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Additional Information

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3 1 **Numerical Simulation of vertical flow wetlands with special emphasis on**  
4 2 **treatment performance during winter**  
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23 13 **Abstract**

24 14 In Austria, single stage vertical flow (VF) wetlands with intermittent loading are a state-of-the-art  
25 15 technology for treating domestic wastewater. They are designed according to the Austrian design  
26 16 standard with a specific surface area of 4 m<sup>2</sup> per person (i.e. 20 g COD/m<sup>2</sup>/d) and thus demand a  
27 17 bigger amount of land to treat the same amount of wastewater compared to intensified technical  
28 18 treatment systems. In order to reduce the amount of land needed, a modified design for VF  
29 19 wetlands have been proposed. The modified design has a specific surface area of 2.5 m<sup>2</sup> per person  
30 20 (i.e. 32 g COD/m<sup>2</sup>/d) and it has been shown to be able to meet the Austrian effluent requirements.  
31 21 To allow higher organic loading, more loadings per day but lower volume of a single loading, a  
32 22 constant loading interval, and increased number of openings per m<sup>2</sup> are applied. A simulation study  
33 23 using the HYDRUS Wetland Module was carried out to compare the treatment efficiencies of  
34 24 single-stage VF wetlands with classical and modified design. Data from a classical Austrian single-  
35 25 stage VF wetland was used for calibration of the model using the standard parameter set for the  
36 26 CW2D biokinetic model. The influent COD fractionation was calibrated to adapt to the  
37 27 wastewater. The simulations showed a good performance of the modified design compared to a  
38 28 classical VF wetland for COD removal with COD effluent concentrations in winter (effluent water  
39 29 temperature of 4.5 °C) of 35 and 29 mg/L, respectively. The simulation study showed that during  
40 30 high-loading events the VF wetland with modified design has lower maximum NH<sub>4</sub>-N effluent  
41 31 concentrations. Single-stage VF wetlands with modified design seem to be very effective and  
42 32 allow applying higher organic loads compared to single-stage VF wetlands with classical design.  
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51 34 **Keywords**

53 35 HYDRUS Wetland Module, numerical simulation, specific surface area, Treatment wetlands,  
54 36 vertical flow.  
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## INTRODUCTION

Treatment wetlands (TWs) are engineered systems designed to optimize the chemical, physical and biogeochemical processes that occur in wetland ecosystems to treat wastewater. TWs have a broad range of applications, ranging from secondary and/or tertiary treatment of urban wastewater (Dotro et al. 2017), treatment of raw wastewater in French VF wetlands (Molle et al., 2005), to specific treatments, such as industrial effluents treatment, water recycling or sludge dewatering (Wu et al., 2015; Martín et al., 2013; Brix, 2017).

In Austria ammonia nitrogen has to be removed for all wastewater treatment plants (WWTPs), the maximum allowed effluent concentration for small WWTPs ( $\leq 500$  persons connected) is 10 mg NH<sub>4</sub>-N/L at effluent water temperatures higher than 12 °C (1.AEVkA, 1996). Thus, only vertical flow (VF) wetlands can be implemented. Therefore, single stage VF wetlands with intermittent loading are a state-of-art technology used for secondary treatment of domestic wastewater. In fact, 20% of all small WWTPs in Austria are VF wetlands (Langergraber et al., 2018) and most of them are designed according to the Austrian design standard (ÖNORM B 2505, 2009) with a specific surface area of 4 m<sup>2</sup> per person. Those VF wetlands are situated mostly in rural areas, where VF wetlands are a very suitable technology for wastewater treatment due to its robustness, the low operation, and maintenance requirements and the high availability of land required. Decreasing the specific area required would allow TWs to be used in other areas with less availability of land.

A modified design approach has been proposed and implemented in Austria by a company. This modified design aims to reduce the specific surface area requirement to 2.5 m<sup>2</sup> per person by still meeting the Austrian effluent standards. Additional differences of the modified design are the higher number of loadings, a controlled loading interval and the density of openings in the distribution pipes (Table 1).

