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Additional Information

1	Valorization of agro-industrial waste from cotton for production of bio-oil and
2	biochar
3	
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15	ABSTRACT
16	Cottonseed residue has a high lignocellulosic content, and as such, is a promising
17	candidate for thermochemical techniques that convert biomass into solid, liquid and
18	gaseous fractions, such as pyrolysis. Cottonseed was submitted to fast pyrolysis in a
19	fixed bed reactor and the pyrolysis products were analyzed using several techniques,
20	including elemental analysis, scanning electron microscopy, Fourier-transform infrared
21	spectroscopy, nitrogen adsorption isotherms and two-dimensional gas
22	chromatography. The detailed composition of the bio-oil was investigated using
23	GC×GC/TOFMS combined with software tools, retention indices and dispersion plots.
24	A total of 257 compounds were tentatively identified and 168 were confirmed by LTPRI.
25	The most abundant components of cottonseed products were phenols and nitrogenous

compounds. The amount of nitrogen compounds in the bio-oil distinguishes it from bio-26 27 oils produced from other sources of biomass. These compounds may be very useful in the chemical, food, pharmaceutical and biofuel industries. The higher heating values 28 of cottonseed and bio-oil were 19.34 MJ kg⁻¹ and 34.25 MJ kg⁻¹, respectively, 29 demonstrating the feasibility of the use of cottonseed in its natural form in energy 30 generation or as a secondary source. The biochar had a significant carbon content and 31 a high heating value (22.12 MJ kg⁻¹), making it attractive for fuel applications as a 32 substitute for coal. The activation methods used were able to improve the physical and 33 chemical characteristics of the biochar, as demonstrated by methylene blue adsorption 34 35 tests. The maximum adsorption capacity of NaOH-activated biochar was 23.82 mg g ¹ while that of K₂CO₃-activated biochar was 332.40 mg g⁻¹.The Langmuir isotherm 36 provided the best fit to the experimental data. These results indicate that chemically-37 38 activated biochar may be used as an adsorbent in addition to many other applications. 39

40 Key Words

Biomass, agro-industrial waste, bio-oil, GC×GC/TOFMS; biochar, chemical activation

43 **1. INTRODUCTION**

The use of renewable energy is becoming increasingly imperative to reduce the environmental impact of human activity, economic development and lifestyle changes observed in recent decades. The global energy matrix has suffered changes over time and although biomass does not represent a major contribution to the global energy is expected to play an important role in the global energy balance in the near future. It can be derived from a variety of sources, including but not limited to wood residue, agricultural residue and organic matter from municipal solid waste [1,2]. Biomass can

also yield organic and inorganic compounds with high added value through thermal
degradation processes such as pyrolysis and liquefaction [3,4].

The thermal decomposition of organic substances requires an inert atmosphere 53 (nitrogen or argon) and high temperatures (generally between 400 and 1000 °C), and 54 yields solid (biochar), liquid (bio-oil) and gaseous (non-condensable) products [5,6]. 55 Bio-oil is commonly considered as a substitute to petroleum, both in the energy and 56 petrochemical industries and on the other hand, biochar is a carbonaceous material 57 with high density of porous and a large amount of fixed carbon. Its applications are 58 varied and depend on the characteristics of the material itself. Many sources of 59 60 biomass can be used in the production of bio-oil/biochar. Some of the most 61 advantageous are agro-industrial and domestic waste, which cannot be used for food production like other types of biomass (e.g., oilseeds used to produce biodiesel) [4,7]. 62 63 Cottonseed is an abundant agricultural residue that can be used as raw material in the conversion of biomass into energy and chemical products. The residue from fiber 64 production (seeds) is used in the extraction of vegetable oil, and the remaining seeds 65 are used as animal feed; but these uses are limited by the high fat content (20%) and 66 toxic phenolic compounds present in the seeds [8,9]. As such, cottonseed residue may 67 68 represent a promising source of clean energy and an alternative to fossil fuels as a primary source of chemical compounds. 69

The use of derived chemical products, bio-oil and biochar establishes a cycle of reuse, where all waste generated by a given process is used to obtain new valuable materials, promoting a zero waste industry. This study provides a complete characterization of cottonseed and its pyrolysis products to encourage the further use of these materials as renewable energy resources.

75

76 2. MATERIALS AND METHODS

77 2.1 Pyrolysis of Cottonseed

Cottonseed was obtained from the Planalto Farm, Costa Rica, Mato Grosso do Sul, 78 Brazil. Prior to pyrolysis, the seeds were mechanically pressed to remove excess oil. 79 Fast pyrolysis was conducted using a tube furnace and quartz reactor. Ten grams of 80 cottonseed were used in the experiment, with a (N₂) gas flow of 150 mL min ⁻¹. The 81 resulting vapors, including the bio-oil, were then passed through an ice-water 82 condenser (with water at T ~-5 °C) to cause partial condensation, with the liquid 83 collected in the collection flask. The sample was heated at 100 °C min ⁻¹ until the 84 temperature reached 550 °C. The target temperature was maintained for 15 minutes. 85 Mass yield (w/w%) was measured in triplicate, and determined by dividing the weight 86 87 of the condensate (crude bio-oil) by the initial weight of solid biomass. The crude biooil was separated into aqueous and organic phases by liquid-liquid extraction (LLE), 88 using five 5 mL portions of dichloromethane (DCM). After the reactor had cooled, the 89 organic phase was analyzed and the biochar was collected and weighed to determine 90 yield [10]. 91

92 2.2 Proximate and Elemental Analysis

The moisture, volatile matter and ash content of the cottonseed biomass were determined according to Spanish standard methods UNE-EN ISO (18134-1/18134-2/18134-3:2015), UNE-ISO 18123:2015, UNE-EN ISO 18122:2015, respectively, and fixed carbon was determined by mass difference. Elemental analysis was performed on samples of biomass, bio-oil and biochar (activated and non-activated) using a CE Instruments CHNS1100 device to determine the percentage of carbon, nitrogen, hydrogen and sulfur in the samples. 100 The higher heating value (HHV) for the bio-oil was measured using a bomb calorimeter

101 (IKA Werke model C5003) in accordance with ASTM D240.

