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Jointly Mapping Hydraulic Conductivity and Porosity by Assimilating Concentration Data via Ensemble Kalman Filter

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Abstract

Real-time data from on-line sensors offer the possibility to update environmental simulation models in real-time. Information from on-line sensors concerning contaminant concentrations in groundwater allow for the real-time characterization and control of a contaminant plume. In this paper it is proposed to use the CPU-efficient Ensemble Kalman Filter (EnKF) method, a data assimilation algorithm, for jointly updating the flow and transport parameters (hydraulic conductivity and porosity) and state variables (piezometric head and concentration) of a groundwater flow and contaminant transport problem. A synthetic experiment is used to demonstrate the capability of the EnKF to estimate hydraulic conductivity and porosity by assimilating dynamic head and multiple concentration data in a transient flow and transport model. In this work the worth of hydraulic conductivity, porosity, piezometric head, and concentration data is analyzed in the context of aquifer characterization and prediction uncertainty reduction. The results indicate that the characterization of the hydraulic conductivity and porosity fields is continuously improved as more data are assimilated. Also, groundwater flow and mass transport predictions are improved as more and different types of data are assimilated. The beneficial impact of accounting for multiple concentration data is patent. *Key words:* Data assimilation; stochastic transport; ensemble Kalman filter; multiple concentration data; hydraulic conductivity and porosity; heterogeneity

1. Introduction

During the last several decades numerical simulation is routinely utilized to evaluate the groundwater resources and predict the fate of contaminant plumes. The adequate characterization of spatially distributed hydrogeological parameters like hydraulic conductivity and porosity plays an important role in groundwater

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5 flow and transport simulations. However, due to the scarcity of measurements in combination with the large
6 spatial heterogeneity it is not trivial how to characterize the spatial distribution of the mentioned parameters,
7 and, consequently, groundwater flow and transport predictions call for an uncertainty assessment. Inverse
8 modeling is often used to reduce model uncertainty by jointly conditioning on hard data (e.g., hydraulic
9 conductivity and porosity) and indirect data (e.g., the observed state information, such as piezometric heads,
10 concentrations and temperatures) to characterize the spatial variation of hydrogeological parameters. The
11 issue of how to condition on the direct measurements has been extensively investigated in the geostatistical
12 literature (e.g., Journel, 1974; Gómez-Hernández and Srivastava, 1990; Strebelle, 2002). Likewise, inverse
13 modeling, i.e., conditioning to indirect data, has been reviewed in the literature (e.g., Yeh, 1986; McLaughlin
14 and Townley, 1996; Zimmerman et al., 1998; Carrera et al., 2005; Hendricks Franssen et al., 2009). Commonly,
15 inverse methods define an objective function that includes the mismatch between calculated and observed
16 state values, as well as the perturbation of the initial parameter estimates. This objective function is
17 minimized by an optimization approach. Examples are the self-calibration method (Sahuquillo et al., 1992;
18 Gómez-Hernández et al., 1997; Capilla et al., 1999; Wen et al., 2002; Hendricks Franssen et al., 2003), the
19 pilot point method (Ramarao et al., 1995; LaVenue et al., 1995; Alcolea et al., 2006), the Markov chain
20 Monte Carlo method (Oliver et al., 1997), and the gradual deformation method (Hu, 2000; Capilla and
21 Llopis-Albert, 2009).

22 Albeit the abundant literature on inverse conditioning of conductivities to piezometric head, only a few
23 works have paid attention on jointly conditioning on head and concentration data to improve the characteri-
24 zation of multiple hydrogeological parameters. Medina and Carrera (1996) extended the maximum likelihood
25 approach (Carrera and Neuman, 1986) to condition on concentration data for a better characterization of
26 zoned hydraulic conductivity maps. The main shortcoming of this approach is that the small-scale hetero-
27 geneity is ignored due to the estimation of hydraulic conductivity for larger zones. Hendricks Franssen et al.
28 (1999) calibrated both hydraulic conductivity and storativity by conditioning to transient head data using
29 the self-calibration method (Gómez-Hernández et al., 1997). More recently, Hendricks Franssen et al. (2003)
30 further extended the self-calibration method to calibrate hydraulic conductivity by conditioning on piezo-
31 metric head and concentration data. Huang et al. (2004) also employed the self-calibration method to jointly
32 identify hydraulic conductivity and sorption partitioning coefficient by conditioning on tracer breakthrough
33 data. Fu and Gómez-Hernández (2009) employed the block Markov chain Monte Carlo method to calibrate
34 conductivity by jointly conditioning to head and travel time data. Llopis-Albert and Capilla (2009) utilized
35 the gradual deformation method to estimate the conductivity by incorporating head, concentration and travel

36 time data. Schwede and Cirpka (2009) used the quasi-linear geostatistical approach of Kitanidis (1995) to
37 estimate conductivity by conditioning on steady-state concentration measurements. Barnhart et al. (2010)
38 employed PEST (Doherty, 2004), a model-independent nonlinear parameter estimation program, to calibrate
39 hydraulic conductivity by conditioning to concentration data collected from wireless sensor networks. These
40 approaches are able to generate multiple equally-likely parameter fields conditional to static and dynamic
41 measurements, thus capable of depicting small-scale variability of hydraulic conductivity. However, the main
42 shortcoming of those methods is that they are CPU-intensive; these methods require running the forward
43 model multiple times during the iterative optimization process of each realization.

44 The Ensemble Kalman Filter (EnKF) (Burgers et al., 1998; Evensen, 2003), based on the sequential
45 Bayesian updating rule, can be used to obtain results similar to those obtained by Monte-Carlo (MC)
46 type inverse methods but with reduced CPU time (see section 2.2), and it is also flexible to incorporate
47 multiple sources of uncertainty. Hendricks Franssen and Kinzelbach (2009) carried out a synthetic exercise
48 and demonstrated that EnKF needs around a factor of 80 less CPU time than the self-calibration method
49 to attain similar results. The EnKF can also handle data from on-line sensors that become available in
50 real-time and assimilate them into an on-line model. The traditional inverse methods are not well suited to
51 assimilate information that becomes available in real-time. EnKF provides an ensemble of updated stochastic
52 realizations which can be used for uncertainty analysis.

53 The EnKF is increasingly applied, in atmospheric sciences, land-atmosphere interaction, petroleum en-
54 gineering and hydrogeology (e.g. Anderson, 2001; Reichle et al., 2002; Wen and Chen, 2005; Chen and
55 Zhang, 2006; Nowak, 2009; Hendricks Franssen et al., 2011; Zhou et al., 2011b). In atmospheric sciences
56 or land surface models in general only the model states are updated, whereas in petroleum engineering and
57 hydrogeology both system parameters and state variables are commonly addressed (Naevdal et al., 2005).

58 The EnKF has been successfully applied to assimilate dynamic piezometric head data to improve model
59 predictions (e.g., Chen and Zhang, 2006; Hendricks Franssen and Kinzelbach, 2008; Sun et al., 2009; Li
60 et al., 2011c; Zhou et al., 2011a). With regard to assimilating concentration data, Huang et al. (2008)
61 conducted a synthetic experiment and calibrated hydraulic conductivity fields by assimilating piezometric
62 head and concentration data. In their experiment, flow was at steady-state. Liu et al. (2008) estimated
63 multiple parameters (i.e., hydraulic conductivity, dispersivities, mobile/immobile porosities) by assimilating
64 piezometric head and concentration data in the steady-state flow model for the MADE site. It is worth to
65 note that they used constant values as the prior estimates for the mentioned parameters, and perturbed
66 the parameters, by assimilating observation data via EnKF, to yield heterogeneous fields. Our aim is to

67 quantify the uncertainty of parameters and states starting with heterogeneous fields by conditioning on the
68 direct measurements. Schöniger et al. (2011) assimilated normal-score transformed concentration data to
69 calibrate hydraulic conductivities. They concluded that the improvement by the normal-score transformation
70 (as compared with the classical EnKF, which uses untransformed data) is limited because after univariate
71 normal transformation of the state variable, the concentration distribution is far from multi-Gaussian.

72 In comparison with the effort devoted to characterize the spatial variability of hydraulic conductivity by
73 conditioning state information, less attention has been paid to identifying the spatial variability of porosity,
74 probably due to its relatively small spatial variability ranging from 0.1 to 0.55 in unconsolidated granular
75 aquifers (Freeze and Cherry, 1979). Additionally, various authors (e.g., Hassan, 2001; Riva et al., 2008;
76 Hu et al., 2009; Jiang et al., 2010) have demonstrated (both in synthetic examples and real aquifers) the
77 significance of accounting for the heterogeneity of porosity on predictions of solute movement.

78 We will demonstrate the capability of the EnKF to jointly map the hydraulic conductivity and porosity
79 fields by assimilating dynamic piezometric head and multiple concentration data. Few studies have considered
80 the conditioning with help of both multiple concentration data and dynamic piezometric head data to
81 characterize unknown parameters. Also, to the best of our knowledge, this is the first work proposing the
82 joint estimation of spatially distributed hydraulic conductivity and porosity fields in hydrogeology.

83 The remaining of this paper is organized as follows. We first summarize in section 2 the mathematical
84 framework of the EnKF and discuss the jointly mapping of hydraulic conductivity and porosity by assimilating
85 multiple concentration data. In section 3, a synthetic example is used to demonstrate the effectiveness
86 of the EnKF. The paper ends with summary and conclusions in section 4.

87 **2. Data Assimilation with the EnKF**

88 First, the flow and transport equations (i.e, the transfer functions) will be presented, and then the
89 algorithm of EnKF is introduced with emphasis on the assimilation of concentration data.

90 *2.1. Flow and Transport Equations*

91 The well known flow equation of an incompressible or slightly compressible fluid in saturated porous
92 media can be expressed by combining Darcy's Law and the continuity equation (Bear, 1972; Freeze and
93 Cherry, 1979):

$$\nabla \cdot (K \nabla h) = S \frac{\partial h}{\partial t} + W \quad (1)$$

94 where K is the hydraulic conductivity [LT^{-1}] (which, without loss of generality, will be considered as a
 95 scalar at the characterization scale), h is the piezometric head [L]; W represents sources or sinks [L^3T^{-1}];
 96 S is the specific storage coefficient [L^{-1}]; t is the time [T]; $\nabla \cdot = (\partial/\partial x + \partial/\partial y + \partial/\partial z)$ is the divergence
 97 operator of a vector field, and $\nabla = (\partial/\partial x, \partial/\partial y, \partial/\partial z)^T$ is the gradient operator of a scalar field.

98 Solute transport with linear equilibrium adsorption is governed by the following differential equation
 99 (Bear, 1972; Freeze and Cherry, 1979):

$$\phi R \frac{\partial c}{\partial t} = \nabla(\phi \mathbf{D} \cdot \nabla c) - \nabla \mathbf{q} c \quad (2)$$

100 where c is solute concentration of solute in the water phase [ML^{-3}]; ϕ is the porosity [dimensionless]; \mathbf{D}
 101 is the local hydrodynamic dispersion coefficient tensor [L^2T^{-1}], with eigenvalues associated with principal
 102 axes parallel and perpendicular to the direction of flow, defined as $D_I = \alpha_L |\mathbf{q}| + D_m$, $D_{II} = \alpha_T |\mathbf{q}| + D_m$,
 103 $D_{III} = \alpha_T |\mathbf{q}| + D_m$ (α_L and α_T are respectively the longitudinal and transverse pore-scale dispersivity; D_m
 104 is the molecular diffusion coefficient set to zero in this study, and \mathbf{q} is the Darcy velocity given by $\mathbf{q} = -K \nabla h$
 105 [LT^{-1}]); R is retardation factor expressed by $R = 1 + \rho_b K_d / \phi$ (ρ_b is the bulk density of soil; K_d is the
 106 distribution coefficient).

107 2.2. Ensemble Kalman Filter

108 Extensive descriptions of EnKF and its algorithm can be found in Burgers et al. (1998) and Evensen
 109 (2003). Here, we mainly focus on the use of EnKF with updating both parameters (i.e., hydraulic conductivity
 110 and porosity) and states (i.e., piezometric head and concentration). It involves a forecast step and an analysis
 111 step, after the generation of the initial ensemble of hydraulic conductivity and porosity realizations.