Table 1: Differences of VF wetland design parameters

<i>Design</i>	<i>Classical Austrian single-stage VF wetland (ÖNORM B 2505, 2009)</i>	<i>Modified design</i>
Specific surface area	4 m <sup>2</sup> per person	2.5 m <sup>2</sup> per person
Organic load	20 g COD/m <sup>2</sup> /d	32 g COD/m <sup>2</sup> /d
Loadings	3-4 times per day	8 times per day
Loading interval	Not controlled, depending on wastewater generation	Controlled, every 3 hours (storage tank required)
Openings distribution pipes	1 per 2 m <sup>2</sup>	1 per m <sup>2</sup>

The aim of this work is to compare the effluent quality of classical Austrian single-stage VF wetlands during winter conditions with different design parameters. For comparing the different VF wetland designs, a numerical simulation study using the HYDRUS Wetland Module (Langergraber and Šimůnek, 2012) was carried out. The simulations have been run for different temperatures and focus on the operation in winter, as these correspond to the most extreme conditions to which the treatment system is subjected.

## MATERIALS AND METHODS

### Experimental Data

Table 2 shows measured effluent flow and quality data used for this work that were obtained in 2008 from the experimental site situated in the WWTP Ernsthofen (Lower Austria). The VF wetland was designed according to the Austrian design standards for secondary treatment with a specific surface area of 4 m<sup>2</sup> per person, i.e. it was loaded with an organic load of 20 g COD/m<sup>2</sup>/d. The total surface area of the VF bed was 20 m<sup>2</sup> (4 m width, 5 m length) and 4 pipes evenly distributed along the width of the bed were used to distribute the wastewater on the surface of the bed. The grain size of the sand used for the 50 cm main layer was 0.06-4 mm (d<sub>10</sub> = 0.2 mm; d<sub>60</sub> = 0.8 mm). Additionally, the filter consisted of a 10 cm intermediate layer with grain size 4-8 mm and a 15 cm drainage layer with grain size 16-32 mm. The VF wetland was planted with common reed (*Phragmites australis*). It was constructed in spring 2003 and has been in operation for 5 years at the time the data were collected.

Table 2: Measured influent and effluent concentrations at different temperatures for the classical Austrian single-stage VF wetland (data derived from Langergraber et al., 2010).

Parameter	Temperature	COD	NH <sub>4</sub> -N	NO <sub>2</sub> -N	NO <sub>3</sub> -N	PO <sub>4</sub> -P
Unit	°C	mg/L	mg/L	mg/L	mg/L	mg/L
<b>Influent</b>	-	495	65	0.015	0.4	11.9
	4.5	27	6.47	0.692	49.2	8.3
	7.5	18	0.02	0.015	57.4	8.3
<b>Effluent</b>	10.5	19	0.02	0.015	72.1	10.6
	13.5	19	0.02	0.015	60.6	10.5
	16.5	21	0.03	0.022	70.0	11.1
	19.5	18	0.02	0.015	60.6	10.5

A constant volume for each loading was applied. The classical VF wetland was loaded intermittently 4 times per day. The times of loading were 07:40, 11:40, 19:00 and 21:40 hours. The loading times have been calculated from the diurnal variation for the influent flow given in the standard ON EN 12556-3 (2005) for testing small wastewater treatment plants. During the

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3 94 experiments, the different loading scenarios described for the 48 week testing procedure of the  
4 95 standard were applied to the VF wetland. For more details of the experimental site and the  
5 96 experiments carried out in this period the reader is referred to Langergraber et al. (2010). In this  
6 97 paper, this system is referred to as "classical VF wetland".

8 98 For wetlands with the modified design, only effluent concentrations collected during the self-  
9 99 monitoring of the plant were available. As no influent data for these wetlands were available, the  
10 100 data could not be used for the simulation study and are thus not presented here. To make a  
11 101 comparison of the two designs possible, influent data from the Ernsthofen site were used for both  
12 102 designs.

### 13 103 14 104 **Simulation Study**

15 105 The HYDRUS Wetland Module (Langergraber and Šimůnek, 2012) was used for this simulation  
16 106 study. The HYDRUS software can be used to simulate partially saturated porous media, such as  
17 107 used in TWs, since HYDRUS numerically solves the Richards Equation for water flow and uses  
18 108 the convection-dispersion equation to simulate heat and solutes transport. Accordingly, given that  
19 109 the TWs used for the study were VF wetlands, the CW2D biokinetic model (Langergraber and  
20 110 Šimůnek, 2005) was chosen. Its mathematical formulation is based on Activated Sludge Models  
21 111 (Henze et al., 2000) and it describes aerobic and anoxic processes to simulate the transformation  
22 112 and degradation processes of compounds (organic matter, nitrogen, and phosphorus).