102 2.3 Thermal gravimetric analysis

103 Thermal gravimetric analysis (TGA) was performed using a Mettler Toledo TGA/SDTA 104 851 (Columbus OH) analyzer, heating the sample from 25 °C to 900 °C at 10°C min ⁻¹ 105 under an inert atmosphere (ultra-pure nitrogen at 25 mL min ⁻¹).

106 **2.4 Infrared Spectroscopy (FTIR)**

107 FTIR was performed using a Cary 630 FTIR-ATR spectrometer (Agilent Technologies).

The samples were scanned over a wide range (400 cm $^{-1}$ to 4000 cm $^{-1}$) for a total of 64 scans at a resolution of 4 cm $^{-1}$.

110 2.5 Scanning Electron Microscope (SEM)

111 The analyses were performed using a scanning electron microscope with a field 112 emission gun (FEG) (HITACHI, S-4800) at an acceleration voltage of 10 kV and 113 magnifications ranging from 100x to 6000x.

114 **2.6** Obtaining the N₂ - BET Adsorption Isotherms

The physical characteristics of the biochar surface were examined using the BET method (Brunauer, Emmett and Teller) and the micropore size distribution was determined using density functional theory (DFT) [11,12]. Prior to evaluating the adsorption isotherms, 50 mg of the sample were degassed for 30h at room temperature in a vacuum chamber. Nitrogen adsorption and desorption isotherms were then measured at the boiling temperature of nitrogen using a Micromeritics Tristar II Kr 3020 surface analyzer (Micromeritics, Germany).

123 2.7 Determination of pH at the point of zero charge (pH_{PZC})

For each material, a curve was obtained by plotting the pH of the equilibrium solution 124 as a function of the solid mass fraction. The pH at the point of zero charge (pHpzc) of 125 126 the materials was determined throw the measure of the pH value of the suspension with the highest solid fraction when pH evolution with solid concentration is low; that 127 128 is, the plateau in the plot of equilibrium pH versus solid weight fraction corresponds to the pHpzc of the sample. pHpzc was determined according to the method proposed by 129 Noh and Schwarz [13] with the modifications proposed by Reymond and Kolenda [14]. 130 Measurements were performed using a HACH Sension⁺ PH1 DL Portable pH meter 131 (GB-Manchester) using 100 mg of the dry sample (biochar) divided into 5 glass vials 132 133 (20 ml), to which appropriate amounts of Milli-Q water were added to achieve mass 134 ratios of 10, 8, 6, 4, and 2% of biochar to water. Samples were kept under continuous agitation at a constant temperature of 25 °C. pH was measured after 24 h of contact 135 to ensure all samples had enough time to reach pH equilibrium. 136

137 **2.8 Chemical activation of biochar**

138 The chemical activation of cottonseed biochar was carried out using two different 139 processes in order to compare their relative effectiveness.

140 **2.8.1** Activation with alkaline solution (NaOH)

Activation with sodium hydroxide was performed using the method proposed by Jin et al., [15]. The process was performed by adding 2 g of biochar to 500 mL 2 mol L⁻¹ sodium hydroxide solution (NaOH, 99 % purity, Sigma Aldrich, Seelze/Germany) kept under stirring for 2 h. The suspension was then filtered and the material recovered with filter paper. After filtration the samples were washed with deionized water and the suspension was stirred several times until the pH of the filtrate remained stable. The
material was then filtered and dried for 12 h in an oven at 105 °C.

148 **2.8.2** Activation with Potassium Carbonate (K₂CO₃) in inert atmosphere

This process was performed according to the methodology proposed by Maciel et al., 149 150 [16] with some modifications. Briefly, the biochar was placed in contact with the activator (K₂CO₃, 99 % pure, Sigma Aldrich, Seelze/Germany) at a biochar/carbonate 151 mass ratio of 1:3 in a porcelain capsule. The mixture in the capsule was then 152 introduced into a horizontal tubular stainless-steel reactor. The reactor temperature 153 was raised by 10°C min ⁻¹ until a target temperature of 800 °C was reached. The 154 temperature was maintained for 2 hours, with continuous flow of nitrogen gas at a rate 155 of 5 mL s ⁻¹. 156

157 **2.9 Methylene blue adsorption tests**

Prior to evaluating the adsorption of activated and non-activated biochar, kinetic experiments were used to determine time of equilibrium. After the activation process, biochar can exhibit different behaviors as a function of adsorbate/adsorbent contact time, concentration and amount of adsorbent, especially after K₂CO₃ activation. As a result, the parameters used for the kinetic models were different for each biochar activation processes.

