998) but in terms of m² per metre of river length—, was
54 calculated for every simulated flow. The SEFA software calculates first, the Composite
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3 Suitability Index (CSI), by multiplying the habitat suitability (between 0 and 1) for the
4 three microhabitat variables, namely depth, velocity and substrate, at each measurement
5 point or simulation cell (Muñoz-Mas et al. 2012; Jowett et al. 2014). The AWS is
6 calculated by evaluating the CSI at each model cell (CSI), then multiplying it by the
7 proportion of the reach area represented by that cell and summing over the reach (Jowett
8 et al. 2014). In order to characterise the spatial variability of habitats available for *L.*
9 *echigonia*, a longitudinal distribution of AWS for several flow rates was calculated.
10 Finally, a time series of AWS was calculated to represent the temporal variability of *L.*
11 *echigonia* habitats along the Yagawa River from June 2015 to October 2018. The latter
12 was performed at the monthly time scale.
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28 **Results**

29 The habitat characterisation demonstrated the diversity of microhabitats measured
30 during the two years of regular surveys (Figure 3), across 108 transects located in
31 1381.5 m long of the Yagawa River. The range of the microhabitat variables reached a
32 maximum velocity of 1.068 m/s, maximum depth of 0.40 m; all the substrate types were
33 present, and channel wetted width ranged from 1.10 to 4.65 m. The distribution of
34 substrate classes along the cross sections in the river reach is shown in Figure 4. A total
35 number of 406 presence data were collected for adult individuals of *L. echigonia*; the
36 number of absence data was 2775. The results of the Mann-Whitney tests showed a
37 non-random use of the microhabitats ($p < 0.05$), i.e. a significant statistical difference
38 (presence *versus* absence data) in both depth and velocity. This indicated that *L.*
39 *echigonia* was performing an active habitat selection in the Yagawa River.
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8 The HSCs (Figure 5) showed that the highly suitable depth ($SI > 0.5$) for the adult *L.*
9 *echigonia* in the Yagawa River ranges from 0.043 to 0.15 m, reaching the maximum
10 suitability at a depth around 0.075 m. Based on our observation, the suitability was set
11 to 0.1 for areas deeper than 0.17 m. The suitable range of mean velocity was observed
12 between 0.00 and 0.09 m/s, with the maximum suitability for velocities equal or smaller
13 than 0.05 m/s. The presence data were scattered beyond 0.20 m/s, thus the suitability
14 index for that range was set to 0.1, founded on the knowledge that the swimming ability
15 of this fish species is very limited at velocities over 0.2-0.3 m/s (Matsuzawa et al.,
16 2017a). In terms of substrate, vegetation and large gravel were the most suitable types.
17 On the contrary, concrete or continuous rock was considered unsuitable. Concerning silt
18 and boulders, these types were not observed in the study site. Nevertheless, there are 8
19 substrate classes listed in Figure 5 (including those aforementioned) because the HSCs
20 must fit the categories described in the SEFA software.
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45 The stream flow was calculated using all the 108 transects surveyed (cross-sections),
46 resulting in the calibration flow of 0.044 m³/s. The precise measurement of the
47 topography, water surface elevation and mean velocity in these transects allowed the
48 calculation of the Manning's roughness coefficient N (ranging between 0.03 and 0.17).
49 The survey results indicated the most reliable data between transects 9 and 76 (with a
50 relevant hydraulic jump above transect 76), thus the calibration of the hydraulic model
51 was performed there. The sensitivity analysis showed negligible sensitivity to errors in
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3 water surface elevations in the transect number 13 and upstream until transect 76, where
4 the physical habitat condition was evaluated. Therefore, the total length of the
5 simulation reach was of 805 m, representing 60.2 % of the total measured length. At the
6 calibration flow, the AWS for the simulation reach was 0.360 m²/m, representing the
7 availability of suitable habitat for the time of the survey, i.e. the week from the 28th May
8 to the 3rd of June 2018. The curve relating AWS with river flow rate showed the highest
9 value at a flow of 0.020 m³/s, for which the AWS representing the whole simulation
10 reach gets a value of 0.465 m²/m (Figure 6). The curve slowly goes down and tends to
11 the lowest value, approximately 0.150 m²/m, which corresponds to the highest
12 simulated flow, 0.200 m³/s.
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33 On the spatial distribution of the AWS per transect an irregular pattern was observed,
34 with ample differences in habitat conditions (i.e. a range from near to zero to 1.89 m²/m
35 (Figure 7). The general trend showed that higher flows produced smaller amounts of
36 habitat availability. Therefore, the highest values of AWS were observed during the
37 lowest flow rate (0.010 m³/s) in five groups of transects alternating with areas of low
38 suitability, in the upper part of the simulation reach. These groups of suitable transects
39 correspond in every case to pool-type mesohabitats, where at the calibration flow the
40 mean depth was larger than the average, mean velocity was lower than the average,
41 vegetation was not present or was very scarce and the proportions of concrete were
42 relevant (XS 37-40, XS 44-45, XS 48-51, XS 58-61, XS 64, in Figure 4). On the
43 contrary, in a small group of transects, the highest values of AWS were achieved at the
44 flow rates of 0.150 and 0.200 m³/s (XS-4 corresponding to station 2; and XS-11, 12,
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3 13). These transects encompassed habitats where at the calibration flow the mean depth
4 was equal or smaller than the average, mean velocity was similar or higher than the
5 average and substrate was natural (concrete was mostly absent) regardless the
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10 mesohabitat type (run, riffle and pool).