23 113 A modified parameter set was introduced by Langergraber (2007) to simulate temperature  
24 114 dependencies in the degradation of COD and NH<sub>4</sub>-N. It was achieved by adjusting the half-  
25 115 saturation constants included in the Monod terms in the reaction rates equations for hydrolysis and  
26 116 nitrification processes. Thus, the standard biokinetic parameters, defined in Langergraber and  
27 117 Šimůnek (2012), have been used for all the temperatures and the different designs to compare its  
28 118 performance. As phosphorus has not been the focus of the study, no adsorption processes have  
29 119 been considered.

30 120 The water flow model was calibrated using measured data from the 2008 experiments (see before)  
31 121 using a constant volume of 175 L for each loading. The reactive transport was calibrated for the  
32 122 classical Austrian VF wetland design using the data shown in Table 2. Firstly, the effluent COD  
33 123 at the highest effluent water temperature of 19.5 °C was fitted by adjusting the inert fraction of the  
34 124 influent COD, secondly, the ratio between slowly and readily biodegradable COD was calibrated  
35 125 at the lowest effluent water temperature of 4.5 °C. For reactive transport simulations, the volume  
36 126 of each loading was 202 L per loading to guarantee the desired organic loading rate of  
37 127 20 g COD/m<sup>2</sup>/d.

38 128 For the modified design the organic load was 32 g COD/m<sup>2</sup>/d (i.e. 2.5 m<sup>2</sup> per person) resulting in  
39 129 a daily hydraulic loading of 1300 L/d. The eight loadings per day had a volume of 162 L each and  
40 130 were applied with a constant interval of 3 hours between loadings. Additionally, simulations were  
41 131 run for a high-load classical VF wetland, designed with a reduced specific surface area of 2.5 m<sup>2</sup>

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3 132 per person. The loadings were applied at 07:40, 11:40, 19:00 and 21:40 hours with the volume of  
4 133 a single load of 323 L.  
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6 134 Additionally, the two designs were compared for the high load phases according to ON EN 12556-  
7 135 3 (2005), i.e. a 14-days simulation was carried out. The load on the first two of the 14 days was  
8 136 150 % of the design load (Rizzo and Langergraber, 2016). Again, these simulations have been  
9 137 carried out again for different temperatures. The results from the previous simulations with design  
10 138 load have been taken as initial conditions for these simulations.  
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### 13 139 14 15 140 **Model set-up**

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17 141 The transport domain was delimited by a two-dimensional finite element mesh and a simple  
18 142 geometry representing a surface area of 20 m<sup>2</sup>. Considering the length of the VF bed of 500 cm,  
19 143 the 2D cross section had a width of 400 cm (horizontal plane) and a depth of 75 cm (vertical plane).  
20 144 The finite element mesh consisted of 1'421 nodes and 2'688 elements for the classical VF wetland  
21 145 design. Each opening was simulated by three nodes in the horizontal top plane that was set as  
22 146 atmospheric boundary condition. For the classical VF wetland design, 4 openings were distributed  
23 147 uniformly on the horizontal top plane. . A node at the bottom right was set as the constant head  
24 148 with a pressure head of -2 cm to simulate the water outlet of the system. For the modified design,  
25 149 the mesh consisted of 2'813 nodes and 5'376 elements and 4 openings were distributed uniformly  
26 150 on the horizontal top plane.  
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31 151 As proposed by Mollner (2005), only two filter materials have been considered. The intermediate  
32 152 layer and the drainage layer are considered as a single 25 cm thick layer. This approach is carried  
33 153 out to simplify the model, as the calibration of the soil hydraulic parameters with two layers (main  
34 154 and drainage) is well adjusted to the measured data and requires less calculation time. Moreover,  
35 155 it was assumed that the drainage layer properties did not change since 2005. The retention time of  
36 156 the water flow in the drainage layer is very short, so the main biological processes take place in  
37 157 the main layer and this is well simulated by the HYDRUS Wetland Module.  
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## 41 158 42 43 159 **RESULTS AND DISCUSSION**