Activated biochar was left in contact with 4 mL of methylene blue dye solution at concentrations ranging from 10 to 450 mg L ⁻¹ (details for each activation process are described in the subsequent section). Methylene blue solutions were prepared by diluting the stock solution in distilled water prior to each adsorption test.

168 Tests were performed in 5 mL screw-cap glass vials covered with aluminum foil to 169 prevent photodegradation. The samples were placed on a Heidolph shaker under

constant stirring at 700 rpm at room temperature (25 °C), for periods ranging from 30
min to 24h depending on the material.

The samples were then centrifuged and the concentration of the remaining adsorbate (methylene blue dye) was determined using a UV-Vis spectrophotometer (Cary-60, Agilent Technologies) at a wavelength of 665 nm, which corresponds to the is wavelength of peak absorption for the methylene blue dye. After measurements were taken, the concentration was calculated using the calibration curve (0 to 10 mg L⁻¹) of methylene blue according to the Lambert-Beer law.

Kinetic study: 5 mg of K₂CO₃-activated biochar and 10 mg each of non-activated and
NaOH-activated biochar were used in this experiment. The following parameters were
defined for each material:

a) For non-activated and NaOH-activated biochar, concentrations of 10, 35, 45 mg L⁻¹
 ¹ were used, and samples were collected at 60, 120, 240, 300, 720 and 1440 min.

b) For inert atmosphere (N₂) and potassium carbonate-activated biochar,
concentrations of 350, 400, 450 mg L ⁻¹ were used and samples were collected at 30,
60, 90, 120, 150, 180, 380 and 480 min.

Equilibrium study: the study of equilibrium adsorption isotherms was conducted according to the results of kinetic studies, after enough time of contact had passed, for the biochar and methylene blue solutions to reach equilibrium.

For non-activated and NaOH-activated biochar, solutions with concentrations of 15, 25, 35, 45, 55 and 75 mg L ⁻¹ were stirred at 700 rpm for 720 min (selected equilibrium time determined from the kinetic studies). For K₂CO₃-activated biochar, the concentrations used were 250, 300, 350, 400 and 450, and the equilibrium time was 180 min. Residual adsorbate concentrations for each material were also analyzed by UV-Vis spectrophotometer.

The maximum adsorption capacity was calculated using **Equation (1)**, with the concentration calculated after equilibrium time. These results were adjusted to Langmuir model using **Equation (2)**, and the Freundlich model using **Equation (3)**, [17].

$$Q_e = \frac{(C_0 - C_e)V}{m} \tag{1}$$

199

$$Q_e = \frac{Q_0 K_L C_e}{1 + K_L + C_e} \tag{2}$$

200

$$\frac{Ce}{Qe} = \frac{1}{K_L Q_0} + \frac{Ce}{Q_0} \tag{3}$$

201 **2.10 Analysis of bio-oil by GC×GC/TOFMS**

Bio-oil samples were analyzed using a GC × GC system (Agilent Technologies, Palo Alto, CA, EUA) coupled to a Pegasus IV (Leco, St. Joseph, MI, USA) time-of-flight mass spectrometer. The analysis was performed on a 60 m, 250 μ m i.d. DB-5 (95% dimethylpolysiloxane and 5 % phenyl groups) capillary column with 0.25 μ m film thickness and a 2.5m, 180 μ m i.d. a DB-17-ms (50 % phenyl and 50 % dimethylpolysiloxane) column with 0.18 μ m film thickness. Both were purchased from J & W Scientific (Folsom, CA, USA).

After optimization, the parameters used for these analyses were as follows: 280 °C injector and ion source temperature; 1 μ L splitless injection volume and 10 mg mL⁻¹ sample concentration. Oven heating started at 60 °C for 1 min, and increased to 280 °C at a rate of 3 °C min ⁻¹. A second oven was kept at 5 °C above the first. The carrier gas was helium at a flow rate of 1 mL min ⁻¹. The modulation period was 8 s with a 0.6 s hot pulse.

The mass spectrometer operated in a range of 45–400 Da and recorded 100 scans per second (100 Hz). The retention index (LTPRI) for each compound was calculated according to Van den Dool & Kratz [18] based on the retention times of a standard mixture of linear n-alkanes (C6-C30). The experimental retention index was then compared with data from the NIST Mass Spectral Library.

The concentration of peaks was estimated based on the analysis of relative peak 221 222 areas. Each compound in the bio-oil was evaluated according to its functional group and family, allowing for the tentative identification of all chemical classes present in 223 bio-oil, though the response factor of each compound was not taken into account. 224 225 Nevertheless, considering the large number of compounds identified and the impossibility of obtaining analytical standards for each compound, this procedure can 226 be used to estimate the percent area of major peaks, to compare different samples 227 228 and screen for possible changes. Further details on this approach are described elsewhere [10,19-21]. 229

230 3. RESULTS AND DISCUSSION

The original biomass and the products of the pyrolysis (bio-oil and biochar-activated and non-activated) were analyzed and characterized by several techniques. The results and discussion for the Thermal gravimetric analysis, Proximate and Elemental analysis, Infrared Spectroscopy (FTIR), Scanning Electron Microscope (SEM), N₂-BET Adsorption Isotherms and Determination of pH at the point of zero charge (pHPzc) can be viewed in the **Supplementary Material Part A**.

