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Additional Information

1 Metagenomic analysis of viruses, bacteria and protozoa in irrigation water

2

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15

16 **Keywords:** Irrigation water, Metagenomics, Virus, Bacteria, Protozoa

17

18 Abstract

19 Viruses (e.g., noroviruses and hepatitis A and E virus), bacteria (e.g., *Salmonella spp.* and pathogenic
20 *Escherichia coli*) and protozoa (e.g., *Cryptosporidium parvum* and *Giardia intestinalis*) are well-
21 known contributors to food-borne illnesses linked to contaminated fresh produce. As agricultural
22 irrigation increases the total amount of water used annually, reclaimed water is a good alternative to
23 reduce dependency on conventional irrigation water sources. European guidelines have established
24 acceptable concentrations of certain pathogens and/or indicators in irrigation water, depending on the
25 irrigation system used and the irrigated crop. However, the incidences of food-borne infections are
26 known to be underestimated and all the different pathogens contributing to these infections are not
27 known. Next-generation sequencing (NGS) enables the determination of the viral, bacterial and

28 protozoan populations present in a water sample, providing an opportunity to detect emerging
29 pathogens and develop improved tools for monitoring the quality of irrigation water. This is a
30 descriptive study of the virome, bacteriome and parasitome present in different irrigation water
31 sources. We applied the same concentration method for all the studied samples and specific
32 metagenomic approaches to characterize both DNA and RNA viruses, bacteria and protozoa.

33 In general, most of the known viral species corresponded to plant viruses and bacteriophages. Viral
34 diversity in river water varied over the year, with higher bacteriophage prevalences during the autumn
35 and winter. Reservoir water contained *Enterobacter cloacae*, an opportunistic human pathogen and an
36 indicator of fecal contamination, as well as *Naegleria australiensis* and *Naegleria clarki*. Hepatitis E
37 virus and *Naegleria fowleri*, emerging human pathogens, were detected in groundwater. Reclaimed
38 water produced in a constructed wetland system presented a virome and bacteriome that resembled
39 those of freshwater samples (river and reservoir water). Viral, bacterial and protozoan pathogens were
40 occasionally detected in the different irrigation water sources included in this study, justifying the use
41 of improved NGS techniques to get a comprehensive evaluation of microbial species and potential
42 environmental health hazards associated to irrigation water.

43 **1. Introduction**

44 Agricultural irrigation accounts for 36% of the total annual water usage in Europe, reaching up to
45 80% in some parts of the Mediterranean region (European Environment Agency, 2012). In Spain 79%
46 of the total irrigated area (3Mha) is irrigated with surface water and only 21% with groundwater
47 (Drewes et al., 2017). River flow rate fluctuation and the overexploitation of groundwater resources
48 are existing problems in Europe, where integrated water resource plans are currently being
49 implemented. Moreover, climate change is expected to intensify problems of water scarcity and affect
50 irrigation requirements in the Mediterranean region (Collins et al., 2009). To reduce dependency on
51 freshwater, reclaimed water can be used for irrigation, provided that the risk associated with pathogen
52 contamination can be minimised. Reclaimed water can be relatively nutrient rich and may reduce the
53 need for additional applications of inorganic fertilizers (Parsons et al., 2001). In Catalonia, around 300

54 wastewater treatment plants (WWTP) treat 700 hm³ of water every year to produce 204 hm³ of
55 reclaimed water (ACA, 2016; Pérez et al., 2011). The use of reclaimed water sources is regulated by
56 different national regulations and European guidelines, which stipulate the acceptable microbial
57 concentrations in irrigation water based on the irrigation system being used and the irrigated crop (RD
58 1620, 2007; WHO, 2006). In 2017, the European Commission published the minimum quality
59 requirements for water reuse in agricultural irrigation and aquifer recharge, producing a guidance
60 document addressing microbiological risks in the primary production of fresh fruits and vegetables
61 (Alcalde-Sanz and Gawlik, 2017; EU C163, 2017).