### 44 45 160 **Soil hydraulic model parameters calibration**

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47 161 Material parameters have been shown to have the strongest influence on effluent concentrations  
48 162 (Langergraber, 2001). Therefore, a successful calibration of the hydraulic behavior determines the  
49 163 quality of the simulation of water quality parameters. According to Langergraber et al. (2010), a  
50 164 volume of 175 L was applied for each loading (700 L/d) during the period of 2008/09. This is the  
51 165 volume used for soil hydraulic parameters calibration.  
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54 166 The objective function for the inverse simulation to fit soil hydraulic parameters from Mollner  
55 167 (2005) was defined using measured effluent flow data obtained in 2008. Table 3 shows soil  
56 168 hydraulic parameters for main and drainage layers. The parameters for the soil hydraulic model as  
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obtained by Mollner (2005) were used as a starting point. Two inverse simulations were carried out to fit hydraulic parameters. Firstly, the residual water content  $Q_r$ , the shape parameters  $\alpha$  and  $n$ , and the tortuosity parameter  $l$  were fitted. A second simulation was carried out to fit hydraulic conductivity,  $K_s$ . Figure 1 shows a good match of simulation results compared to measured data for the daily loading pattern.

Table 3: Soil hydraulic parameters of the van Genuchten - Mualem model used for simulations and results of the simulations using Inverse Solution of HYDRUS.

Layers	Parameter	$Q_r$	$Q_s$	$\alpha$	$n$	$K_s$	$l$
	Unit	-	-	1/cm	-	cm/h	-
Main	Mollner (2005)	0.130	0.42	0.063	3.90	128	0.73
	Simulation 1	0.045 <sup>a</sup>	0.42	0.065 <sup>a</sup>	3.164 <sup>a</sup>	128	0.974 <sup>a</sup>
	Simulation 2	0.045	0.42	0.065	3.164	124.5 <sup>a</sup>	0.974
Drainage		0.045 <sup>b</sup>	0.43 <sup>b</sup>	0.145 <sup>b</sup>	2.68 <sup>b</sup>	150,000 <sup>c</sup>	0.50 <sup>b</sup>

<sup>a</sup> Fitted value

<sup>b</sup> Standard parameters for "sand" as provided by the HYDRUS software

<sup>c</sup> Mollner (2005)