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14 [Here Figure 7]

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19 The temporal distribution of AWS, during forty-one consecutive months, indicated that,
20 on average, the best habitat conditions for adult *L. echigonia* occurred during the period
21 between January and July (Figure 8). However, the year 2018 showed a particular
22 situation because there was an extreme event when the river flow was almost depleted
23 during February 2018 (0.002 m³/s); and in the following months, the availability of
24 suitable habitat, in terms of AWS, was relatively low in comparison with the previous
25 two years. In general, the period with the lesser habitat availability corresponded to the
26 months between August and December. In September 2015, 2016 and 2018 the lowest
27 values of habitat availability were recorded, with AWS between 0.142 and 0.158 m²/m.
28 Similarly, the minimum habitat in November 2017 corresponded to 0.142 m²/m. On the
29 contrary, the highest values corresponded to February 2016 (0.464 m²/m), with a flow
30 rate of 0.02 m³/s, the period of April-June 2017 (ca. 0.464 m²/m), with similar flow rate,
31 and January 2018 with the same values in AWS and flow rate. This means that for the
32 recorded range of flows, the AWS can span up to three-fold the lowest value.

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Discussion

The first objective of developing habitat suitability curves (HSCs) for adult *Lefua echigonia* in the Yagawa River was successfully completed. The HSCs allow for the implementation of the physical habitat simulation in this and other similar rivers with similar geomorphology and hydrological dynamics, although a validation procedure (i.e. transferability test) should be strongly encouraged (e.g. Thomas & Bovee, 1993). This is one of the first studies of habitat simulation in small spring-fed urban rivers of Tokyo (see Matsuzawa et al. 2017b), which makes it relevant for the biodiversity conservation of urban small spring-fed rivers. Consequently, any derived management plan should help those rivers (very abundant around the fluvial terrace of the Tama River where a part of metropolitan area spreads) to sustain healthy aquatic environments and native species. Furthermore, the accumulation of knowledge of small-bodied fish can provide relevant information for conservation and restoration planning and for the management of diverse ecosystems or species (Fukuda 2011).

Although stream flow and other factors affect the microhabitat use by fish (e.g. Orth 1987; Lambert 1994; Martínez-Capel 2000), in this study, the bias related with survey data was considered very improbable because the microhabitat data were very abundant ($N = 406$ presence data) and the test for non-random use, demonstrated that habitat selection was correctly identified. In addition, microhabitat data were collected during a relatively long period (24 months), and under a wide range of microhabitat conditions considerably ampler than those selected by the fish.