62

63 The number of reported illnesses linked to contaminated fresh produce has increased in both
64 developed and developing countries, with the majority of reported outbreaks being in Europe
65 (Chatziprodromidou et al., 2018). Viruses (e.g., noroviruses and hepatitis A and E virus), bacteria
66 (e.g., *Salmonella spp.* and pathogenic *Escherichia coli*) and protozoa (e.g., *Cryptosporidium parvum*
67 and *Giardia intestinalis*) are well-known contributors to these reported food-borne diseases; however,
68 the number of food-borne infections is known to be underestimated and all the different pathogens,
69 including new and emerging strains, contributing to these infections are not known. The risk
70 associated with the consumption of fresh produce irrigated with water containing pathogens depends
71 on many variables, the most important being the concentration of pathogens in the irrigation water
72 (Gonzales-Gustavson et al., 2019). The levels of exposure to solar radiation, the irrigation system, the
73 time from irrigation to consumption and the type of crop are other important factors (Pachepsky et al.,
74 2011; Thebo et al., 2017). If crops are irrigated with reclaimed water, the concentration loads of the
75 pathogens and their specific persistence in the environment are also important factors.

76 Current regulations and monitoring programs involve the measurements of selected microbial
77 indicators, which do not necessarily correlate with the reference pathogens (Drewes et al., 2017).
78 Therefore, more resistant pathogens, zoonotic microorganisms excreted by livestock or even
79 uncharacterized pathogenic strains may also be present or underestimated in irrigation water.

80

81 While pathogen-specific tests for water quality assessment are appropriate for estimating irrigation
82 water risks, next-generation sequencing (NGS) techniques can detect a wide range of microbial
83 populations in a given sample. Using NGS technologies, most studies have focused on bacterial
84 communities instead of viral or protozoan populations in different environments. Several studies have
85 demonstrated that sewage reflects the bacteriome (Newton et al., 2015), virome (Cantalupo et al.,
86 2011; Fernandez-Cassi et al., 2018; Ng et al., 2012) and parasitome (Bradley et al., 2016) of human
87 populations. Different studies have also characterized virus, bacteria or protozoa diversity in
88 groundwater (D'Auria et al., 2018), river water (Uyaguari-Diaz et al., 2016), and reclaimed water
89 (Rosario et al., 2009). Metagenomic methods have been used to tackle viruses and bacteria found in
90 fresh fruits and vegetables surfaces, highlighting the potential role of irrigation water as vector of
91 waterborne pathogens (Aw et al., 2016; Fernandez-Cassi et al., 2017; Jackson et al., 2013; Leff and
92 Fierer, 2013). Viruses lack a common marker for their taxonomical classification and identification,
93 making difficult to monitor the entire population in a single assay. Uyaguari-Diaz and colleagues
94 selected the gene 23 (encoding the major capsid protein of T4 bacteriophages) and the gene encoding
95 the RNA-dependent RNA polymerase (present in most RNA viruses) as the target amplicons for viral
96 metagenomics (Shi et al., 2016; Uyaguari-Diaz et al., 2016). As with the use of 16S or 18S ribosomal
97 RNA analysis to identify bacteria or protozoa, respectively, the use of a common marker for
98 identifying viruses might result in an imprecise taxonomical classification due to the poor resolution
99 caused by using a single gene (Poretzky et al., 2014). Viruses show great diversity in their genome
100 structure and organization. Thus, important waterborne pathogenic DNA viruses (e.g. adenovirus)
101 lacking an RNA-dependent RNA polymerase might be overlooked when using this marker to identify
102 viruses. In this study, we applied a metagenomic approach to characterize the viral populations in
103 different water sources used for irrigation, including both DNA and RNA viruses. To our knowledge,
104 this is the first study to assess the virome, bacteriome and parasitome of the same irrigation water
105 samples using metagenomics after a single concentration method for all the studied microorganisms.

106
107
108

109 2. Materials and methods

110 2.1. Water samples and concentration of microorganisms using the skimmed milk flocculation 111 method (SMF).

112 Four conventional irrigation water sources (drinking water, reservoir water, groundwater and river
113 water) and one reclaimed water source, currently used for irrigation, were selected for the study. Raw
114 sewage was included as a reference of the microbial contamination circulating within the studied area.
115 Samples were collected every month over a period of one year, from April 2016 to March 2017.