237 **3.1 Thermal gravimetric analysis**

The cottonseed demonstrated a high content of volatile and semi-volatile matter (~75 %), indicating that thermal degradation may be an effective method of extracting compounds of interest from this type of biomass. The deconvolution of the derivative (DTG) curve helped define the temperature ranges with the highest percentage of mass loss for each of the main components of the cottonseed (cellulose, hemicellulose and lignin).

244 3.2 Pyrolysis of cotton seed

The mass yields of the products obtained in cottonseed pyrolysis process were calculated as shown in **Figure 1**.



Figure 1: Effects of temperature on yields of cottonseed pyrolysis products.



The maximum bio-oil yield of 20 % was obtained at 550 °C. Above this temperature, the bio-oil yield decreased to approximately 16 %. This can be predominantly attributed to secondary reactions (breakdown and rearrangement) of pyrolysis vapors at high temperatures [22,23].

Previous studies of cottonseed biomass do not separate the aqueous phase from the 253 254 bio-oil as was done in this study, and refer to the entire liquid fraction as bio-oil, resulting in much higher yields than observed in this paper. The application of this 255 approach to the present findings would result in crude bio-oil (including the aqueous 256 phase) yields of 48 % (at 450 °C), 49 % (at 500 °C), 48 % (at 550 °C), 47 % (at 600 257 °C) and 55 % (at 650 °C). These values are in line with previous literature, but as they 258 include large amounts of water, they are only useful for comparison purposes [24-26]. 259 Apaydin-Varol et al., [25] performed the fast pyrolysis of cottonseed in order to define 260 process variables such as temperature (400 – 700 °C), heating rate (5 – 700 °C min⁻ 261 ¹) and nitrogen flow rate (100 - 800 cm 3 min $^{-1}$); the maximum bio-oil yield obtained by 262 these authors was ~ 49 % at a temperature of 500 °C under a nitrogen flow of 200 cm³ 263 min ⁻¹ and a heating rate of 300 °C min ⁻¹. 264

According to several authors, maximum bio-oil yield occurs between 350 and 500 °C since pyrolysis reactions occur at different temperatures, supporting the hypothesis that higher temperatures cause the decomposition of molecules in the solid state, producing smaller molecules and a subsequent increase in gas yield [2,24].

Biochar yield decreased from 32 % to 24.2 % when the temperature increased from 450 °C to 650 °C. This reflects the severe decomposition of organic molecules at higher temperatures, increasing the yield of gaseous product at the expense of the biochar ratio [22,27]. Similar observations of higher biochar yields at lower temperature have been recorded by other researchers for various types of biomass [28], including crambe seeds [29], spent coffee grounds [27], sorghum [30], cotton stalk [26] and wood [31].

At higher temperatures, the gas yield increases due to the secondary reactions of pyrolysis vapors within the reactor, while the decrease in char yield decreased is attributed to the primary decomposition of the feedstock and the secondary decomposition of char at higher temperatures. During secondary decomposition, noncondensable gas products also contribute to the elevated gas yields which occur with increasing temperature [5,24].

282 The aqueous fraction of pyrolysis oil comes from the moisture in the biomass and the dehydration of organic compounds (especially cellulose and hemicellulose) during the 283 284 pyrolysis process [5,32]. The main degradation products of cellulose and hemicellulose are water-soluble. A fraction of the water yield forms a separate aqueous phase while 285 the rest is dispersed into oil. The water-insoluble fraction of the oil consists mainly of 286 287 lignin-derived materials. The recovery of organic compounds from the aqueous phase may be of interest as a potential source of hydrogen, alkanes and polyols [33,34]. The 288 water content of the bio-oil ranged from approximately 28 to 38 %, and was highest at 289 290 650 °C.

Longer residence times in the pyrolysis reactor can lead to further breakdown (dehydration, decarboxylation and condensation) of these compounds, increasing the production of water and non-condensable with smaller molecules such as carbonyls and aliphatic hydrocarbons [5,23] affecting the yield and promoting repolymerization reactions. A short residence time favors the formation of liquid products (bio-oil) since the rapid passage of organic vapors through the reactor minimizes side reactions and preserves the fatty acids and esters in the bio-oil [22,23].

The results showed that the reaction temperature significantly affected the yield of pyrolysis products. The highest yields recorded were ~ 20 % for bio-oil at 550 °C and $\sim 32\%$ for biochar at 450 °C.

301 **3.3 Analysis by Infrared Spectroscopy (FTIR)**

This analysis confirmed the efficiency of the project by comparing the FTIR of the solids before and after and the pyrolysis with the complete withdrawn of organic compounds from the biomass. On the other hand, the bio-oil shows typical absorptions of OH, COOH, COOR, =CO and alkyl and aromatic chains.

306 **3.4 Scanning Electron Microscope (SEM)**

The results for this characterization allow the best visualization and demonstrated the efficiency of the applied activation methods, especially with K₂CO₃ that allowed for the development of a larger surface area for the biochar.

310 **3.5** N₂ - BET adsorption isotherms

These results suggest that the pyrolytic process used in the present study, along with chemical activation with potassium carbonate, allow for the production of porous material from residual cottonseed.

314 **3.6 Determination of pH at the point of zero charge (pH_{PZC})**

The results of pHPzc measurements of activated biochars suggested that the surface of these materials possesses basic characteristics that may aid in the adsorption of the methylene blue cationic dye in aqueous solutions; adsorption can be favored by ensuring that the solution pH is above 7.