The HSCs exhibited the microhabitats selected by this fish species, offering the first description of the fish behaviour at the microhabitat scale in the wild, and consequently of the habitat dynamics as the flow changes across seasons. The HSC for velocity showed the selection of *L. echigonia* for quiet and very slow waters, preferably

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3 below 0.09 m/s; this result resembles previous studies, which indicated that adults
4 usually stay in slow currents or static waters (Mitsuo et al. 2007). To the best of our
5 knowledge, there exist some reports on microhabitat use of this fish species, but they
6 are limited to specific habitats like spring-fed pool-like habitats and ponds. Other fish of
7 the same family have received more attention, especially concerning some Eurasian
8 species, such as the stone loach, *Barbatula barbatula* (L., 1758). This fish has been
9 associated with an ample variability of microhabitat use (Copp & Vilizzi, 2004), in
10 relation with different factors, e.g. ontogenetic status, time of day, and food availability,
11 but in general selecting low flow velocity (0.05-0.10 m/s) (Copp, 1992). Thus,
12 demonstrating a great coincidence with our results for *L. echigonia*. On the other hand,
13 in accordance with that plasticity in habitat use demonstrated by other similar species,
14 Lamouroux & Capra (2002) described the interval of suitable mean velocity for the
15 stone loach (suitability index > 0.5) between ca. 0.1 and 0.8 m/s. These results suggest
16 that *L. echigonia* may depict a greater plasticity in terms of habitat selection in other
17 rivers with different habitat conditions.

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In the Yagawa River, the maximum recorded velocity (> 1 m/s) demonstrated
the ample range of conditions of this study area. However, the highest high velocities
coincided in a greater proportion to large-sized gravel substrates that can create a slowly
flowing microhabitat around them. Further study is needed for a deeper understanding
of *L. echigonia* microhabitat uses under high velocity conditions, leading to an
improved habitat assessment even under what can be currently considered unsuitable
conditions. The habitat selection in terms of depth indicated a high suitability of very
shallow waters, mainly between 0.04 and 0.13 m, while the available microhabitats
sampled reached 0.40 m depth, which foregrounds the very small size of the river. This
observation generally matches former studies, which indicated that *L. echigonia*

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3 inhabits shallow, slowly-flowing streams and ponds where the natural environment has
4 been well conserved (Hosoya, 2003). The selection of shallow waters could be related
5 with the banks where vegetation was frequent, since different authors have identified
6 vegetation as a favourite substrate for depositing the eggs (Mitsuo et al., 2009).
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12 An analogous reasoning can be performed regarding the relatives of the family
13 *Cobitidae* of Europe, since the stone loach *Noemacheilus barbatulus* (L., 1758), the
14 kissing loach *Leptobotia curta* (Temminck & Schlegel, 1846), the small striated spined
15 loach *Cobitis sp.*, and the middle striated spined loach *Cobitis sp.*, are known to use
16 spaces among submerged plants for their spawning site (see Aoyama & Doi, 2006,
17 including further references therein). Similarly, Bohlen (2003) demonstrated that
18 *Cobitis taenia* (L., 1758) showed a strong preference for dense vegetation as a spawning
19 substrate in German rivers; it was then hypothesized that vegetation provides sheltering
20 environments as decreasing egg predation and as preventing eggs from drifting away
21 even in rapid currents (Bohlen, 2003). Another study on stone loach (*B. barbatula*), also
22 indicated a strong preference for shallow microhabitats with depositional substrate,
23 sometimes combined with some type of cover (Prenda et al, 1997). In accordance,
24 Welton et al. (1983) found that, in the Mill Stream (England), stone loach had a clear
25 preference for macrophyte areas with a substratum of silt.
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44 The results for substrate indicated preference for vegetation (usually
45 accompanied with fine substrate) as well as for medium- and large-sized gravel in
46 accordance with the lithophilous character of *L. echigonia* (Aoyama & Doi, 2006). Our
47 observations partly coincide with a close relative species of the family *Cobitidae* in
48 Japan, *Cobitis kaibarai* (Nakajima, 2012), for which the population seemed to be
49 correlated with emergent hydrophytes with silt substrate in non-spawning season (Kim
50 et al., 2012). Concerning the Eurasian relative stone loach, there is abundant
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3 information about the preference of juveniles and adults for coarse substratum (Copp &
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5 Vilizzi, 2004). Thus, it is generally considered that stone loach prefers fine-to-medium-
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7 size substrate in rivers (Copp et al., 1994; Zweimüller, 1995), which may be interpreted
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9 as a resemblance of the habitat use of *L. echigonia*. As partly mentioned earlier, the
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11 function of medium- and large-sized gravel to alleviate high velocity should also be
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13 considered here as observed in the Yagawa River.
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17 The physical habitat simulation in a highly representative river reach (more than
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19 a 60% of the total length) showed an irregular spatial distribution of the suitable
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21 habitats, related with microhabitat and mesohabitat types, as well as the high influence
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23 of channelised sections (see Figure 4). Nonetheless, the coefficient of correlation
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25 between percentage of vegetation and concrete, for all the cross sections, was -0.706 . In
26
27 urban streams, it is common to observe hydrological and morphological modifications,
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29 with consequent morphological adjustments degrading the instream habitats and
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31 simplifying the streambed structure leading to pools aggradation and riffles erosion
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33 (Gregory et al., 1994). The larger reductions on the AWS were observed in transects
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35 located in the upstream segment, with concrete in proportions generally between 20 and
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37 30% (Figure 4). In these sections, the low flow favours low flow velocities in pools and
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39 the opposite in channelised sections where the effect is aggravated because the shelter
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41 against swift current provided by natural vegetation is generally scarce. In such
42
43 conditions, when the flow rises fish tend to avoid environments with large fluctuations
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45 in velocity because turbulent flows can cause fish dragging (Hockley et al, 2014). On
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47 the contrary, some transects in the downstream segment (generally from section 3 to
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49 22), with a high proportion of vegetation (Figure 4), showed an increase of the AWS for
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51 higher flows, regardless the mesohabitat type. These results indicated that bank
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3 channelisation could nowadays be the main threat for the species conservation, which
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5 could be extrapolated to other taxa adapted to such kind of streams.
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9 **Conclusions and recommendations**