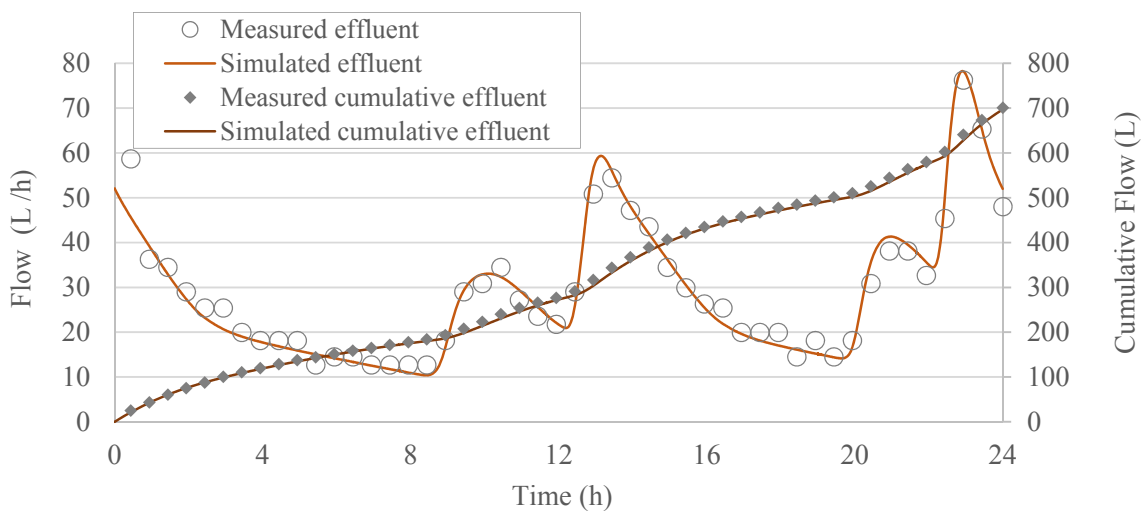


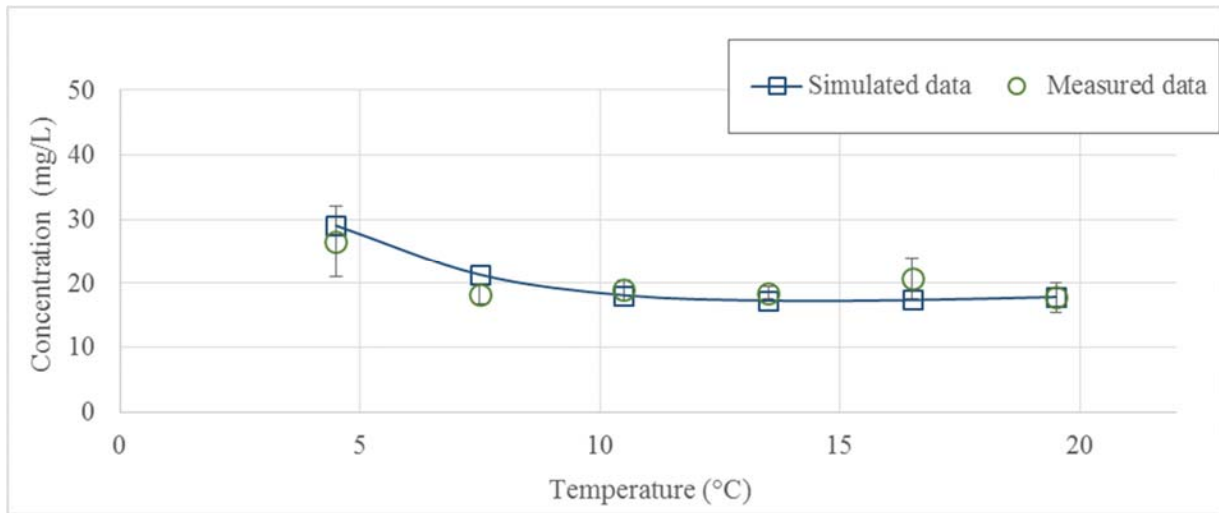
Figure 1: Measured and simulated data for effluent flow and cumulative effluent flow.

### Calibration of the reactive transport model

Default biokinetic model parameters of CW2D (Langergraber and Šimůnek, 2012) have been used for the simulations. Influent COD fractionation have been calibrated in order to adapt the model to the wastewater simulated in this study case. An inert COD (CI) of 7 mg/L has been used to match measured COD data at 19.5 °C by trial and error method. A CR: CS ratio (readily and slowly

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3 186 biodegradable soluble COD, respectively) of 2:1 was applied as described by Langergraber (2007)  
4 187 for the wastewater of the experimental plant.  
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6 188 Figure 2 shows measured COD effluent concentrations (median values) with error bars (95%  
7 189 confidence interval) compared to simulated COD effluent concentrations. Simulated effluent  
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9 190 concentrations follow the COD pattern measured for different temperature ranges well.



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192 Figure 2: Measured and calibrated COD effluent concentration for classical Austrian single-stage  
193 VF wetland. Error bars indicate the 95% confidence interval for measured data.

### 194 195 **Comparison of classical VF wetland with the modified design**

196 Table 4 and Table 5 show influent and simulated effluent concentrations for COD, NH<sub>4</sub>-N, NO<sub>2</sub>-  
197 N and NO<sub>3</sub>-N as well as soluble organic matter fractions as a function of temperature for classical  
198 single-stage VF wetland and modified design, respectively.

199 Table 4: Measured influent concentrations and simulated effluent concentrations for classical  
200 single-stage VF wetland.

Parameter	Temp.	CR	CS	CI	COD	NH <sub>4</sub> -N	NO <sub>2</sub> -N	NO <sub>3</sub> -N
Unit	°C	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
<b>Influent</b>	-	325	163	7	495	65	0.015	0.4
<b>Effluent</b>	4.5	0.1	13.5	15.4	29.1	4.09	0.017	56.3
	7.5	0.1	5.4	15.8	21.4	2.13	0.019	56.2
	10.5	0.1	1.7	16.3	18.2	1.01	0.019	55.2
	13.5	0.1	0.4	16.7	17.3	0.46	0.020	54.8
	16.5	0.2	0.1	17.2	17.4	0.25	0.022	53.8
	19.5	0.2	0.0	17.7	17.9	0.13	0.027	52.6

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202 Table 5: Measured influent concentrations and simulated effluent concentrations for modified  
 203 design.