319 3.7 Dye adsorption tests: Methylene blue as adsorbate for activated biochars

320 3.7.1 Adsorption kinetics

Contact time is an important parameter for determining the equilibrium time and kinetics of the adsorption process. The kinetics of an adsorption system describes the rate of solute adsorption, which determines the required residence time for adsorption to occur at the solid-liquid interface. Kinetic studies showed that the equilibrium and maximum adsorption time differed between the materials examined.

326 Adsorption kinetics were investigated at room temperature, in varying concentrations

327 (from 10 to 450 mg L⁻¹) of methylene blue dye. Adsorption time curves are shown in

328 **Figure 2 (a, b, c)**.

Figure 2: Kinetic curve with percentage of methylene blue removal as a function of contact time for
(a) non-activated biochar and (b) NaOH-activated biochar at concentrations of 10, 35 and 45 mg
L⁻¹ and for (c) K₂CO₃-activated biochar at concentrations of 350, 400 and 450 mg L⁻¹. The other
conditions used for these experiments are described in experimental section.



The kinetic analysis of non-activated biochar (**Figure 2-a**) showed that the system reached equilibrium in approximately 300 minutes. However, the material had a very low dye adsorption capacity. At 300 min, the adsorption percentages at concentrations of 35 and 45 mg L ⁻¹ were 68.65% and 46.71 %, respectively, while in the solution with the lowest concentration (10 mg L⁻¹), 88.63% of the dye was adsorbed.

For NaOH-activated biochar (**Figure 2-b**), the adsorption process reached equilibrium in approximately 720 minutes. At this time, 95.55% and 86.71% of the dye had been removed at concentrations of 35 and 45 mg L⁻¹. In the solution with the lowest concentration (10 mg L⁻¹), 99.63% of the dye was removed after in 120 minutes.

The kinetic studies of K₂CO₃- activated biochar (Figure 2-c) revealed that the system 344 with the most concentrated solution (450 mg L⁻¹) reached equilibrium in 180 minutes 345 after which approximately 87% of the dye was adsorbed. This percentage remained 346 347 stable for the remainder of the test period. In solutions with concentrations of 350 and 400 mg L⁻¹, the highest concentration decrease occurs after 60 minutes of contact, and 348 was measured to be 99.55% and 97.69%, respectively. At initial contact, far from the 349 point of equilibrium, kinetics are governed by the rate of reactions on the adsorbent 350 surface; when the system approaches equilibrium, the predominant surface kinetics 351 mechanism shifts to intraparticle diffusion [35]. 352

353 3.7.2 Adsorption isotherms

Adsorption isotherms provide important information about the adsorption process, such as the ratio of adsorbate per unit mass of adsorbent to the distribution of adsorbate molecules at the solid/liquid interface when the system reaches equilibrium. Additionally, isotherms indicate how the adsorbent will effectively interact with the adsorbate and whether the desired process is feasible. As such, the analysis of

359 isothermal data and its adjustment it to different models is an important step toward 360 the identification of a suitable adsorption mode [36,37].

The adsorption capacity of non-activated and activated biochar was tested in the time 361 interval determined in the kinetic studies. Non-activated and NaOH-activated biochar 362 were tested for a period of 720 minutes, while K₂CO₃-activated biochar was tested for 363 180 minutes. Assays were performed with a concentration within the equipment 364 detection range and the experiment was conducted at 25 °C. 365

As observed in the kinetic studies, the amount of adsorbate increased with increasing 366 solute concentration in the liquid phase, and the adsorbate distribution between the 367 368 two phases represented adsorption equilibrium. This equilibrium is usually represented 369 by the amount of solute adsorbed per unit mass of the adsorbent as a function of the concentration of solute in the equilibrium solution [38,39]. 370

371 After the equilibrium data were obtained, it was adjusted to the non-linear isotherm models of Langmuir and Freundlich. Figure 3 (a-c) shows the adsorption isotherms 372 373 along with the adjusted curves for biochar activated with alkaline solution, while **Table** 374 1 presents the constants obtained in the adjustment to each mathematical model.

375 Table 1: Constants obtained for the maximum adsorption capacity of the cottonseed biochar non-activated and activated, adjusted to the mathematical models of Langmuir and Freundlich 376 using as adsorbate the methylene blue dye.

Cotton seed biochar	Constants of Langmuir				Constants of Freudlich		
	K∟(L mg⁻¹)	Q₀(mg g⁻¹)	RL	R ²	n	K _F (mg g⁻¹)	R ²
Non-activated	0.35	7.82	0.17	0.99	1.42	5.29	0,93
Activated with NaOH	0.85	24.87	0.01	0.98	3.38	11.15	0.89
Activated with K ₂ CO ₃	3.75	333.30	5.92×10 ⁻⁴	0.99	11.70	254.74	0.94

The amount of methylene blue adsorbed until the equilibrium time was reached ranged from **2.76** to **9.56** mg g ⁻¹ for non-activated biochar at concentrations of 15 to 75 mg L $^{-1}$, while alkaline activated biochar adsorbed **5.89** to **23.82** mg g ⁻¹ of dye under the same time and concentration conditions.

The highest adsorption capacity was obtained for K₂CO₃-activated biochar, which increased from **199.90** mg g⁻¹ to **332.40** mg g⁻¹ as concentration went from 250 to 450 mg L⁻¹. The rise in adsorption capacity with increased initial concentration is mainly due to the higher number of available molecules for adsorption, but may also be attributed to an increase in mass transfer, promoting greater interaction between adsorbent and dye molecules [40,41].