- 112 • Step 1: Forecast model. The flow equation (1) or transport equation (2) is solved, i.e.,

$$\mathbf{Y}_k = f(\mathbf{X}_{k-1}, \mathbf{Y}_{k-1}), \quad (3)$$

113 where \mathbf{Y}_k is the state of the system (piezometric heads and/or concentration data) at time step
 114 t_k , f represents the groundwater flow and transport model (including boundary conditions, external
 115 stresses, and known parameters), and \mathbf{X}_{k-1} denotes the model parameters (hydraulic conductivity

116

and/or porosity) after the latests update at time t_{k-1} . Specifically, \mathbf{X} and \mathbf{Y} are expressed as:

$$\left\{ \begin{array}{ll} \text{Case A : } \mathbf{X} = [\ln K]^T & \mathbf{Y} = [h]^T, \quad \text{if only } h \text{ data are available.} \\ \text{Case B : } \mathbf{X} = [\ln K, \phi]^T & \mathbf{Y} = [c]^T, \quad \text{if only } c \text{ data are available.} \\ \text{Case C : } \mathbf{X} = [\ln K, \phi]^T & \mathbf{Y} = [h, c]^T, \quad \text{if } h \text{ and } c \text{ data are available.} \end{array} \right. \quad (4)$$

117

- Step 2: Analysis step. Using the observed dynamic piezometric head and concentration data, the model parameters are updated as follows:

118

1. Build the joint vector Ψ_k , which includes the parameters (\mathbf{X}) and the forecasted state values (\mathbf{Y}).

119

This vector can be split into as many members as there are realizations in the ensemble, with

120

$$\Psi_{k,j} = \begin{bmatrix} \mathbf{X} \\ \mathbf{Y} \end{bmatrix}_{k,j}, \quad (5)$$

121

being the j^{th} ensemble member of the augmented state vector at time t_k .

122

As an example, if the number of discretization blocks in the domain is N_k and we are in case C,

123

i.e., updating both $\ln K$ and ϕ using both h and c data, the dimension of vector Ψ_k will be $4 \times N_k$.

124

2. The joint vector is updated, realization by realization, by assimilating the observations (\mathbf{Y}_k^{obs}):

$$\Psi_{k,j}^a = \Psi_{k,j}^f + \mathbf{G}_k \left(\mathbf{Y}_{k,j}^{obs} + \epsilon - \mathbf{H} \Psi_{k,j}^f \right), \quad (6)$$

125

where the superscripts a and f denote analysis and forecast, respectively; ϵ is a random observation

126

error vector; \mathbf{H} is a linear operator that interpolates the forecasted heads to the measurement

127

locations, and, in our case, is composed of 0's and 1's since we assume that measurements are

128

taken at block centers. Therefore, equation (6) can be expressed as:

$$\Psi_{k,j}^a = \Psi_{k,j}^f + \mathbf{G}_k \left(\mathbf{Y}_{k,j}^{obs} + \epsilon - \mathbf{Y}_{k,j}^f \right), \quad (7)$$

129

where the Kalman gain \mathbf{G}_k is given by:

$$\mathbf{G}_k = \mathbf{P}_k^f \mathbf{H}^T \left(\mathbf{H} \mathbf{P}_k^f \mathbf{H}^T + \mathbf{R}_k \right)^{-1}, \quad (8)$$

130

where \mathbf{R}_k is the measurement error covariance matrix, and \mathbf{P}_k^f contains the covariances between

131 the different components of the state vector. \mathbf{P}_k^f can be estimated from the ensemble of forecasted
 132 results as:

$$\begin{aligned} \mathbf{P}_k^f &\approx E \left[\left(\Psi_{k,j}^f - \overline{\Psi}_{k,j}^f \right) \left(\Psi_{k,j}^f - \overline{\Psi}_{k,j}^f \right)^T \right] \\ &\approx \sum_{j=1}^{N_e} \frac{\left(\Psi_{k,j}^f - \overline{\Psi}_{k,j}^f \right) \left(\Psi_{k,j}^f - \overline{\Psi}_{k,j}^f \right)^T}{N_e}, \end{aligned} \quad (9)$$

133 where N_e is the number of realizations in the ensemble, and the overbar denotes average over the
 134 ensemble.

135 In the implementation of the algorithm, it is not necessary to calculate explicitly the full covariance
 136 matrix \mathbf{P}_k^f (of dimensions $(4 \times N_k) \times (4 \times N_k)$ for case C). The matrix \mathbf{H} is very sparse, and,
 137 consequently, the matrices $\mathbf{P}_k^f \mathbf{H}^T$ and $\mathbf{H} \mathbf{P}_k^f \mathbf{H}^T$ can be computed directly at a strongly reduced
 138 CPU cost.

- 139 • Step 3: Loop back. The updated states become the current states and the forecast-analysis loop is
 140 started again.

141 When the number of observation locations used in the assimilation step is not very large, the computa-
 142 tional cost of calculating the covariances is limited. The main cost is related with the forward simulations
 143 for each of the stochastic realizations.

144 During the updating step, the forecasted state variables may have no physical meaning, e.g., negative
 145 concentrations. In our case, we remove negative values resetting them to zero. We have checked that when
 146 this may happen at locations far from the concentration plume and always with small values. This approach
 147 follows the one by Gu and Oliver (2006), who had a similar problem when dealing with water saturation in
 148 a reservoir characterization exercise.

149 The algorithm is implemented in the C software EnKF3D which is used in conjunction with finite-
 150 difference program MODFLOW (Harbaugh et al., 2000), to solve the confined transient flow equation (1),
 151 and the solute transport code MT3DMS (Zheng et al., 1999). MT3DMS uses a third-order total-variation-
 152 diminishing (TVD) solution scheme, to solve the transport equation (2).

153 3. Synthetic Example

154 In this section, a synthetic example will be presented to demonstrate the capability of the EnKF to
 155 calibrate the hydraulic conductivities and porosities by assimilating piezometric head and concentration

156 data. The resulting ensemble of realizations will be used also for uncertainty characterization; in a real-
157 world case study, uncertainty may stem both from the conceptual model (e.g., the boundary conditions,
158 aquifer geometry) and from the parameters. Here, we only consider the uncertainty due to the heterogeneity
159 of hydraulic conductivity and porosity, no conceptual uncertainty is considered.

160 3.1. Experiment Setup

161 3.1.1. Reference Field

162 The reference conductivity and porosity fields are generated using the code GCOSIM3D (Gómez-Hernández
163 and Journel, 1993) over a domain of $250 \text{ m} \times 250 \text{ m} \times 1 \text{ m}$, which is discretized into grid cells of size 5 m
164 by 5 m by 1 m (see Figure 1A and 1C). Here, we assume that the two variables are independent of each
165 other. The parameters of each random function are listed in Table 1. From these reference realizations
166 nine conductivity and nine porosity data are sampled for conditioning purposes. The locations are shown in
167 Figure 1B and 1D.

168 It is assumed that the sampled data have the same support as the grid cell. If the data support would be
169 much smaller than the grid cell size, the additional problem of upscaling must be considered for generating
170 the parameter realizations conditional to the direct measured data (e.g., Li et al., 2011a; Zhou et al., 2010;
171 Li et al., 2011b).

172 The aquifer is assumed to be confined with impermeable boundaries on south and north, prescribed
173 head values on the western boundary and constant flow rate on the eastern boundary (see Figure 2). The
174 prescribed head value is 0 m along the western boundary. The total flow rate through the eastern boundary
175 is $-25 \text{ m}^3/\text{d}$, distributed uniformly along the boundary. The initial head value is 0 m over the entire domain.
176 The total simulation time is 500 days, and this period is discretized into 100 time steps following a geometric
177 sequence of ration 1.05. Specific storage is assumed constant with a value of 0.003 m^{-1} . The simulated
178 dynamic piezometric heads at the observation wells #1 to #9 in Figure 2 are sampled and will be used as
179 assimilating data. The simulated heads at the wells #10 and #11 will be used as validation data.

180 The boundary conditions for the transport model are no-mass flux boundaries on the western, northern,
181 and southern borders of the model. The eastern border is a specified advective mass flux boundary, acting
182 as a line of sinks taking mass out of the aquifer (see Figure 2).

183 The code MODFLOW (Harbaugh et al., 2000) is used to solve the transient groundwater flow equation
184 for the reference field and the pore velocities across the grid cell interfaces are calculated using the porosities
185 in Figure 1C. This velocity field is used as input for solving the solute transport problem with help of the

186 MT3DMS code (Zheng et al., 1999). We only consider advection and dispersion as transport mechanisms
187 with $\alpha_L = 1.0$ m and $\alpha_T = 0.1$ m. Conservative solute is uniformly placed over a line transverse to the
188 groundwater flow at time $t = 0$ (see Figure 2). The source concentration is 900 ppm. To avoid the boundary
189 effect as described by Naff et al. (1998), the contaminant source is separated 20 m from the western boundary
190 and 50 m from the northern and southern boundaries. The plume snapshots at 300, 400 and 500 days will
191 be used here to compare the EnKF solutions with the reference plume maps (see Figures 3A, 3C and 3E).
192 The concentration is measured at 63 wells, uniformly distributed over the domain (see Figure 2). These
193 measured multiple concentration data (see Figures 3B, 3D, 3F) will serve as assimilating data.

194 3.1.2. Scenario Studies

195 Six simulation scenarios are considered for which different types of measurement data are assimilated
196 (see Table 2). Scenario 1 (S1) is an unconditional case. In Scenario 2 (S2) geostatistical simulation (Gómez-
197 Hernández and Journel, 1993) is used in order to condition on the nine measured hydraulic conductivities
198 and the nine porosities shown in Figures 1B and 1D, respectively. For S1 and S2, 500 realizations of hydraulic
199 conductivity and porosity are generated using the same random functions as for the reference fields. Flow
200 and transport are calculated for each of the 500 $\ln K$ - ϕ realization couples, without conditioning to head or
201 concentration data.

202 For scenario 3 (S3) dynamic piezometric head data are used to update the geostatistical realizations
203 conditioned on hydraulic conductivity and porosity data of scenario S2. Piezometric head data from wells
204 #1 to #9 are sequentially assimilated for the first 60 time steps (approximately 67.7 days).

205 In scenarios 4, 5 and 6 (S4, S5, S6) concentration data are assimilated by EnKF, in addition to hydraulic
206 conductivity data and piezometric head data. S4 uses concentration data at 400 days, S5 uses concentration
207 data at 300 and 400 days, and S6 uses concentration data at 300, 400 and 500 days.

208 The piezometric head and concentration data are sampled from the reference simulations without error.
209 However, during the assimilation process it is considered that the data might contain measurement errors
210 and therefore a diagonal error covariance matrix was used, with all non-zero terms equal to 0.0025 m^2 for
211 head data and 0.0025 ppm^2 for concentration data. We note, in practice, the errors for the heads and
212 concentration data would be not the same, and the observation errors would change with the time. From an
213 operational point of view, it is straightforward to integrate them into the assimilation procedure.

214 3.2. Assessment Measures

215 The results for the six scenarios will be analyzed with the help of two metrics:

216 1. The average absolute bias (*AAB*) is a measure of accuracy defined as follows:

$$AAB(X) = \frac{1}{N_b} \sum_{i=1}^{N_b} \frac{1}{N_e} \sum_{r=1}^{N_e} |X_{i,r} - X_{ref,i}| \quad (10)$$

217 where X_i is, either the logconductivity $\ln K$, porosity ϕ , hydraulic head h or concentration c , at location
218 i , $X_{i,r}$ represents its value for realization r , $X_{ref,i}$ is the reference value at location i , N_b is number of
219 nodes, and N_e is the number of realizations in the ensemble (500, in this case).

220 2. The ensemble spread (*AESP*) represents the estimated uncertainty defined as follows:

$$AESP(X) = \left(\frac{1}{N_b} \sum_{i=1}^{N_b} \sigma_{X_i}^2 \right)^{1/2}, \quad (11)$$

221 where $\sigma_{X_i}^2$ is the ensemble variance at location i .

222 The smaller the values for *AAB* and *AESP*, the better the prediction of variable X .

223 3.3. Data Assimilation Results

224 3.3.1. Hydraulic Conductivities and Porosities

225 Figures 4 and 5 show the ensemble mean and variance of the 500 logconductivity realizations for all
226 six scenarios. Figures 6 and 7 show the ensemble mean and variance of the 500 porosity realizations and
227 scenarios. The ensemble mean is used to check whether the main patterns of variability of the parameter
228 are captured. In contrast to the individual realization showing distinctive patterns of high and low values,
229 the ensemble means are smoothed representations of the spatial variability of the parameters. The ensemble
230 variance illustrates how conditioning reduces the differences between the realizations.