Parameter	Temp.	CR	CS	CI	COD	NH <sub>4</sub> -N	NO <sub>2</sub> -N	NO <sub>3</sub> -N
Unit	°C	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
<b>Influent</b>	-	325	163	7	495	65	0.015	0.4
<b>Effluent</b>	4.5	0.1	20.2	14.3	34.6	4.22	0.018	54.9
	7.5	0.1	8.8	14.7	23.7	2.14	0.019	54.8
	10.5	0.2	3.1	14.9	18.1	1.02	0.019	53.3
	13.5	0.2	0.8	15.2	16.1	0.50	0.020	51.4
	16.5	0.2	0.1	15.4	15.7	0.25	0.022	49.2
	19.5	0.2	0.0	15.6	15.8	0.13	0.027	47.7

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 205 Figure 3 compares the simulated COD effluent concentrations of a classical VF wetland with  
 206 nominal load (20 g COD/m<sup>2</sup>/d), a high loaded single-stage VF wetland (32 g COD/m<sup>2</sup>/d) and a VF  
 207 wetland with modified design (32 g COD/m<sup>2</sup>/d). It shows - as expected - that by increasing the  
 208 input organic load per day, removal efficiency of organic matter at low temperatures is reduced.  
 209 Efficiency does not vary significantly at high temperatures. For the modified design, the constant  
 210 loading interval and an increase in the number of loadings in the modified design (3-4 to 8 loadings  
 211 per day) ensure lower hydraulic load at each single loading, lower flow velocities in the filter,  
 212 increased contact time between pollutant and bacteria and thus better COD removal. The higher  
 213 number of openings in the distribution pipes further contribute to these effects by a more even  
 214 distribution of the wastewater onto the surface of the VF wetland. At the lowest simulated  
 215 temperature (4.5 °C), effluent concentrations of COD are 29.1, 34.6 and 47.5 mg/L for classical  
 216 VF wetland design, modified design, and high-load VF wetland design, respectively. The results  
 217 are still far below the legal requirements required for COD effluent concentration (90 mg/L).

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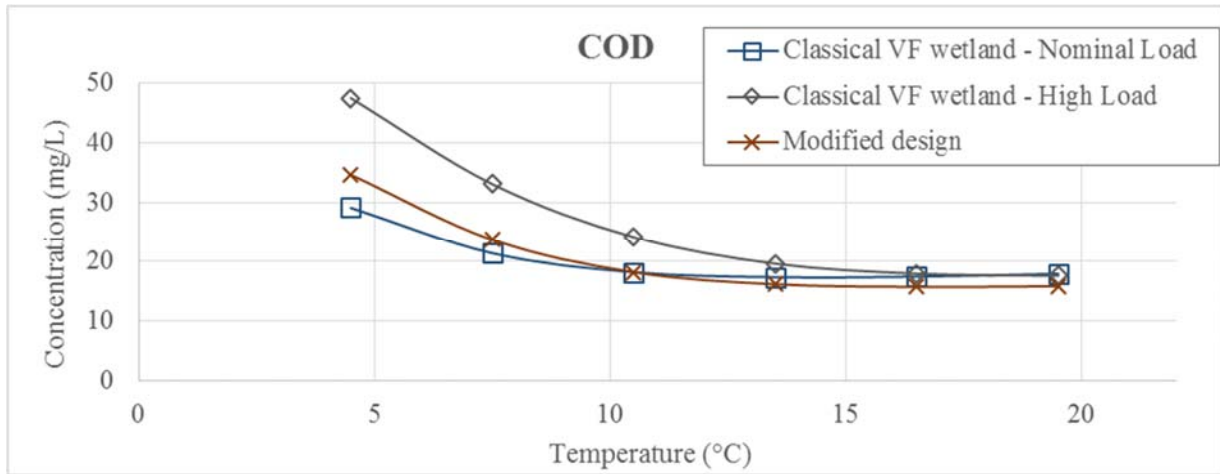


Figure 3: Comparison of COD effluent concentration for a single-stage VF wetland loaded with nominal loading (ÖNORM B 2505, 2009), high loading and the modified design, for different temperatures.