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11 The habitat time series during more than three years of study, allowed us to
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13 confirm that the critical periods, in terms of habitat, for *L. echigonia* are related with
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15 high flows. The optimum habitat conditions occur during a great part of the year,
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17 including the breeding season (approximately March-June). However, the critical
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19 periods usually start in August, when the juveniles are approximately four months old,
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21 thus potentially jeopardizing the sustainability of the fish population. The spatial
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23 patterns and the temporal trends described above indicate that during high flows some
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25 mesohabitats could offer good protection (small pools in the upper part), but the
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27 coincidence of those slow-water mesohabitats with sections where concrete is more
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29 abundant is hindering their utility as natural shelter during high flows. Therefore, we
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31 hypothesize that high flows and habitat degradation may produce a critical bottleneck in
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33 the fish population; this is an aspect that deserves further research in the Yagawa River
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35 and other small urban rivers in Japan.
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41 On the other hand, the time series show a relatively consistent pattern of
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43 seasonal flows, with small inter-annual variability, which is characteristic of spring-fed
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45 river systems, where the aquifer recharge and groundwater quality should be protected.
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47 Indeed, urbanisation has hampered rainfall infiltration, which has resulted in decreases
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49 or in the worst case depletion of spring water in many parts of Tokyo. To cope with it,
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51 Kunitachi City Government has launched the basic plan for water cycle (Kunitachi City
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53 Government, 2016) to achieve sustainable water management considering the balance
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55 between nature and human water needs.
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3 Although it is quite difficult in the urbanized areas of Tokyo, where possible, it would
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5 be advisable to naturalize the form of the banks to some extent and to facilitate the re-
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7 naturalisation, in order to facilitate the natural colonisation by aquatic vegetation. Some
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9 studies have developed and evaluated pool-riffle structures specifically design for
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11 constrained, low gradient channels of urban streams, producing improvements in the
12
13 mesohabitat structure, fish abundance, biomass and diversity during the two years after
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15 construction, especially concerning the fish of slow-water species (Schwartz and
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17 Herricks, 2007). In addition, the enhancement of river connectivity is an interesting
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19 measure, given the high sensitivity of *L. echigonia* and other benthic fish to any type of
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21 barriers; the conservation of this fish species requires not only the protection of habitats
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23 at the local level, but also planning and management at the regional level, considering
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25 the river network as the proper study area to maintain the genetic diversity of this
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27 species (Mitsuo et al. 2009).
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33 In the general context of urban streams, it becomes relevant the citizens'
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35 environmental education and specific programs for citizen science, which can make a
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37 difference to involve the neighbours in river conservation and help them understand the
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39 importance of conserving and rehabilitating the ecosystems biodiversity (Smith et al.
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41 2016). Taking and processing data, engages people with environmental issues, and
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43 surely with their immediate environment (Pocock et al. 2014). For instance, a
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45 programme involving school activities on the fish species and the river ecosystem, as
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47 well as the citizens' participation in the rehabilitation and monitoring of the Yagawa
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49 River, including the river habitats and fish populations, should help river management
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51 and conservation. A carefully designed program of citizen science potentially allows to
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53 collect data at much larger spatial and temporal extent and much finer resolutions than
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55 would otherwise be possible, representing a cost-effective way of collecting such data
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3 while also providing an excellent opportunity for people to become engaged (Pocock et
4
5 al. 2014).