The isotherms obtained for K₂CO₃-activated biochar (**Figure 3-b**) demonstrate high affinity to the dye and favorable adsorption capacity, with adsorption occurring rapidly and the dye remaining bound to the adsorbent until equilibrium was reached.

The Langmuir model provided the best fit to the experimental data for both activated materials, as can be seen in **Table 1**. The adjustment was confirmed by R² values. This model is assuming that the adsorption process takes place on a homogeneous surface with monolayer adsorption (the atom or molecule of the adsorbent binds to specific sites on the surface of the adsorbent).

The efficiency of the adsorption process, predicted by the dimensionless equilibrium parameter R_L, provides further evidence of the satisfactory adsorption properties of activated biochars. The adsorption nature is considered unfavorable if R_L> 1, linear if R_L = 1, and favorable if 0 < R_L <1. As can be observed in **Table 1**, non-activated biochar had an R_L value of 0.17, while NaOH- and K₂CO₃-activated biochar had values of 0.01 and 5.92 ×10⁻⁴, respectively [42,43].

403

404

Figure 3: Adsorption isotherms construct at 25 °C for (a) non-activated biochar and (b) NaOHactivated biochar after 720 minutes of contact and for (c) K₂CO₃-activated biochar after 180 minutes in contact with methylene blue solutions in different concentrations (Q_{eq} = adsorbed amount and C_{eq} = equilibrium concentration).



The Langmuir (KL) constant refers to the amount of substance adsorbed per unit mass of adsorbent at equilibrium. This value was higher for K₂CO₃-activated biochar than non-activated and NaOH-activated samples, which points to greater interaction between the dye and the surface of K₂CO₃-activated biochar relative to the other samples [17].

The Freundlich constant (K_F) is related to adsorption capacity. However, unlike the Langmuir constant Q_0 , K_F does not inform the maximum removal capacity, since the Freundlich model does not predict surface saturation [36,43].

The constants obtained for the Freundlich model were also analyzed. This model assumes that adsorption occurs on a heterogeneous layer, and includes a parameter (n) which represents surface heterogeneity and adsorption intensity. When n=1, adsorption is linear, while n < 1 suggests that the adsorption process is chemical, while n > 1 indicates physical adsorption.

In the present study, non-activated biochar had an *n* value of 1.42, while the *n* value 423 424 of NaOH-activated biochar was 3.38, and that of K₂CO₃-activated biochar, 11.70. These values are indicative of favorable adsorption, and are compatible with previous 425 literature on methylene blue adsorption, where studies report values in the 1-10 range 426 427 [44,45]. The parameter 1/n indicates the intensity of adsorption and the degree of surface heterogeneity, where values near zero are indicative of more heterogeneous 428 surfaces. The 1/n values for non-activated, NaOH-activated and K2CO3-activated 429 biochar were 0.70, 0.29 and 0.08 respectively. These values provide further evidence 430 431 that the isotherms were favorable, though the Freundlich model was not the best fit to the data. 432

433 Once the experimental data were fitted to the Langmuir isotherm, the values obtained 434 for the Q_0 parameter, which describes the monolayer saturation capacity of the

435 adsorbent, were 7.82 mg g⁻¹ for non-activated biochar, 24.87 mg g⁻¹ for NaOH-436 activated samples and 333.30 mg g⁻¹ for K₂CO₃-activated biochar.

These findings together with the initial characterization of these materials suggest that NaOH activation resulted in basic functional groups on the biochar surface, which enabled methylene blue, a cationic dye, to interact with the material through negatively charged active sites.

Previous studies using this technique [46–48], found that alkaline-activated biochar showed efficient removal of methylene blue dye, though the largest differences in maximum adsorption capacity were obtained for K₂CO₃-activated biochar, whose surface had a higher and more ordered pore density develop during activation at high temperatures and, according to characterization results, presented basic functional groups that interacted with the cationic dye and facilitated the access of molecules to the inside of the pores.

These results suggest that intrinsic characteristics such as total surface area and larger pore volumes play a very important role in the dye removal process. The values obtained in the present study were very promising, since they are close to the maximum adsorption capacity of methylene blue by commercial activated charcoal, which is 388 mg g⁻¹, according to the Langmuir model [49,50].

The maximum adsorption capacity of K₂CO₃-activated cottonseed biochar observed in this study using methylene blue was also higher than that of many other adsorbents described in the literature.

Recently, T. Sarat Chandra et al., [39] investigated the use of defatted algal biomass (DAB) as a non-conventional low cost adsorbent and found the raw, defatted and pretreated DAB had maximum adsorption capacities of 6.0, 7.73 and 7.80 mg g⁻¹, respectively.

A. Nasrullah et al., [40] prepared activated carbon-alginate (AC-alginate) beads by entrapping activated carbon powder from mangosteen fruit peel into calcium-alginate beads, and tested their efficacy at removing methylene blue dye from an aqueous solution. The results showed that the beads had a maximum adsorption capacity of 230 mg g⁻¹. Foo and Hameed [51], prepared activated carbon from date seeds using microwave-assisted KOH activation and obtained an adsorption capacity of 316.11 mg g⁻¹ for methylene blue.