231 In scenario 1, with no conditional data, the ensemble mean and variance of $\ln K$ and ϕ are very close to
232 the prior mean and variance. In scenario 2, using 9 conditioning hydraulic conductivities and porosities, the
233 overall spatial patterns are captured, resulting in typical kriging maps. The ensemble variance maps show
234 the typical bull-eye look of kriging maps, with zero variance at the sample locations and increasing variance
235 away from them. The dynamic piezometric head data included in S3 help to capture better the main patterns
236 of hydraulic conductivity with a further reduction of the variance. S4, S5 and S6 also include concentration
237 data for conditioning. The ensemble mean maps better delineate the main patterns of variability, and at the
238 same time, unlike previous scenarios with strongly smoothed representations of $\ln K$, also show some degree
239 of the small-scale variability. The characterization of the main patterns of $\ln K$ improves quite remarkable

240 with the conditioning of concentration data. For the scenario that uses the largest amount of conditioning
 241 data (S6), the patterns in the left upper corner of the area are identified very well, whereas this is not the
 242 case if only hydraulic conductivity, porosity and piezometric heads are used for conditioning. The role of
 243 concentration data on the characterization of porosity is also observable. The main patterns of porosity
 244 are clearer than without conditioning, and closer to the reference distribution. As expected, the ensemble
 245 variance, both for $\ln K$ and ϕ , reduces further in S6 as compared with the other scenarios.

246 From a more quantitative point of view, the calculated two metrics (see Table 3) lead to similar con-
 247 clusions. When the measured hydraulic conductivity, porosity, piezometric head and multiple concentration
 248 data are all used for conditioning (S6), the average absolute bias and the ensemble spread have the smallest
 249 values. More precisely, when conditioning to $\ln K$, the $AAB(\ln K)$ decreases 14% (S2 vs. S1), if we further
 250 condition to piezometric head, $AAB(\ln K)$ further decreases a 10% (S3 vs. S2), and there is an additional
 251 reduction of a 7% when conditioning to concentrations (S6 vs. S3). Likewise $AESP(\ln K)$ goes down 12%
 252 from S1 to S2, an additional 16% from S2 to S3 and 14% more from S3 to S6. Similar results can be observed
 253 when analyzing the evolution of $AAB(\phi)$ and $AESP(\phi)$. The $AAB(\phi)$ shows an 18% reduction as a conse-
 254 quence of conditioning to measured ϕ (S2 vs. S1), and 7% additional reduction related with conditioning to
 255 concentration data (S6 vs. S2) and $AESP(\phi)$ shows a 15% reduction as a consequence of conditioning to
 256 measured ϕ (S2 vs. S1), and 9% additional reduction related with conditioning to concentration data (S6
 257 vs. S2).

258 From these results, we can conclude that: (1) The direct measured hard data play the most important
 259 role to reduce the absolute bias of parameters; (2) The indirect measured head and concentration data reduce
 260 both the absolute bias and ensemble spread; (3) The best characterization of the aquifer in terms of $\ln K$
 261 and ϕ is achieved by combining all the data.

262 3.3.2. Piezometric Heads Reproduction

263 Figure 8 shows the piezometric head evolution at well #2 and #10 for scenarios S1, S2, S3 and S6.
 264 Recall that the piezometric head data continuously collected from well #1 to #9 are used for conditioning,
 265 while wells #10 and #11 are for validation. Figure 8 shows that for S1 uncertainty is largest and that
 266 the uncertainty is reduced for increasing amounts of conditioning data. For S2 the uncertainty is still
 267 considerable, but if piezometric head data are used for conditioning (S3) the conditional well #2 has a good
 268 head reproduction and the control well #10, also shows a large reduction of spread. The measured head
 269 data play a critical role to reduce the uncertainty of predicted heads. The concentration data do not result
 270 in a further improvement of the characterization of hydraulic head since the dynamic heads are already

271 reproduced very well in S3.

272 Table 4 shows the metrics regarding the piezometric head characterization at time $t = 67.7$ days (i.e., the
273 60th time step). The introduction of measured hydraulic conductivities attains around 27% reductions both
274 for the $AAB(h)$ and $AESP(h)$. An additional 66% reduction of $AAB(h)$ and 73% reduction of $AESP(h)$
275 is achieved by conditioning to head data. The reductions of $AAB(h)$ and $AESP(h)$ almost can be ignored
276 when concentration data are used for conditioning in S4, S5 and S6.

277 The main conclusions are: (1) of all the data, the measured piezometric head data are most informative
278 for improving head predictions and reducing the prediction uncertainty; (2) the impact of concentration data
279 for characterizing piezometric head is very small.

280 3.3.3. Concentrations Reproduction

281 Figure 9 to 12 show the ensemble mean and variance of 500 concentration realizations at time 300 and
282 500 days resulting from the transport simulation for all the six scenarios.

283 These ensemble mean maps of concentration for scenario 1 show that even though each realization
284 will have a non-Gaussian plume similar to those in the reference, the random location of high and low
285 concentrations makes that the ensemble mean maps of plume show a Gaussian shape. Introducing the
286 hydraulic conductivity data (S2) rectifies the plume but still does not reproduce the reference. The ensemble
287 mean of the plume is further rectified when the conductivity, porosity and head data are jointly used for
288 conditioning (S3). The reproduction of piezometric heads is very good in S3, but the limited improvement
289 of the plume characterization indicates the importance of further conditioning on concentration data. The
290 results for scenarios S4 and S6 show that conditioning remarkably improves the characterization of the plume.
291 Conditioning to concentration data at $t = 300$ days (S4) also improves strongly the prediction for 500 days
292 (although the concentration data sampled at $t = 500$ days are not used for conditioning in scenario S4).
293 For scenarios S5 and S6 the additional concentration data from $t = 400$ and 500 days improve further the
294 characterization of the plume so that they are very close to the reference plumes.

295 The ensemble variance maps of the concentration fields show that the ensemble variance decreases away
296 from the barycenter of the plume and is close to zero outside of the plume. The ensemble variance of
297 concentration decreases continuously if more data are used for conditioning.

298 Table 5 shows the AAB and $AESP$ values for the concentrations at three times. Conditioning to
299 hydraulic conductivity and porosity data (S2) results in an average AAB reduction of 15% (compared with
300 S1) and $AESP$ reduces around 5%. Additional conditioning to piezometric head data (S3), results on
301 average in an additional 5% reduction of AAB and an $AESP$ reduction of around 7%. Further conditioning

302 to concentration data (S4, S5 and S6), yields prominent reduction of *AAB* and *AESP* (on average 19% and
303 20%, respectively).

304 We can see from the results: (1) concentration data is the type of data to most reduce the absolute bias
305 and uncertainty of predicted concentration; (2) the direct measured data and indirect head data also have an
306 important impact on the predicted concentrations; (3) when all the data are considered, the concentration
307 fields are best characterized.

308 *3.4. Reactive Transport Prediction Analysis*

309 In this subsection, a reactive transport prediction experiment is conducted with modified flow bound-
310 ary conditions using the conductivity and porosity obtained in the data assimilation exercise to further
311 demonstrate the robustness of EnKF.

312 The flow and transport configurations are the same as before but the flow is at steady-state. The eastern
313 constant flow rate boundary condition is replaced with the constant head boundary condition ($h = -15$ m)
314 and the solute mass is subject to sorption. Besides advection and dispersion also sorption according to a
315 reversible linear equilibrium isotherm is considered with $\rho_b = 1.81$ g/cm³ and $K_d = 0.52$ cm³/g (similar to
316 the values reported in the Borden aquifer (Mackay et al., 1986; Burr et al., 1994)). The reactive tracer is
317 also released near the western boundary (see Figure 2) with the same total initial concentration. The plume
318 snapshot at time $t = 500$ days (see Figure 13) is used to evaluate the worth of the different data.

319 MODFLOW and MT3DMS are employed to solve the flow equation (1) and reactive transport equation
320 (2), respectively.

321 Figure 14 shows the ensemble mean and variance of predicted concentration fields at $t = 500$ days for
322 the fields estimated from the scenarios S2, S3 and S6. It clearly shows that the predicted plume is close to
323 the reference when multiple types of information are used for conditioning. Besides, the ensemble variance
324 is the smallest for S6.

325 **4. Conclusion**

326 We have presented and demonstrated the Ensemble Kalman Filter, a data assimilation algorithm, to
327 jointly estimate hydraulic conductivity and porosity by assimilating dynamic piezometric head and multiple
328 concentration data in a hydrogeological stochastic model. Some of the attractive features of EnKF are the
329 capability of assimilating data in real-time, CPU efficiency, ease of implementation without need of an adjoint
330 model and the flexibility with regard to accounting for multiple sources of uncertainty jointly.

331 We have used a synthetic example (1) to demonstrate the potential EnKF has to condition in a CPU
332 efficient way to concentration data and (2) to analyze the worth of data for the characterization of aquifer
333 parameters and states (with a special focus on solute concentrations). We have found that the head data
334 have a distinctive impact to reduce the uncertainty of predicted piezometric head, but only a limited influ-
335 ence for improving the characterization of concentration distributions. Additional conditioning to multiple
336 concentration data was shown to improve strongly the predicted solute plume and also the characterization
337 of hydraulic conductivity and porosity.

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342 **References**

- 343 Alcolea, A., Carrera, J., Medina, A., 2006. Pilot points method incorporating prior information for solving
344 the groundwater flow inverse problem. *Advances in Water Resources* 29 (11), 1678–1689.
- 345 Anderson, J., 2001. An ensemble adjustment Kalman filter for data assimilation. *Monthly weather review*
346 129, 2884–2903.
- 347 Barnhart, K., Urteaga, I., Han, Q., Jayasumana, A., Illangasekare, T., 2010. On Integrating Groundwater
348 Transport Models with Wireless Sensor Networks. *Ground Water* 48 (5), 771–780.
- 349 Bear, J., 1972. *Dynamics of fluids in porous media*. American Elsevier Pub. Co., New York.
- 350 Burgers, G., van Leeuwen, P., Evensen, G., 1998. Analysis scheme in the ensemble Kalman filter. *Monthly*
351 *Weather Review* 126, 1719–1724.
- 352 Burr, D., Sudicky, E., Naff, R., 1994. Nonreactive and reactive solute transport in three-dimensional hetero-
353 geneous porous media: Mean displacement, plume spreading, and uncertainty. *Water Resources Research*
354 30 (3), 791–815.
- 355 Capilla, J., Llopis-Albert, C., 2009. Gradual conditioning of non-Gaussian transmissivity fields to flow and
356 mass transport data: 1. Theory. *Journal of Hydrology* 371 (1-4), 66–74.

- 357 Capilla, J. E., Rodrigo, J., Gómez-Hernández, J. J., 1999. Simulation of non-gaussian transmissivity fields
358 honoring piezometric data and integrating soft and secondary information. *Math. Geology* 31 (7), 907–927.
- 359 Carrera, J., Alcolea, A., Medina, A., Hidalgo, J., Slooten, L., 2005. Inverse problem in hydrogeology. *Hydro-*
360 *geology Journal* 13 (1), 206–222.
- 361 Carrera, J., Neuman, S., 1986. Estimation of aquifer parameters under transient and steady state conditions:
362 1. Maximum likelihood method incorporating prior information. *Water Resources Research* 22 (2), 199–
363 210.
- 364 Chen, Y., Zhang, D., 2006. Data assimilation for transient flow in geologic formations via ensemble Kalman
365 filter. *Advances in Water Resources* 29 (8), 1107–1122.
- 366 Doherty, J., 2004. PEST model-independent parameter estimation, user manual. Watermark Numerical
367 Computing, Brisbane, Australia, 3349.
- 368 Evensen, G., 2003. The ensemble Kalman filter: Theoretical formulation and practical implementation.
369 *Ocean dynamics* 53 (4), 343–367.
- 370 Freeze, R. A., Cherry, J. A., 1979. *Groundwater*. Prentice-Hall.
- 371 Fu, J., Gómez-Hernández, J., 2009. Uncertainty assessment and data worth in groundwater flow and mass
372 transport modeling using a blocking Markov chain Monte Carlo method. *Journal of Hydrology* 364 (3-4),
373 328–341.
- 374 Gómez-Hernández, J. J., Journel, A. G., 1993. Joint sequential simulation of multi-Gaussian fields. *Geo-*
375 *statistics Troia* 92 (1), 85–94.
- 376 Gómez-Hernández, J. J., Sahuquillo, A., Capilla, J. E., 1997. Stochastic simulation of transmissivity fields
377 conditional to both transmissivity and piezometric data, 1, Theory. *Journal of Hydrology* 203 (1–4), 162–
378 174.
- 379 Gómez-Hernández, J. J., Srivastava, R. M., 1990. ISIM3D: an ANSI-C three dimensional multiple indicator
380 conditional simulation program. *Computers & Geosciences* 16 (4), 395–440.
- 381 Gu, Y., Oliver, D., 2006. The ensemble Kalman filter for continuous updating of reservoir simulation models.
382 *Journal of Energy Resources Technology* 128, 79.