116 Drinking water was sampled directly from water tanks across the water distribution network of
117 Barcelona. This water source is frequently used in urban community orchards. Reservoir water, a
118 common source of irrigation water in the Catalan territory, was collected from the Foix river
119 reservoir. Groundwater, which is a conventional irrigation water source in the Mediterranean region,
120 was sampled from the north-east of Catalonia, a region with intensive farming and agricultural
121 activities. River water samples were collected from the Fluvià river, which receives effluents from 24
122 small WWTPs treating up to 100,000 Hab/Eq. The orchard sites surrounding the river are supplied
123 directly with water diverted or pumped from the river.

124 The outlet water from a sustainable wetland system treating the urban WWTP effluents, was selected
125 as the reclaimed water source. The WWTP uses mechanical treatment to remove sand and large
126 particles and a conventional activated sludge to digest the organic matter. About 70% of the
127 secondary effluent is chlorinated and directly discharged into the River Congost, while the remaining
128 30% is discharged into the constructed wetland system where its nitrogen and phosphorus contents are
129 reduced by natural plant absorption. This system produces 120,000 m³ of recycled water every year,
130 which nowadays is mainly used to irrigate green areas. Raw sewage was directly collected from the
131 urban WWTP.

132 All microorganisms were concentrated using a standardized SMF protocol (Calgua et al., 2013a,
133 2013b; Gonzales-Gustavson et al., 2017). This protocol achieves high recovery efficiencies, with low
134 variability for selected fecal indicators and a wide range of bacterial, viral and parasitic pathogens

135 (Gonzales-Gustavson et al., 2017). Briefly, irrigation water samples (10L), as well as raw sewage
136 (500ml), were acidified to a pH 3.5, using 1 N HCl. Conductivity was also measured and adjusted
137 with artificial sea salt (Sigma, Aldrich Chemie GMBH, Steinheim, Germany) to achieve a minimum
138 conductivity of 1.5 mS/cm². A pre-flocculated skimmed milk solution (PSM) was prepared by
139 dissolving 10 g of skimmed milk powder (Difco-France) in 1 L of artificial seawater (33,33 gr of sea
140 salts), adjusting the pH to 3.5. The PSM was then added to the samples to obtain a final concentration
141 of 0.01 % of skimmed milk. The samples were stirred for 8 h at room temperature, and the flocs were
142 allowed to settle by gravity for another 8 h. The supernatants were then removed, and the sediment
143 was collected and transferred to 500-mL centrifuge containers and centrifuged at 8,000 g for 30 min at
144 4 °C. Pellets were suspended in 5 mL of 0.2 M phosphate buffer, pH 7.5 (1:2, v/v of 0.2 M Na₂HPO₄
145 and 0.2 M NaH₂PO₄). The samples were pooled according to the season they were collected in,
146 distributed in refrigerated boxes among the collaborating laboratories and stored at -20 °C.

147 *2.2. Viral DNA and RNA pre-NGS amplification and bioinformatics analysis*

148 Season-pooled concentrates (900 µL) were filtered through 0.45 µm Sterivex filters (Millipore). Free
149 viral DNA and non-viral DNA was removed using TURBO DNA-free Kit (Ambion). The nucleic
150 acids (NA) of viral origin were then extracted with the Viral RNA Mini Kit (Qiagen), without the
151 carrier RNA. Sequence-independent, single-primer amplification (SISPA) was applied, as previously
152 described (Fernandez-Cassi et al., 2018). Briefly, NAs were retrotranscribed using SuperScript III
153 (Life Technologies) and tagged with random Primer A for the detection of both RNA and DNA
154 viruses (Table 1). The second cDNA strand was constructed using Sequenase version 2.0 Kit
155 (USB/Affymetrix). PCR amplification (10 min at 95°C, 25 cycles of 30 s at 94°C, 30 s at 40°C, and
156 30 s at 50 °C, and a final step of 1 min at 70°C) was performed using the AmpliTaq Gold Master Mix
157 (Life Technologies) and Primer B to obtain enough dsDNA for library preparation (Table 1). PCR
158 products (100 µL) were cleaned and concentrated into smaller volumes (15 µL) with the DNA Clean
159 & Concentrator Kit (Zymo research). Amplified DNA samples were quantified using Qubit 2.0
160 fluorometer (Life Technologies, Oregon, USA) and libraries were constructed using a Nextera XT
161 DNA Sample Preparation kit (Illumina Inc.) following the manufacturer's instructions. Samples were

162 sequenced on an Illumina MiSeq platform, 2×300 bp, producing paired-end reads.