Regarding the  $\text{NH}_4\text{-N}$  effluent concentrations, the differences between the different designs are not significant. Figure 4 shows that simulated  $\text{NH}_4\text{-N}$  effluent concentrations remain close to zero and increase substantially as the temperature decreases in all compared designs. Simulated effluent concentrations meet legal requirements (10 mg  $\text{NH}_4\text{-N/L}$ ) even at all simulated temperatures below 12 °C.

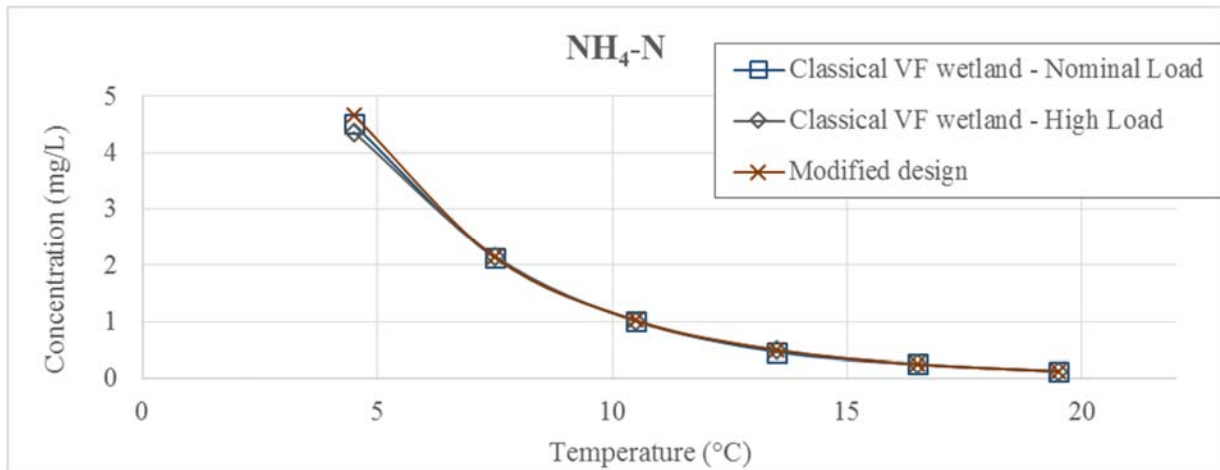
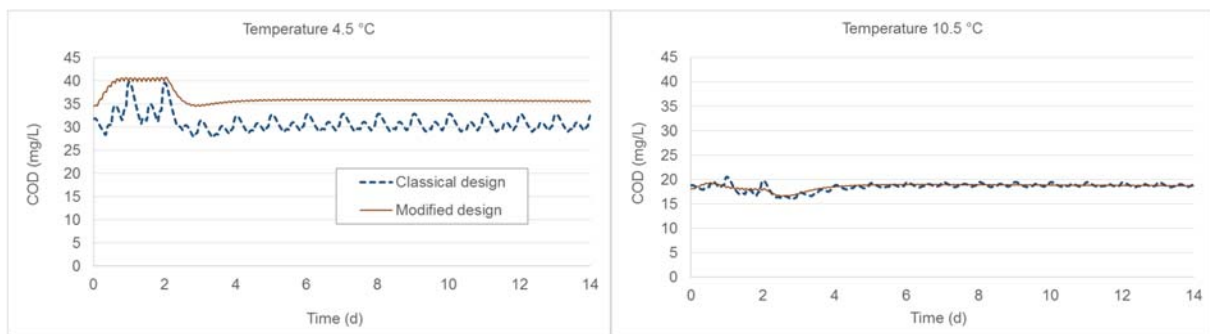


Figure 4: Comparison of  $\text{NH}_4\text{-N}$  effluent concentration for a single-stage VF wetland loaded with nominal loading (ÖNORM B 2505, 2009), high loading and the modified design, for different temperatures.