- 383 Harbaugh, A. W., Banta, E. R., Hill, M. C., McDonald, M. G., 2000. MODFLOW-2000, the U.S. Geological
384 Survey modular ground-water model. U.S. Geological Survey, Branch of Information Services, Reston, VA,
385 Denver, CO.
- 386 Hassan, A., 2001. Water flow and solute mass flux in heterogeneous porous formations with spatially random
387 porosity. *Journal of Hydrology* 242 (1-2), 1–25.
- 388 Hendricks Franssen, H., Alcolea, A., Riva, M., Bakr, M., van der Wiel, N., Stauffer, F., Guadagnini, A.,
389 2009. A comparison of seven methods for the inverse modelling of groundwater flow. application to the
390 characterisation of well catchments. *Advances in Water Resources* 32 (6), 851–872.
- 391 Hendricks Franssen, H., Gómez-Hernández, J., Capilla, J., Sahuquillo, A., 1999. Joint simulation of trans-
392 missivity and storativity fields conditional to steady-state and transient hydraulic head data. *Advances in*
393 *Water Resources* 23 (1), 1–13.
- 394 Hendricks Franssen, H., Gómez-Hernández, J., Sahuquillo, A., 2003. Coupled inverse modelling of groundwa-
395 ter flow and mass transport and the worth of concentration data. *Journal of Hydrology* 281 (4), 281–295.
- 396 Hendricks Franssen, H., H. P. Kaiser, U. Kuhlmann, G. B. F. S. R. M., Kinzelbach, W., 2011. Operational
397 real-time modeling with EnKF of variably saturated subsurface flow including stream-aquifer interaction
398 and parameter updating. *Water Resources Research*, in press,doi:10.1029/2010WR009480.
- 399 Hendricks Franssen, H., Kinzelbach, W., 2008. Real-time groundwater flow modeling with the Ensemble
400 Kalman Filter: Joint estimation of states and parameters and the filter inbreeding problem. *Water Re-*
401 *sources Research* 44 (9), W09408.
- 402 Hendricks Franssen, H., Kinzelbach, W., 2009. Ensemble Kalman filtering versus sequential self-calibration
403 for inverse modelling of dynamic groundwater flow systems. *Journal of Hydrology* 365 (3-4), 261–274.
- 404 Hu, B., Meerschaert, M., Barrash, W., Hyndman, D., He, C., Li, X., Guo, L., 2009. Examining the influence
405 of heterogeneous porosity fields on conservative solute transport. *Journal of contaminant hydrology* 108 (3-
406 4), 77–88.
- 407 Hu, L. Y., 2000. Gradual deformation and iterative calibration of gaussian-related stochastic models. *Math.*
408 *Geology* 32 (1), 87–108.

- 409 Huang, C., Hu, B. X., Li, X., Ye, M., 2008. Using data assimilation method to calibrate a heterogeneous con-
410 ductivity field and improve solute transport prediction with an unknown contamination source. *Stochastic*
411 *Environmental Research and Risk Assessment* 23 (8), 1155–1167.
- 412 Huang, H., Hu, B., Wen, X., Shirley, C., 2004. Stochastic inverse mapping of hydraulic conductivity and
413 sorption partitioning coefficient fields conditioning on nonreactive and reactive tracer test data. *Water*
414 *Resources Research* 40 (1), W01506.
- 415 Jiang, X., Wan, L., Cardenas, M., Ge, S., Wang, X., 2010. Simultaneous rejuvenation and aging of ground-
416 water in basins due to depth-decaying hydraulic conductivity and porosity. *Geophysical Research Letters*
417 37 (5), L05403.
- 418 Journel, A., 1974. Geostatistics for conditional simulation of ore bodies. *Economic Geology* 69 (5), 673.
- 419 Kitanidis, P., 1995. Quasi-linear geostatistical theory for inversing. *Water Resources Research* 31 (10), 2411–
420 2419.
- 421 LaVenue, A. M., Ramarao, B. S., de Marsily, G., Marietta, M. G., 1995. Pilot point methodology for
422 automated calibration of an ensemble of conditionally simulated transmissivity fields, 2, Application.
423 *Water Resour. Res.* 31 (3), 495–516.
- 424 Li, L., Zhou, H., Gómez-Hernández, J. J., 2011a. A comparative study of three-dimensional hydraulic conduc-
425 tivity upscaling at the macrodispersion experiment (MADE) site, on columbus air force base in mississippi
426 (USA). *Journal of Hydrology*, doi:10.1016/j.jhydrol.2011.05.001.
- 427 Li, L., Zhou, H., Gómez-Hernández, J. J., 2011b. Transport upscaling using multi-rate mass transfer in
428 three-dimensional highly heterogeneous porous media. *Advances in Water Resources* 34 (4), 478–489.
- 429 Li, L., Zhou, H., Hendricks Franssen, H., Gómez-Hernández, J. J., 2011c. Modeling transient flow by coupling
430 ensemble kalman filtering and upscaling. *Water Resources Research*, submitted.
- 431 Liu, G., Chen, Y., Zhang, D., 2008. Investigation of flow and transport processes at the MADE site using
432 ensemble kalman filter. *Advances in Water Resources* 31 (7), 975–986.
- 433 Llopis-Albert, C., Capilla, J., 2009. Gradual conditioning of non-Gaussian transmissivity fields to flow and
434 mass transport data: 2. Demonstration on a synthetic aquifer. *Journal of Hydrology* 371 (1-4), 53–65.

- 435 Mackay, D., Freyberg, D., Roberts, P., Cherry, J., 1986. A natural gradient experiment on solute transport
436 in a sand aquifer: 1. Approach and overview of plume movement. *Water Resources Research* 22 (13),
437 2017–2029.
- 438 McLaughlin, D., Townley, L., 1996. A reassessment of the groundwater inverse problem. *Water Resources*
439 *Research* 32 (5), 1131–1161.
- 440 Medina, A., Carrera, J., 1996. Coupled estimation of flow and solute transport parameters. *Water Resources*
441 *Research* 32 (10), 3063–3076.
- 442 Naevdal, G., Johnsen, L., Aanonsen, S., Vefring, E., 2005. Reservoir monitoring and continuous model
443 updating using ensemble kalman filter. *SPE Journal* 10 (1).
- 444 Naff, R., Haley, D., Sudicky, E., 1998. High-resolution Monte Carlo simulation of flow and conservative
445 transport in heterogeneous porous media 1. Methodology and flow results. *Water Resources Research*
446 34 (4), 663–677.
- 447 Nowak, W., 2009. Best unbiased ensemble linearization and the quasi-linear kalman ensemble generator.
448 *Water Resources Research* 45 (4), W04431.
- 449 Oliver, D., Cunha, L., Reynolds, A., 1997. Markov chain Monte Carlo methods for conditioning a perme-
450 ability field to pressure data. *Mathematical Geology* 29 (1), 61–91.
- 451 Ramarao, B. S., LaVenue, A. M., de Marsily, G., Marietta, M. G., 1995. Pilot point methodology for
452 automated calibration of an ensemble of conditionally simulated transmissivity fields, 1, Theory and
453 computational experiments. *Water Resour. Res.* 31 (3), 475–493.
- 454 Reichle, R., Walker, J., Koster, R., Houser, P., 2002. Extended versus ensemble Kalman filtering for land
455 data assimilation. *Journal of hydrometeorology* 3, 728–740.
- 456 Riva, M., Guadagnini, A., Fernandez-Garcia, D., Sanchez-Vila, X., Ptak, T., 2008. Relative importance of
457 geostatistical and transport models in describing heavily tailed breakthrough curves at the lauswiesen site.
458 *Journal of Contaminant Hydrology* 101 (1-4), 1–13.
- 459 Sahuquillo, A., Capilla, J. E., Gómez-Hernández, J. J., Andreu, J., 1992. Conditional simulation of transmis-
460 sivity fields honouring piezometric head data. In: Blair, W. R., Cabrera, E. (Eds.), *Hydraulic Engineering*
461 *Software IV, Fluid Flow Modeling. Vol. II. Elsevier Applied Science, London, UK*, pp. 201–214.

- 462 Schöniger, A., Nowak, W., Hendricks Franssen, H. J., 2011. Parameter estimation by ensemble Kalman
463 filters with transformed data: approach and application to hydraulic tomography, in preparation for
464 Water Resources Research.
- 465 Schwede, R., Cirpka, O., 2009. Use of steady-state concentration measurements in geostatistical inversion.
466 Advances in Water Resources 32 (4), 607–619.
- 467 Strebelle, S., 2002. Conditional simulation of complex geological structures using multiple-point statistics.
468 Mathematical Geology 34 (1), 1–21.
- 469 Sun, A. Y., Morris, A. P., Mohanty, S., Jul. 2009. Sequential updating of multimodal hydrogeologic parameter
470 fields using localization and clustering techniques. Water Resources Research 45, 15 PP.
- 471 Wen, X., Deutsch, C., Cullick, A., 2002. Construction of geostatistical aquifer models integrating dynamic
472 flow and tracer data using inverse technique. Journal of Hydrology 255 (1-4), 151–168.
- 473 Wen, X. H., Chen, W., 2005. Real-time reservoir model updating using ensemble Kalman filter. In: SPE
474 reservoir simulation symposium.
- 475 Yeh, W., 1986. Review of parameter identification procedures in groundwater hydrology: The inverse prob-
476 lem. Water Resources Research 22 (2), 95–108.
- 477 Zheng, C., Wang, P., TUSCALOOSA., A. U., 1999. MT3DMS: A modular three-dimensional multispecies
478 transport model for simulation of advection, dispersion, and chemical reactions of contaminants in ground-
479 water systems; Documentation and user's guide.
- 480 Zhou, H., Gómez-Hernández, J. J., Hendricks Franssen, H., Li, L., 2011a. Handling non-
481 gaussian distributions with Ensemble Kalman Filter. Advances in Water Resources, in press,
482 doi:10.1016/j.advwatres.2011.04.014.
- 483 Zhou, H., Li, L., Gómez-Hernández, J. J., 2010. Three-dimensional hydraulic conductivity upscaling in
484 groundwater modelling. Computers & Geosciences 36 (10), 1224–1235.
- 485 Zhou, H., Li, L., Hendricks Franssen, H., Gómez-Hernández, J. J., 2011b. Pattern recongition in a bimodal
486 aquifer with normal-score Ensemble Kalman Filter. Mathematical Geosciences, under review.
- 487 Zimmerman, D., De Marsily, G., Gotway, C., Marietta, M., Axness, C., Beauheim, R., Bras, R., Carrera,
488 J., Dagan, G., Davies, P., et al., 1998. A comparison of seven geostatistically based inverse approaches to

489 estimate transmissivities for modeling advective transport by groundwater flow. *Water Resources Research*
490 34 (6), 1373–1413.