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8 In conclusion, this work presents reliable ecological information and the tools
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10 needed to help managers conserve the endemic fish species and prioritize actions to
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12 rehabilitate urban rivers, not only for *L. echigonia* but also for other benthic fish
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14 species. Finally, this work should also serve as a reference for further ecological studies
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16 on habitat use by fish of the family *Nemacheilidae* and other close relatives in Japan and
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18 East Asia.
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21 22 23 **Geolocation information**

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25 Coordinates of the study site; 139.429210 East; 35.679981 North.
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28 29 **Acknowledgements**

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32
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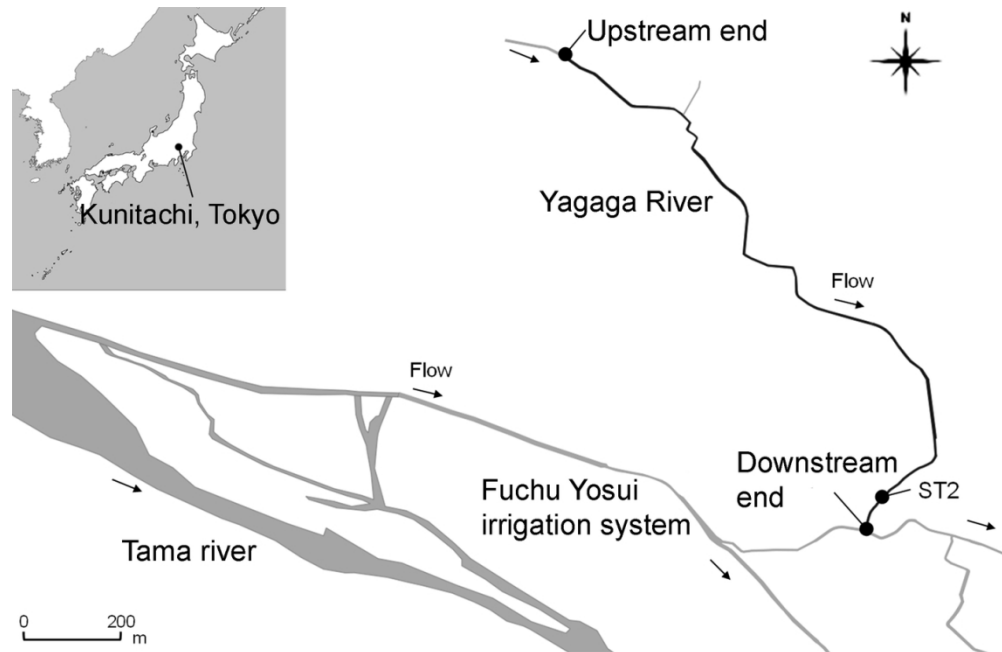


Figure 1. Location of the study area in Kunitachi (Tokyo, Japan), and map showing the upstream and downstream end of the study reach in the Yagawa River. Station 2 near the downstream end (ST2) was selected to display the river flow patterns (Figure 2).

119x77mm (300 x 300 DPI)

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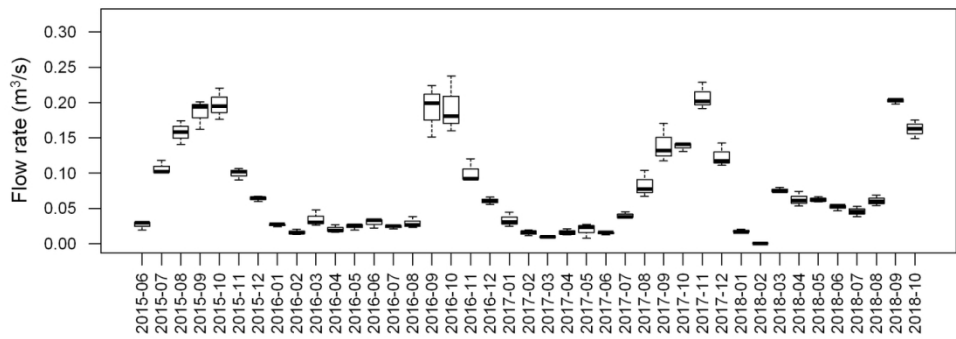


Figure 2. Flow regime (m³/s) monitored monthly at station 2 in the Yagawa River (see Figure 1) from June 2015 to October 2018.