There are no reports of production and activation of cottonseed biochar for use as an adsorbent. Previous studies of this material focused on the production of biochar from cotton stalks. In a study by Deng et al. [52], cotton stalks were chemically activated with KOH (KAc) and K₂CO₃ (KCAc) under microwave irradiation and tested for methylene blue adsorption. Adsorption equilibrium data for both activated carbons were best fitted to the Langmuir isotherm. The maximum adsorption capacities of KAc and KCAc were 294.12 mg g⁻¹ and 285.71 mg g⁻¹, respectively.

The data showed that the activation processes used in the present study were effective at improving the adsorption capacity of biochar for methylene blue dye, especially in the case of K₂CO₃ activation.

However, the differences in adsorption capacity of the materials indicated that the adsorption of organic compounds by carbonaceous materials depends on the physical and chemical properties of biochar. The pore structure influences physical adsorption, with surface area and pore volume emerging as the most important factors. However, the surface chemistry of the materials also has a major effect on the adsorption of organic compounds, through its influence on surface charge, electron density and hydrophilicity.

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485 **3.8 Analysis of bio-oil by GC×GC/TOFMS**

The chromatographic profile of cottonseed bio-oil analyzed by GC×GC/TOFMS can be seen in the two-dimensional diagram in **Figure 4**.

The sample has a complex profile, and the presence of large amounts of oxygenated compounds with very close retention times in a specific region of the chromatogram requires a highly efficient separation technique.

491 **Figure 4:** Two-dimensional color diagram (GC×GC/TOFMS) of the organic fraction of the 492 cottonseed bio-oil.



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A total of 257 compounds were tentatively identified in the cottonseed bio-oil, and 168 were confirmed by linear temperature programmed retention indices (LTPRI), with good agreement (within a range of \pm 20) between calculated values and those in the NIST library. The use of LTPRI allowed for the identification of compounds with greater reliability.

The list of compounds identified by GC×GC/TOFMS is presented in **Table B1** (**Supplementary Material**) and includes the components for which LTPRI values were not located in the literature. These compounds were tentatively identified by comparing the mass spectra with the equipment database and the locations of peaks in dispersionprofiles.

The following classes of compounds were identified in the cottonseed bio-oil: nitrogen 504 505 compounds (56 compounds), phenols (42 compounds), hydrocarbons (90 compounds), acids (10 compounds) and ketones (24 compounds). The comparison of 506 relative peak areas revealed that the most abundant compounds in the bio-oil sample 507 were phenol (9.2%), 3-methyl phenol (5.2%) and Indol (4.7%). The yield from 508 mechanical extraction of cottonseed oil is, on average, 17-24%; this value is 509 considered low relative to other raw materials such as canola, sunflower and castor 510 bean, whose oil yield ranges from 30 to 50% [8,9]. The cotton oil transesterification 511 has generated significant interest, however this method involves mainly of extraction, 512 methanolysis and purification, which are time consuming and require expensive 513 chemicals [25]. Fast pyrolysis is a single-step thermochemical conversion method, 514 which is also less expensive than other techniques, making it an attractive alternative 515 for the production of fuel products. Figure 5-c shows the major compounds 516 (concentration > 1 %) found in the bio-oil. 517

Figure 5: Distribution of classes of compounds in the bio-oil of pyrolysis of CS at 550 °C, sorted according to the relative percent area (a) number of compounds (b) and (c) major compounds



521 (conc. % > 1.0 %).

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The analysis of relative areas showed that there is no direct relationship between the 523 524 number of peaks in each class and the area occupied by them. This is especially clear in the case of hydrocarbons and acids. However, for oxygenated compounds such as 525 phenols and ketones, as well as nitrogenous ones, total area is more evenly distributed 526 among the peaks. This is most evident in Figure 5 (a, b), which compares the relative 527 peak area (total percentage area occupied by the compounds in each class) between 528 529 qualitatively identified hydrocarbons, phenols, ketones and other compounds (number of peaks present in each class) 530

531 **3.9 Dispersion Plots**

Dispersion plots illustrate the spatial distribution of different compounds according to 532 molecular weight, number of substituents and number of branches. In a complex 533 sample, compounds with similar structures elute in the same areas of the two-534 dimensional space, allowing for an analysis of the spatial distribution of chemical 535 classes. This method sheds light on the global composition of bio-oil and contributes 536 to qualitative analysis. Dispersion curves were plotted on Microsoft Excel[™] using the 537 538 retention times for each identified peak. The use of dispersion curves facilitates the analysis of spatial distributions, as observed in previous studies conducted by our 539 research group [20,53]. 540

Figure 6 (a-d) shows the spatial distribution of chemical classes in the bio-oil sample. A clear distribution trend for some chemical subclasses can be observed in the 2D separation space.

Figure 6. Dispersion Graphics for the chemical classes on the organic fractions of the bio-oil from
cotton seed: (a) oxygenated compounds; (b) phenols; (c) N-compounds; (d) hydrocarbons (e) acids
and esters.



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549 Some phenols, ketones, aldehydes and nitrogen compounds show a clear separation 550 in the 2D due to significant polarity differences, while acids, esters and hydrocarbons 551 demonstrated efficient separation in the 1D.

552 Phenolic subclasses (Figure 6 (b)) (alkyl, methoxy and benzenediols) can be seen in 553 different parts of the separation space. Alkyl phenols are distributed according to the number of carbons in the side chain (from phenol C₀ to phenols with 1, 2 and 3 carbons
in the side chain). Methoxy phenols and benzenediols have longer elution times in the
second dimension as a result of their high polarity relative to alkyl phenols.