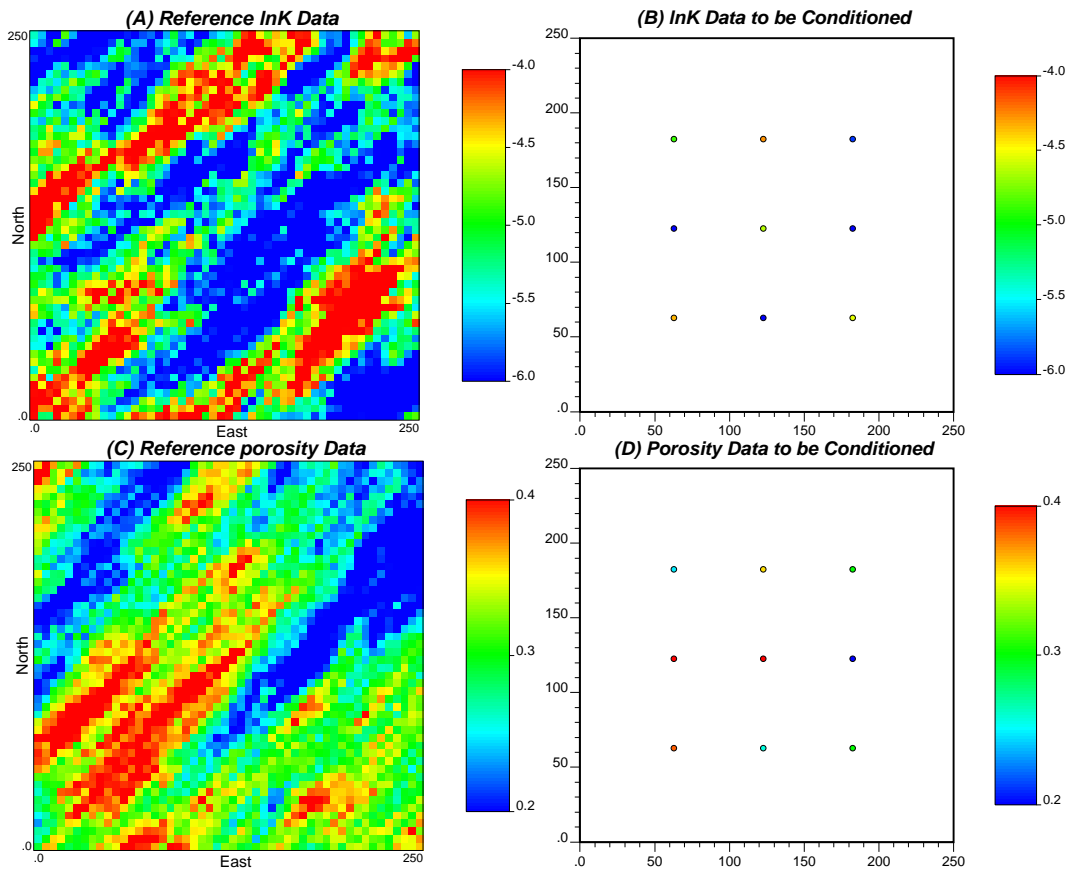


Figure 1: (A) Reference $\ln K$ field, (B) Conditioning $\ln K$ data, (C) Reference porosity(ϕ) field, (D) Conditioning porosity(ϕ) data.

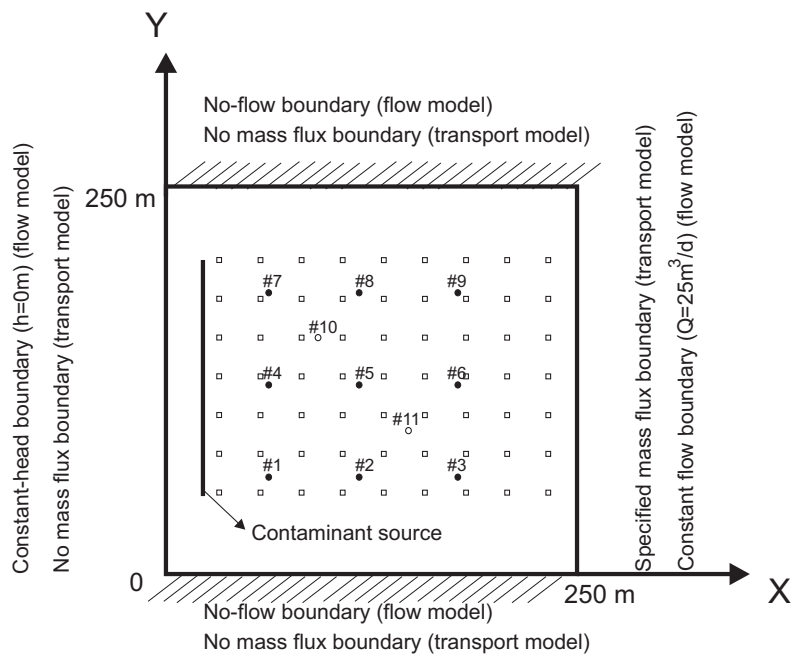


Figure 2: Sketch of the flow and transport simulation with boundary conditions and observation and prediction wells. Filled circles correspond to the pressure head observation wells (#1-#9); Open circles denote the control wells (#10-#11). Empty squares indicate the wells where concentration is sampled.

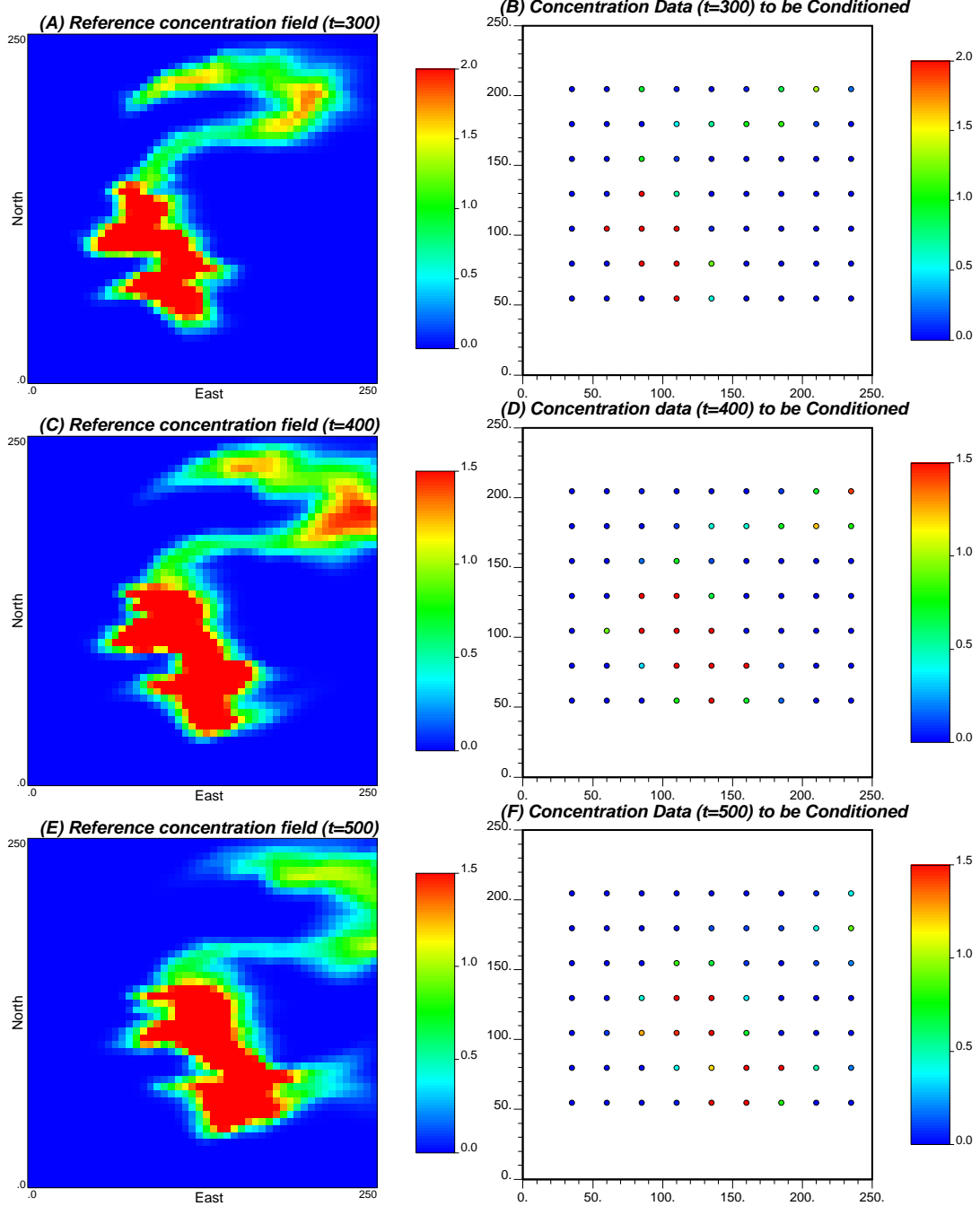


Figure 3: Reference concentration fields at time 300(A), 400(C) and 500(E) days. Conditioning concentration data at time 300 (B), 400 (D), and 500 (F) days.

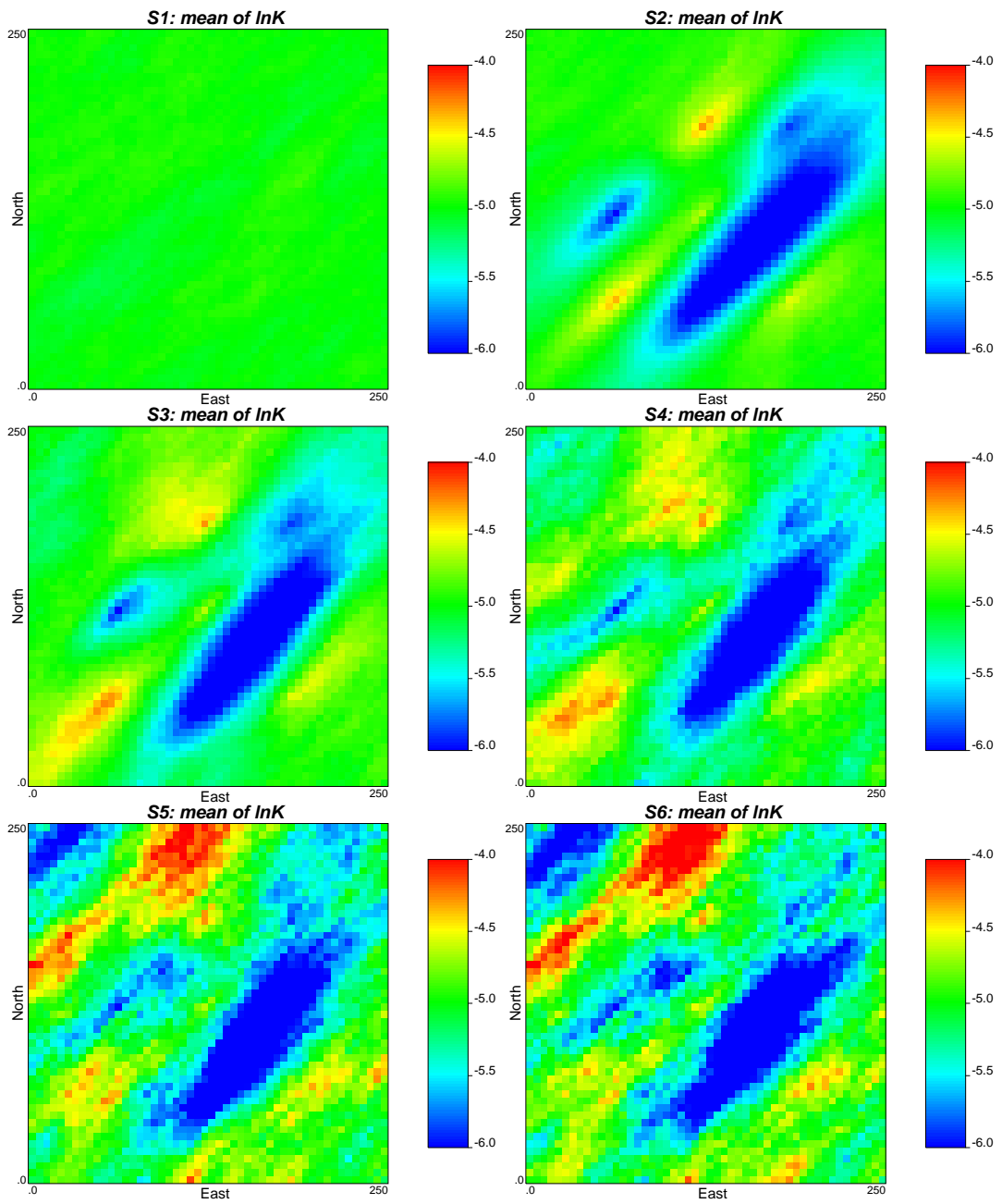


Figure 4: Ensemble average logconductivity fields for the different scenarios.