124x47mm (300 x 300 DPI)

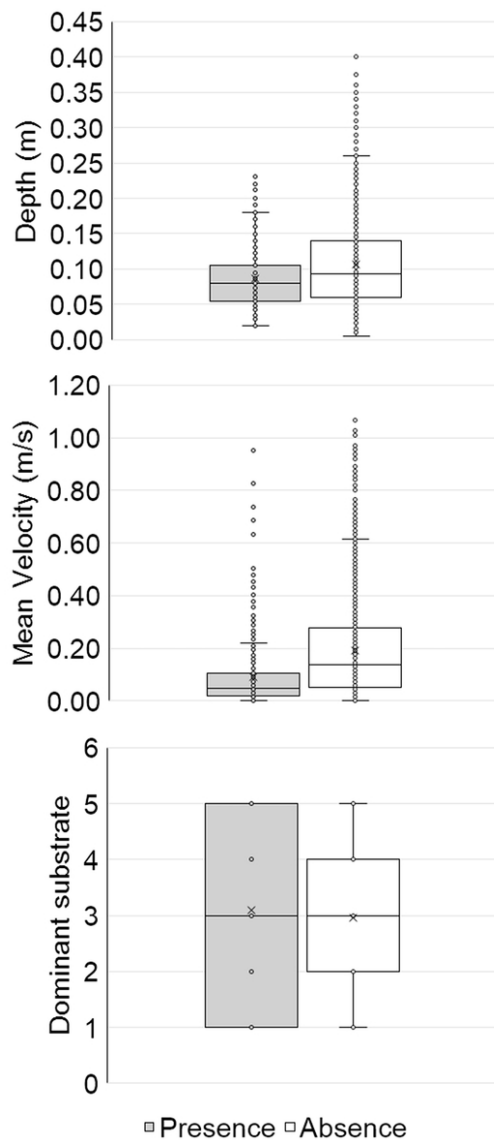


Figure 3. Box-plots showing the habitat variability in the Yagawa River during the study period (June 2015 – May 2017) in terms of velocity (m/s), depth (m) and substrate index (1-6). Substrate codes were; 1, vegetation; 2, sand (< 2 mm); 3, small-sized gravel (2–16 mm); 4, medium-sized gravel (16–64 mm); 5, large-sized gravel (> 64 mm); 6, concrete and continuous rock.

49x104mm (300 x 300 DPI)

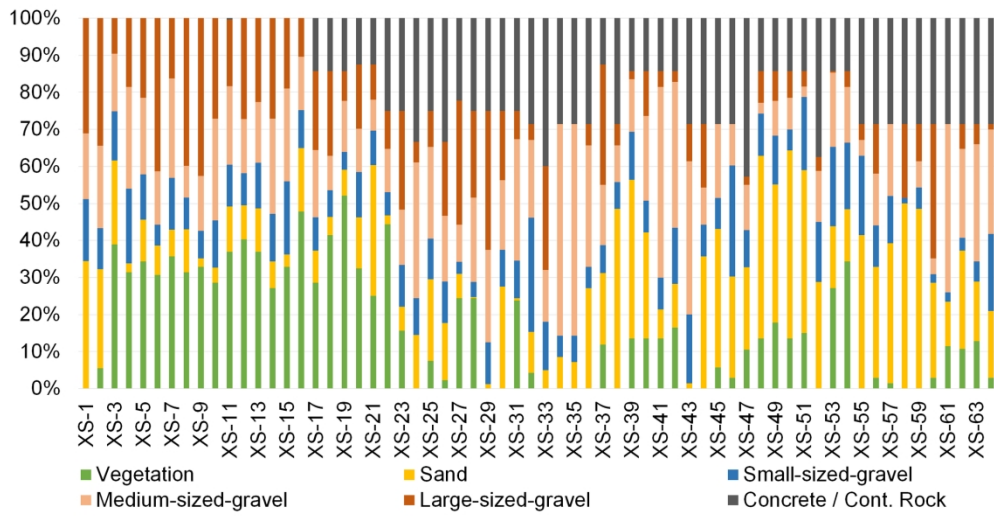


Figure 4. Proportion of substrate types recorded in the field survey in the Yagawa river, at the calibration flow (0.044 m³/s) per transect (from the downstream to upstream end) over the simulation reach.