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3 235 Álvarez et al. (2017) presented results from two parallel operated VF beds in India. They found  
4 236 lower COD and NH<sub>4</sub>-N effluent concentrations in the VF bed that had less but higher single loads  
5 237 compared the VF bed that received more frequent but lower doses. However, no details on the  
6 238 loading regime (intervals between single doses and volume of a single dose) were presented. It  
7 239 seems that a loading interval much less than 3 hours (as used for the modified design) was applied  
8 240 and led to limited oxygen exchange and reduced nitrification capacity, respectively. More frequent  
9 241 loadings and smaller single doses as applied at the modified design did not result in limited oxygen  
10 242 exchange and reduced treatment efficiency.

14 243 Figure 5 and Figure 6 compare the behaviour of the single-stage VF wetlands during the high-load  
15 244 phase according to ON EN 12556-3 (2005). The COD effluent concentrations (Figure 5) are  
16 245 similar for higher temperatures. For lower temperatures, the more frequent loading of the VF  
17 246 wetland with modified design results in less fluctuations of the COD effluent concentrations  
18 247 compared to the classical design.  
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35 250 Figure 5: COD effluent concentration for a single-stage VF wetland with classical design  
36 251 (ÖNORM B 2505, 2009) and modified design during the high-load phase according to  
37 252 ON EN 12556-3 (2005) for 4.5 °C (left) and 10.5 °C (right).  
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41 254 For the NH<sub>4</sub>-N effluent concentrations the difference is more important at higher temperatures.  
42 255 During the 2-days high-loading the NH<sub>4</sub>-N effluent concentration for the VF wetland at 10.5 °C  
43 256 increases from about 1 mg/L up to 4 mg/L whereby for the modified design the maximum effluent  
44 257 concentration is only about 2.5 mg/L (Figure 6, right). This can be explained by the uniform  
45 258 loading intervals and thus more uniform of the wastewater distribution for the modified design. At  
46 259 cold temperatures, the behaviour of the VF wetlands with classical and modified design are similar  
47 259 (Figure 6, left).  
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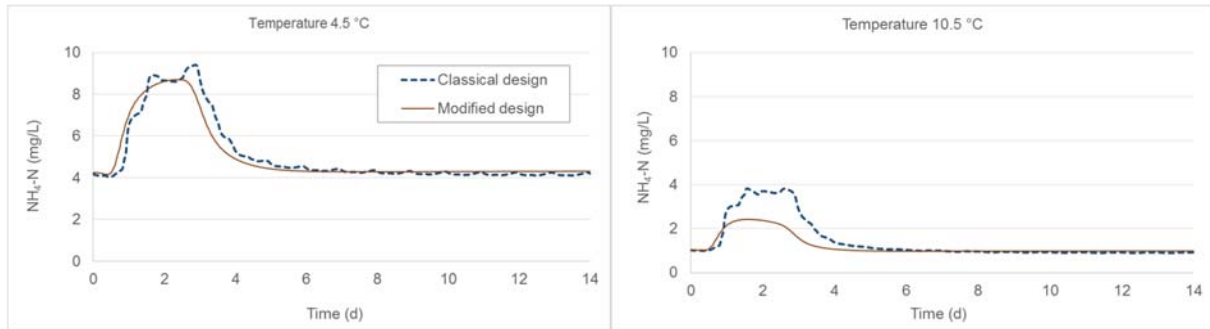


Figure 6: NH<sub>4</sub>-N effluent concentration for a single-stage VF wetland with classical design (ÖNORM B 2505, 2009) and modified design during the high-load phase according to ON EN 12556-3 (2005) for 4.5 °C (left) and 10.5 °C (right).

## CONCLUSIONS

From the simulation study, it can be concluded that:

1. The model of the classical single-stage Austrian VF wetland could be calibrated well to measured effluent water flow and effluent concentrations, respectively.
2. By increasing the organic loading for the classical Austrian VF wetland, the COD effluent concentrations are significantly higher for lower temperatures.
3. By applying the modified design, lower COD effluent concentrations can be reached again.
4. During high-loading events at higher temperatures, VF wetlands with modified design have lower maximum NH<sub>4</sub>-N effluent concentrations.
5. The main modifications in the design, i.e. more loadings per day but lower volume of a single loading, constant loading interval, and increased number of openings per m<sup>2</sup>, seem to be very effective to allow applying higher organic loads.

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