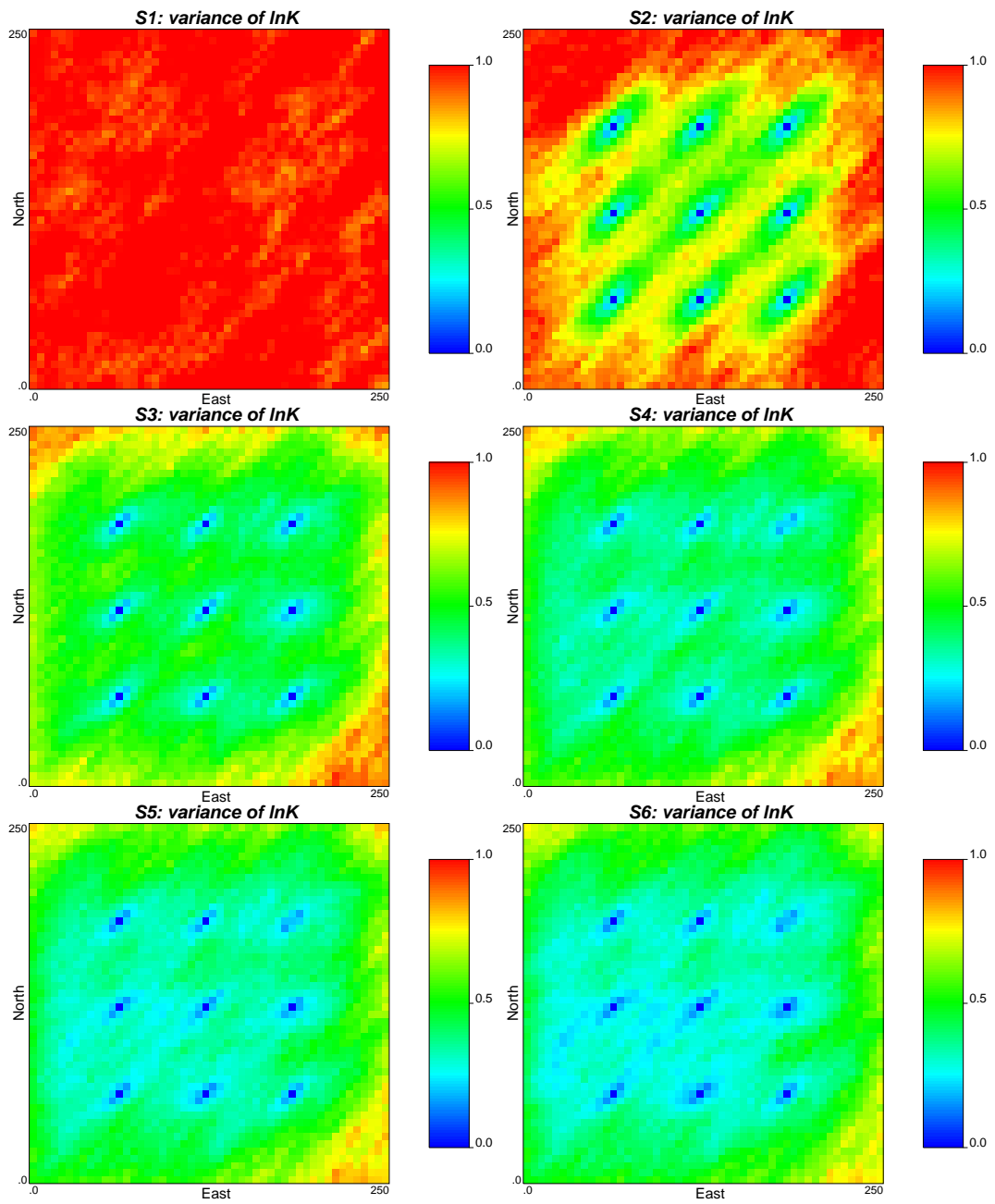


Figure 5: Ensemble logconductivity variance fields for the different scenarios

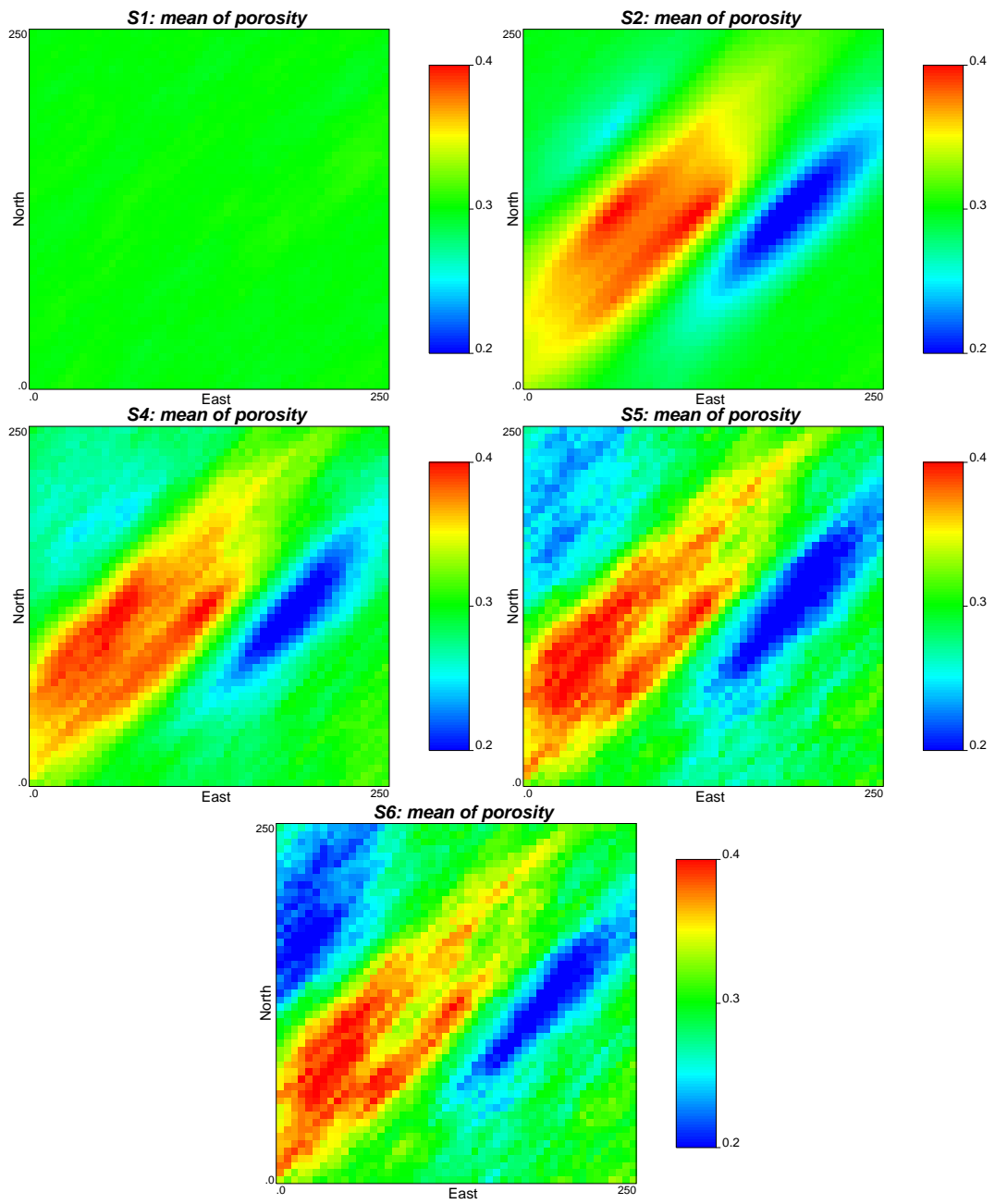


Figure 6: Ensemble average porosity fields for the different scenarios.

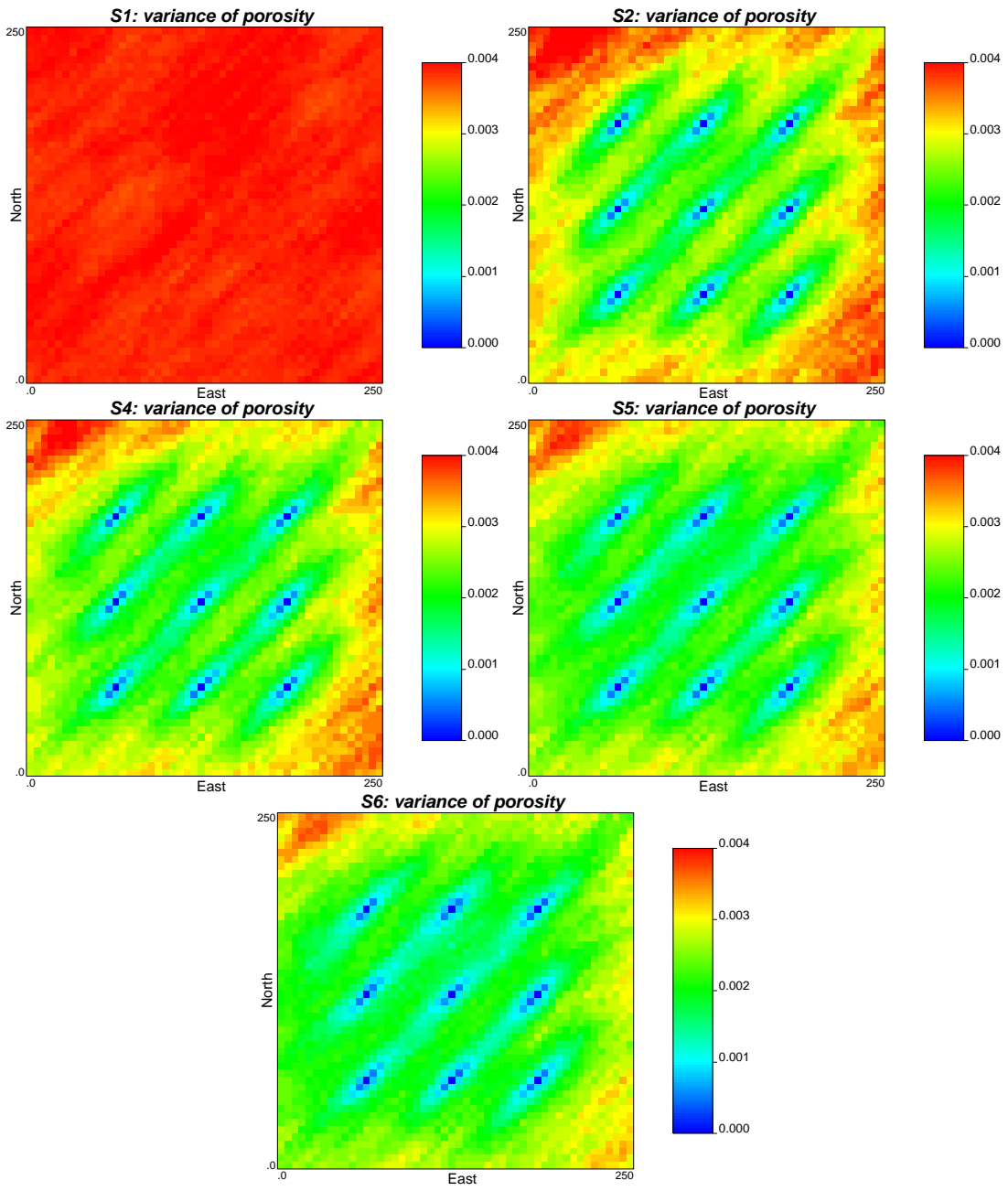


Figure 7: Ensemble porosity variance fields for the different scenarios.