93x51mm (600 x 600 DPI)

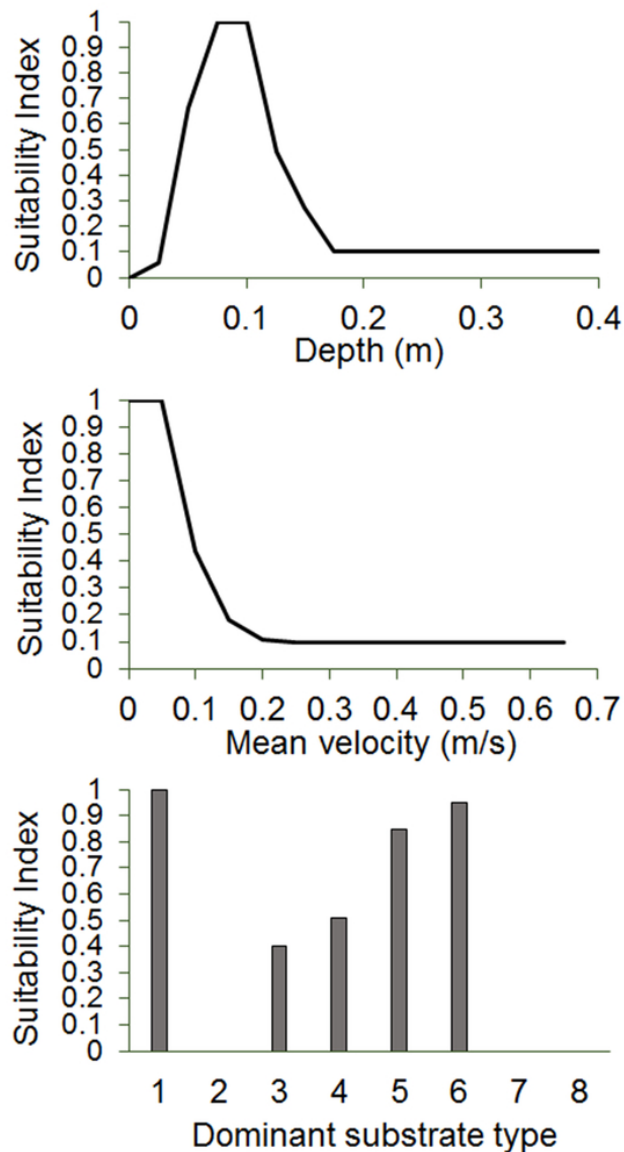


Figure 5. Microhabitat suitability curves for adult *Lefua echigonia*; (a) depth (m), (b) velocity (m/s) and (c) dominant substrate type; 1-vegetation; 2-silt; 3-Sand (< 2 mm); 4-small gravel (2-16 mm); 5-medium-sized gravel (17-64 mm); 6-large-sized gravel (65-256 mm); 7-Boulders (>256 mm); 8-concrete and continuous rock.

49x87mm (300 x 300 DPI)

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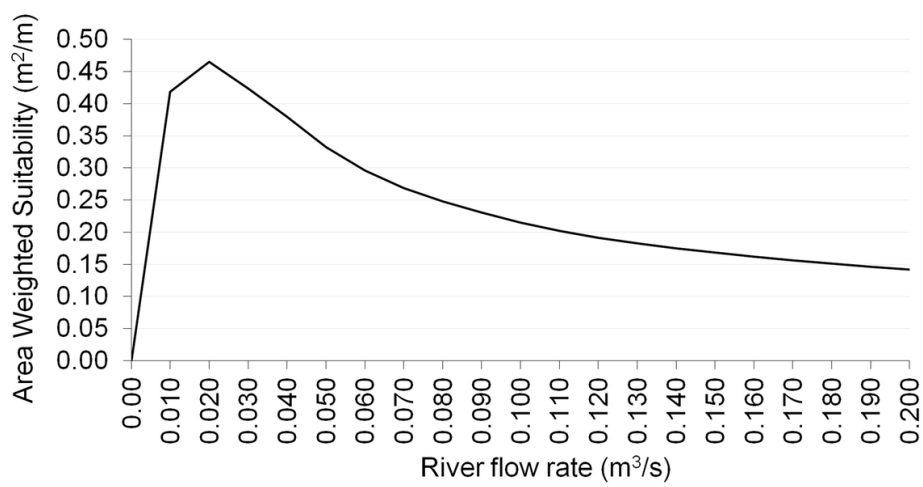