557 Phenolic compounds, mainly derived from lignin degradation [2,54], are some of the 558 most abundant in bio-oil. This class of compounds has great importance and is widely 559 used for industrial purposes. As such, cottonseed bio-oil appears to be an 560 economically viable source of phenols.

As for the applications of the identified compounds, phenol may be of great importance as a substitute for petrochemical phenol in the production of phenolic resins (phenolformaldehyde) in the electronic and wood treatment industries. Some studies report that up to 50% of phenols can be replaced bio-oil equivalents with no damage to the quality of the resin [54,55].

As expected, the spatial arrangement of hydrocarbons was determined by the number of carbons in the side chain, as can be seen in **Figure 6 (d)**. The separation of hydrocarbons occurs mainly in the first dimension, and is also related to the increase of carbon atoms in the chain. Paraffins have carbon chains from C₁₀ to C₂₉, much like olefins and monoaromatics. Polyaromatic hydrocarbons have longer elution times in the second dimension (3.8 and 5.3 s).

The separation of carboxylic acids and esters **Figure 6** (e), occurs mainly in 1D due to their volatility, but the presence of these compounds was not as evident in this bio-oil sample. The significant decrease in oxygen content of bio-oil compared with the original feedstock is important. This decrease was confirmed by elemental analysis of the raw material and the bio-oil produced [55].

577 The composition of cottonseed bio-oil was distinct from that of bio-oils produced from 578 different biomasses [56], such as spent coffee grounds [10], mango seed [53], rice

husk and peach pit [57]. This peculiar characteristic is the richness of nitrogen compounds. The main distinguishing feature of cottonseed bio-oil was the abundance of nitrogen compounds. Proteins account for up to ~25% of cottonseed weight, and the presence of protein results in a higher content of nitrogen compounds in the bio-oil produced [8]. **Figure 6 (c)** shows that the nitrogenous compounds present in bio-oil do not have an ordered spatial organization.

The 56 peaks of nitrogen compounds found in the bio-oil include pyridines, pyrroles, indoles, amines and nitriles. Pyridine has pharmaceutical significance and is rarely found in its pure form. Industrially it is obtained from coal tar but is only available in small amounts (approximately 0.1%).

The literature provides limited information on the specific and global composition of the 589 bio-oil produced by cottonseed pyrolysis. Apaydin-Varol et al., [25] performed fast 590 591 pyrolysis of cottonseed and tentatively identified the chemical composition of the resulting bio-oil using GC / MS. The major compounds identified were: palmitic acid, 592 593 linoleic acid, phenolic (4-methyl phenol) and nitrogenous compounds (1H Indol) and mostly straight chain hydrocarbons distributed in the range of C₁₂ to C₂₈, which is 594 similar to findings obtained in transport fuels. S. Seal et al., [58] performed the chemical 595 596 characterization of the liquid product of cottonseed pyrolysis using GC-MS, identifying n-hexadecanoic acid as the most abundant compound, and noting the presence of a 597 large number of organic compounds, such as hydrocarbons, organic acids, ketone and 598 599 phenol, in addition to a minor amount of alkanes (>C10).

The use of GC×GC/TOF-MS associated with LTPRI software tools and dispersion plots performed in the present study, allowed for a comprehensive characterization of the global composition of cottonseed bio-oil. This classification increases the accuracy of compound identification and the reliability of the results, which are crucial for further

studies and for the development of biorefineries to process cottonseed residue for usein several industrial applications.

606 **4. CONCLUSIONS**

607 Cottonseed bio-oil demonstrated a richness of nitrogen compounds which is not often 608 found in other types of bio-oil. This characteristic enhances the potential of this bio-oil 609 as a substitute for fossil fuel compounds that are difficult to obtain industrially. 610 Additionally, one of the major groups of compounds identified in the bio-oil were 611 phenolic compounds, which have important applications in a variety of fields, with the 612 major compound (phenol) being especially used in the production of phenolic resins.

In addition to nitrogenous and phenolic compounds, cottonseed bio-oil contained large amounts of fatty acids and other organic compounds that could be transformed into important raw materials. Fatty acids are useful for the production of biodiesel, while hydrocarbons may be used as alternative fuels to petroleum derivatives and oxygenated compounds (ketones, phenols) can be used as raw materials for fine chemical engineering in the pharmaceutical, food and chemical industry.

The biochar obtained in the pyrolysis process and submitted to chemical activation (K₂CO₃) possesses great potential as an adsorbent and generated surprising results with regards to its high surface area and adsorption capacity. The application of mathematical models to the adsorption results showed that the Langmuir isotherm provided the best fit to the experimental data of all tests, indicating that the adsorption of methylene blue dye by the activated biochar occurs mainly at the monolayer surface.

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626 **5. ACKNOWLEDGEMENTS**

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631 6. SUPPLEMENTARY MATERIAL DATA

The results and discussion for the Thermal gravimetric analysis, Proximate and Elemental analysis, Infrared Spectroscopy (FTIR), Scanning Electron Microscope (SEM), N₂-BET Adsorption Isotherms and Determination of pH at the point of zero charge (pHpzc) with the respective references included, can be viewed in the **Supplementary Material Part A** and the complete list of compounds identified by GC×GC/TOFMS is presented in **Table B1 (Supplementary Material – Part B)**.

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