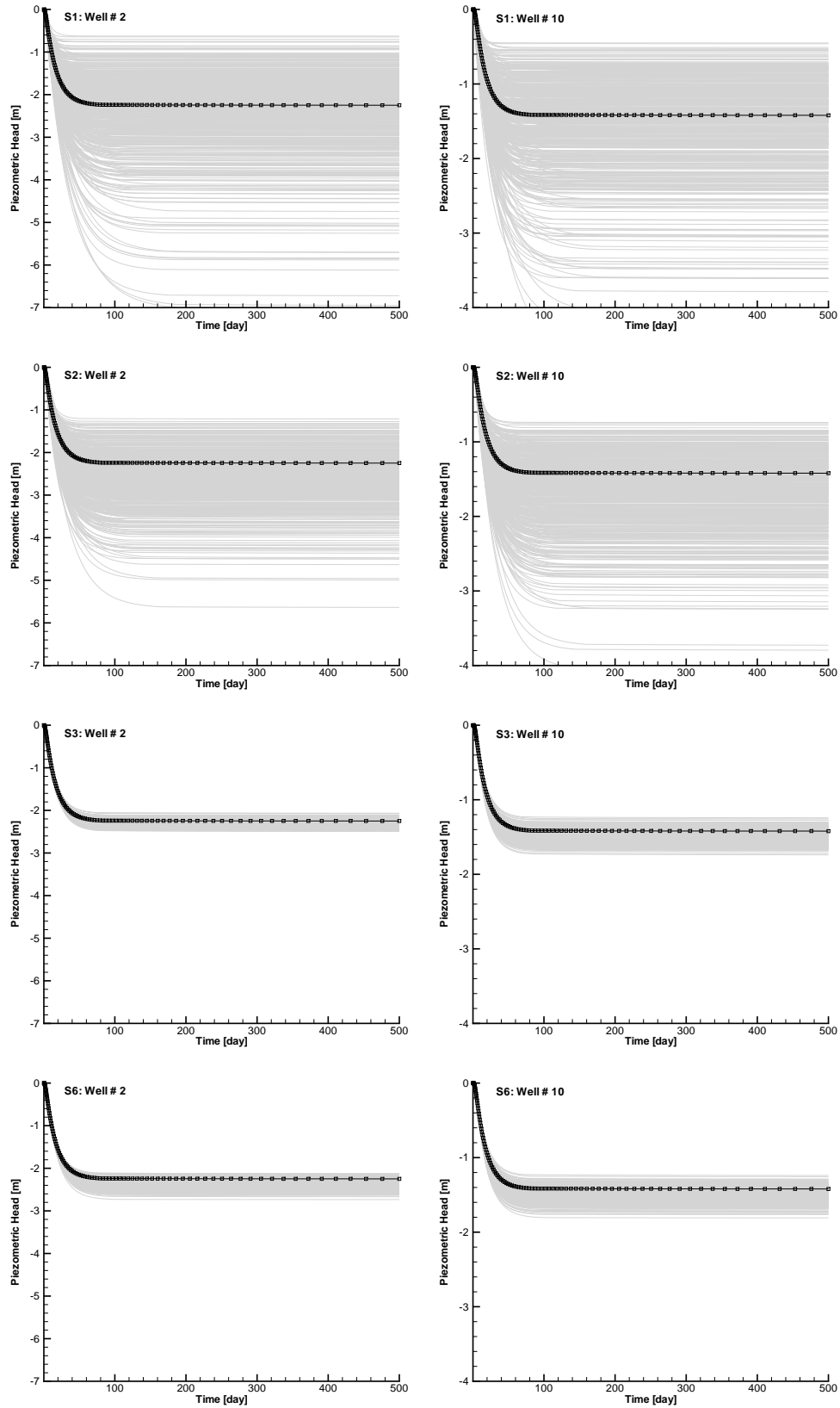


Figure 8: Piezometric head time series for the reference field (black) and simulated ones (gray lines) for the S1, S2, S3 and S6 scenarios at the conditioning well W2 (left column) and verification well W10 (right column).

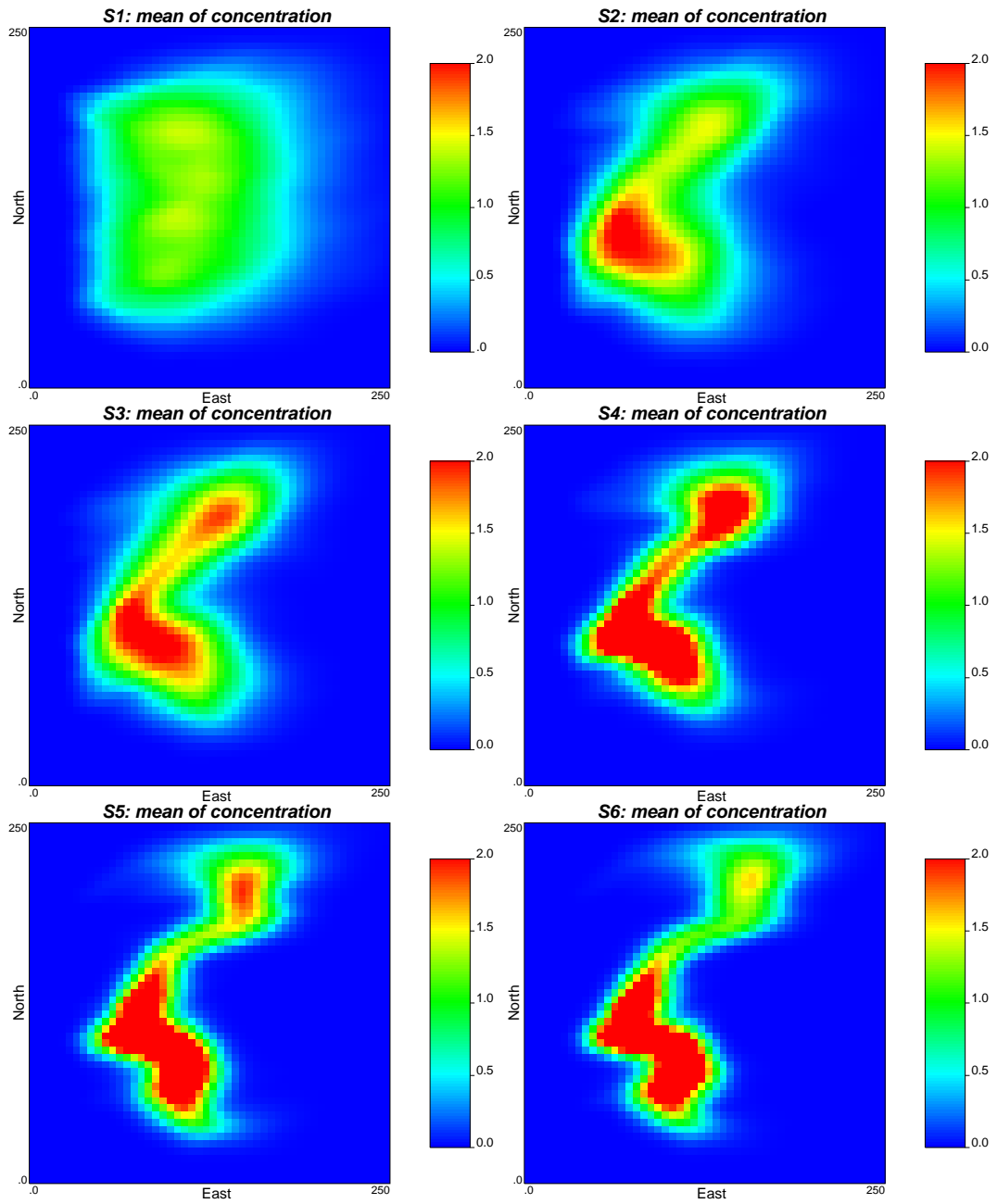


Figure 9: Ensemble average concentration fields at $t = 300$ day for the different scenarios.

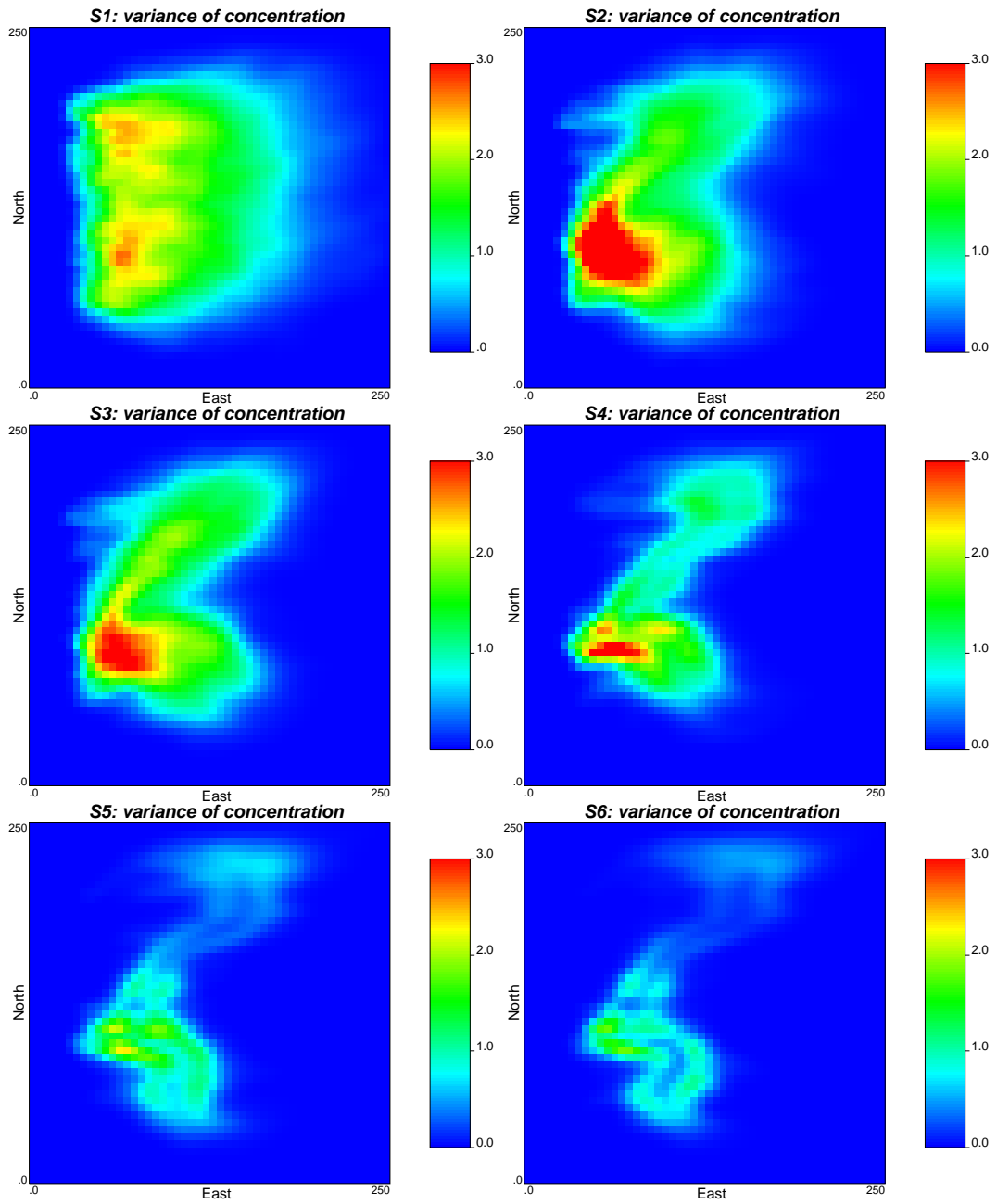


Figure 10: Ensemble variance of concentration fields at $t = 300$ day for the different scenarios.

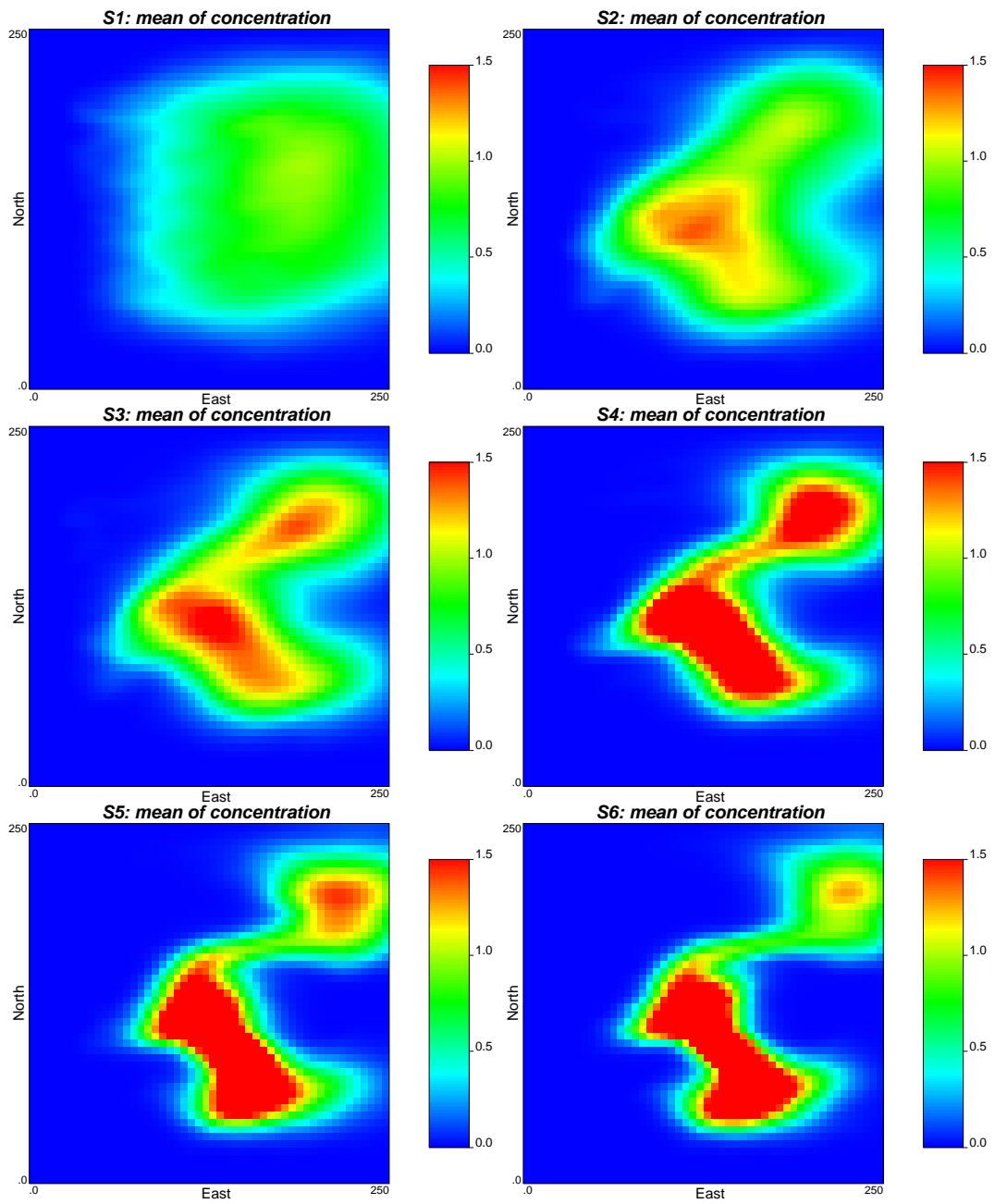


Figure 11: Ensemble average concentration fields at $t = 500$ day for the different scenarios.