Figure 6. Habitat availability in terms of Area Weighted Suitability (AWS) versus river flow (m³/s); AWS is equal to Weighted Usable Area (WUA) per metre of river length.

104x54mm (300 x 300 DPI)

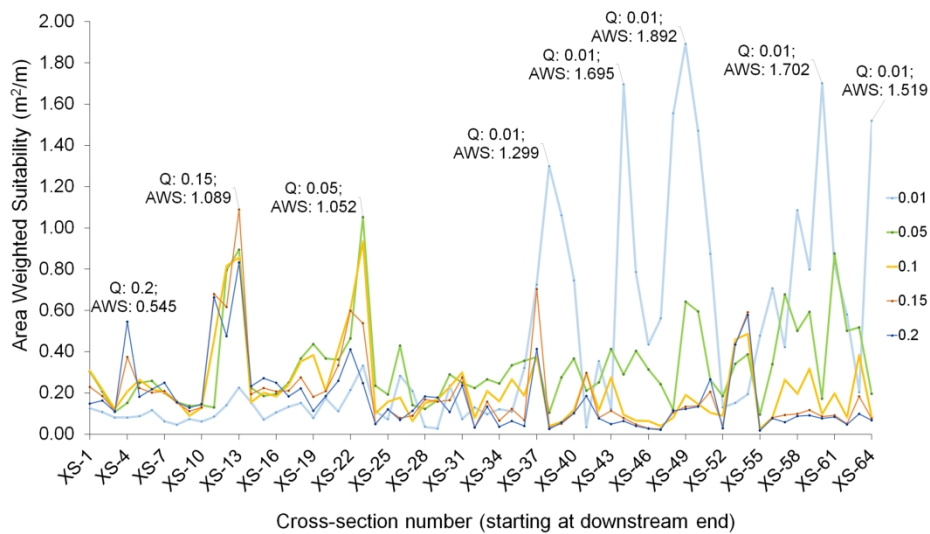


Figure 7. Area Weighted Suitability (AWS) evaluated per transect (from the downstream to upstream end) over the simulation reach, for five values of river flow rate. Transect 4 (XS-4) corresponds to station 2 (see Figure 1).

114x69mm (600 x 600 DPI)

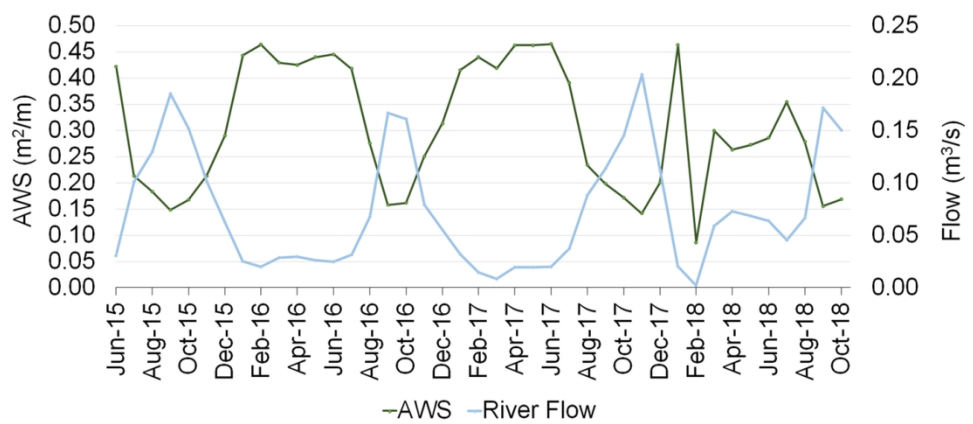


Figure 8. River flow recorded in the Yagawa River and Area Weighted Suitability (AWS) evaluated for the period between June 2015 and October 2018.

126x59mm (300 x 300 DPI)