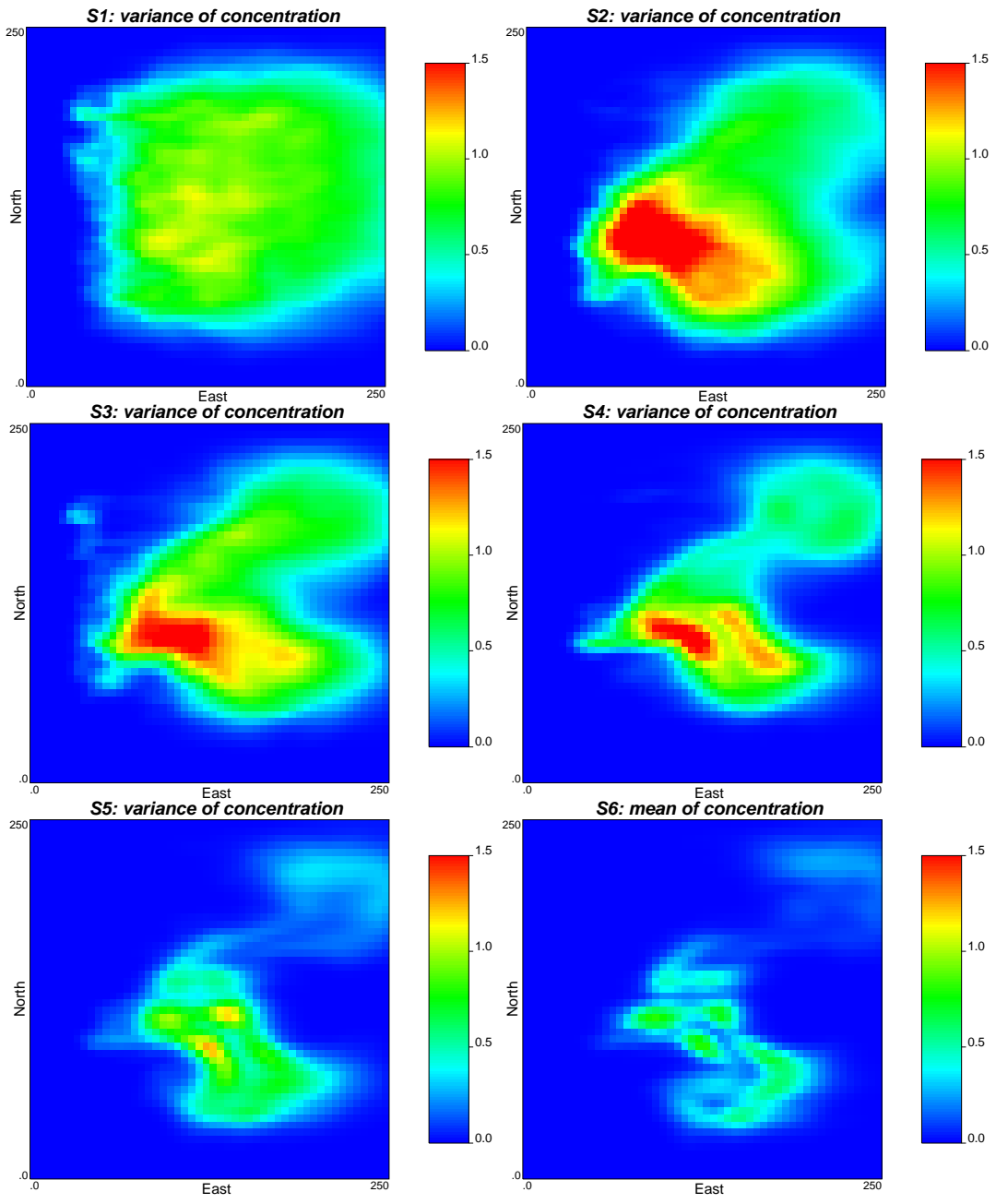


Figure 12: Ensemble variance of concentration fields at $t = 500$ day for the different scenarios.

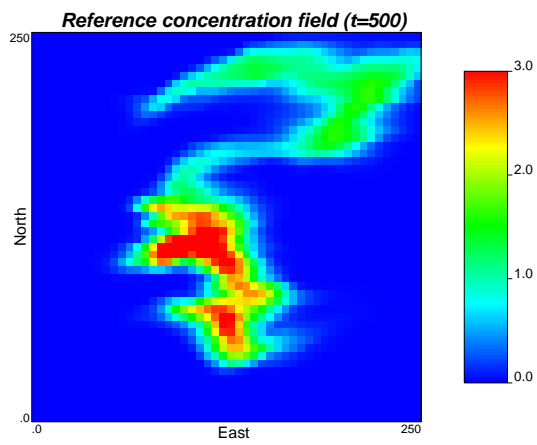


Figure 13: The reference concentration field at $t = 500$ days for the reactive transport prediction experiment.

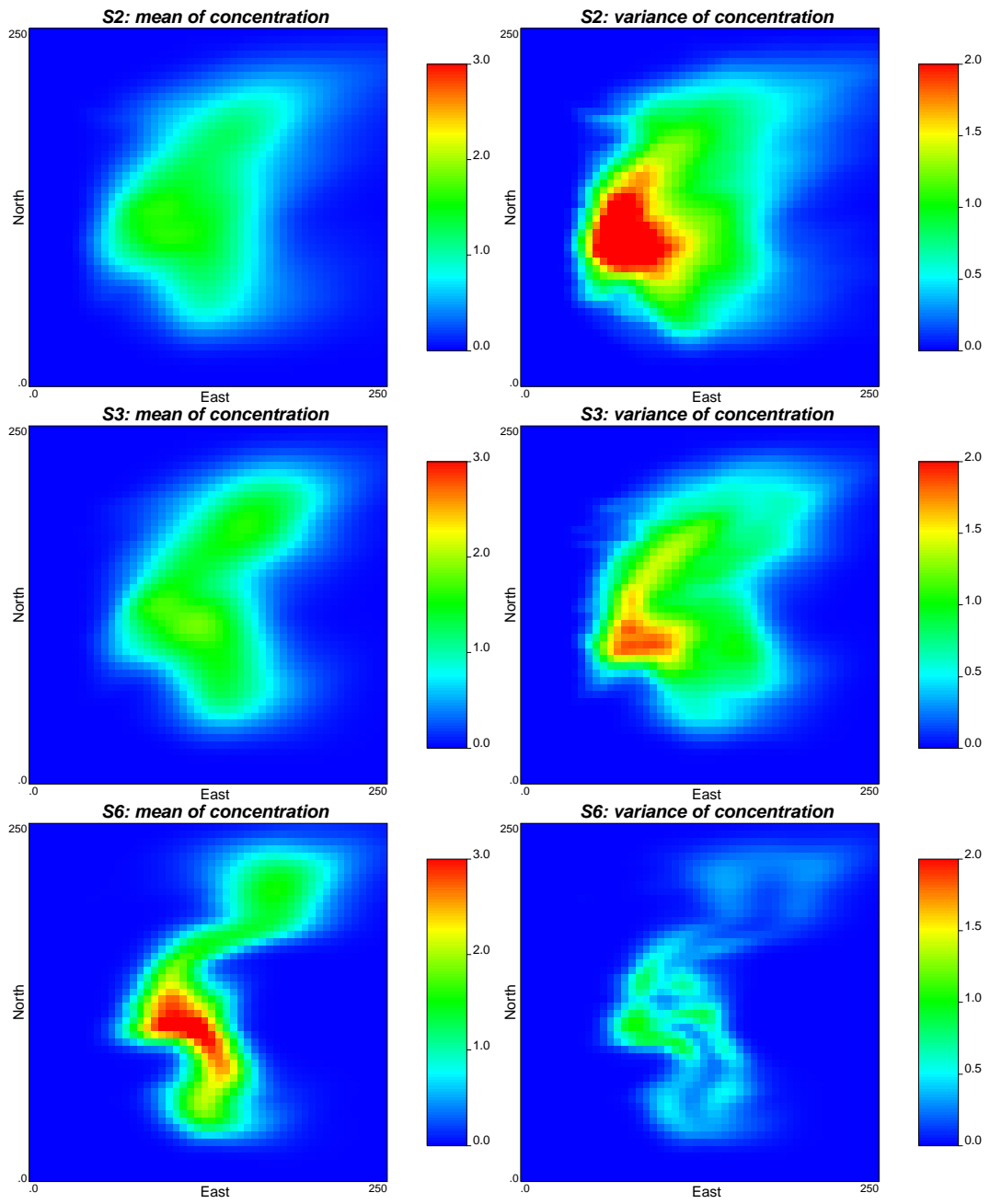


Figure 14: Ensemble mean and variance of concentration fields at $t = 500$ day for the S2,S3 and S6.

Table 1: Parameters of the random functions for modeling the spatial distributions of $\ln K$ and porosity

	Mean	Variance	Variogram type	λ_x [m]	λ_y [m]	rotation angle β
$\ln K$	-5	1	exponential	180	60	45°
ϕ	0.3	0.0036	exponential	240	60	45°

β denotes the rotation angle of one clockwise rotation of positive y axis.

Table 2: Definition of scenarios based on the different sets of conditioning data.

Scenario	S1	S2	S3	S4	S5	S6
Hydraulic conductivities (K)	No	Yes	Yes	Yes	Yes	Yes
Porosity (ϕ)	No	Yes	Yes	Yes	Yes	Yes
Dynamic piezometric heads (h)	No	No	Yes	Yes	Yes	Yes
Concentrations ($t = 300$ day)	No	No	No	Yes	Yes	Yes
Concentrations ($t = 400$ day)	No	No	No	No	Yes	Yes
Concentrations ($t = 500$ day)	No	No	No	No	No	Yes

Table 3: Bias and spread of $\ln K$ and porosity for the different scenarios.

Scenario	S1	S2	S3	S4	S5	S6
$AAB(\ln K)$	1.112	0.949	0.852	0.816	0.796	0.790
$AESP(\ln K)$	1.001	0.874	0.728	0.680	0.650	0.624
$AAB(\phi)$	0.072	0.059	-	0.057	0.056	0.055
$AESP(\phi)$	0.060	0.051	-	0.049	0.047	0.046

Table 4: Bias and spread of predicted piezometric heads at time $t = 67.7$ days for the different scenarios.

Scenario	S1	S2	S3	S4	S5	S6
$AAB(h_{t=67.7})$	0.690	0.503	0.169	0.170	0.179	0.179
$AESP(h_{t=67.7})$	0.901	0.649	0.175	0.172	0.169	0.162

Table 5: Bias and spread of predicted concentrations at time $t = 300$, $t = 400$, $t = 500$ days for the different scenarios.

Scenario	S1	S2	S3	S4	S5	S6
$AAB(c_{t=300})$	0.493	0.402	0.384	0.318	0.252	0.225
$AESP(c_{t=300})$	0.781	0.703	0.652	0.496	0.400	0.337
$AAB(c_{t=400})$	0.506	0.422	0.403	0.331	0.249	0.209
$AESP(c_{t=400})$	0.710	0.662	0.613	0.470	0.371	0.300
$AAB(c_{t=500})$	0.452	0.393	0.374	0.303	0.226	0.176
$AESP(c_{t=500})$	0.634	0.624	0.577	0.457	0.358	0.274