163 The quality of the raw and clean sequences was assessed using the FASTX-Toolkit version 0.0.14
164 software (Hannon Lab; <http://www.hannonlab.org>). The sequences were cleaned using Trimmomatic
165 version 0.32 (Bolger et al., 2014), with the trimming of low-quality reads presenting a Phred score
166 above Q15 over a running window of 4 nucleotides. Duplicated reads were removed in a subsequent
167 step to accelerate the downstream assembly. Viral reads were assembled into contigs, based on 90%
168 identity over a minimum of 50% of the read length (coverage), using CLC Genomics Workbench 4.4
169 (CLC bio USA). The resulting contigs were queried for sequence similarity, using NCBI-BLASTN
170 against the NCBI Viral Genomes database (Brister et al., 2015) and the viral sequences from the
171 GenBank database (Benson et al., 2015). On the other hand, NCBI-BLASTX was also run to compare
172 the translated nucleotide contig sequences against the viral protein sequences from the UniProt
173 database (UniProt Consortium, 2015). Based on the best BLAST result from those three comparisons
174 on each contig, considering a minimum 90% identity cut-off and an E-value lower than 10^{-5} , each
175 sequence was classified into its likely taxonomic group.

176 *2.3. NGS of bacterial populations*

177 Total DNA was extracted from 300 μ l of the SMF concentrates using the DNeasy PowerSoil kit
178 (Qiagen), following the manufacturer's instructions and was measured using Qubit 2.0 (Life
179 Technologies, Invitrogen, Oregon, USA). The metagenomic analysis targeted the V3-V4 region of the
180 16S rRNA gene with previously described primers (Claesson et al., 2009; Herlemann et al., 2011) and
181 the overhang adaptor sequences from Illumina were added to the locus-specific primer for the targeted
182 region (Table 1). Paired-end libraries were built using the Nextera XT DNA sample preparation kit
183 (Illumina Inc.) and sequenced on an Illumina MiSeq platform, 2×250 bp. End reads with a quality
184 Phred score above Q30 were trimmed with QIIME 1.9.1. (Caporaso et al., 2010) and the cleaned reads
185 were merged with the FLASH software (Magoč and Salzberg, 2011). All sequences with ambiguous
186 base calls were also discarded. Taxonomic classification was performed using the SILVA database
187 (Quast et al., 2012), with a 95% cut-off for sequence identity.

188 *2.4. NGS of protozoan populations*

189 A volume of 300 µl of each SMF concentrate was lysed using the FastPrep®-24 instrument (MP
190 Biomedicals). DNA was extracted with the FastDNA® SPIN Kit (MP Biomedicals) for soil, according
191 to the manufacturer's instructions. A modified protocol for the lysing step was followed, as described
192 previously (Shields et al., 2013). Samples were first homogenized for 60 s. After the bead beating
193 step, samples were placed on ice for 1 min and then homogenized for another 60 s. The final DNA
194 products were eluted in a final volume of 50 µL. All the DNA extractions were performed in
195 duplicate. Afterwards, the duplicate samples were mixed and purified with the OneStep™ PCR
196 Inhibitor Removal Kit (Zymo Research) to remove inhibitors that could have affected downstream
197 enzymatic reactions such as the PCR. Genomic DNA concentrations were measured with the Qubit®
198 dsDNA BR Assay Kit (Thermo Fisher Scientific). The 260:280 ratio was measured using the
199 NanoDrop ND-1000 UV/Vis spectrophotometer (Thermo Fisher Scientific). The amplicon sequencing
200 protocol for protozoa, targeting the hypervariable V4 region of the 18S rRNA gene, was performed as
201 previously described (Moreno et al., 2018), using the primers described in Table 1. Briefly, DNA
202 amplicon libraries were built using the Nextera XT DNA Sample Preparation Kit (Illumina Inc.) and
203 sequenced on an Illumina MiSeq platform. The obtained reads were analyzed using QIIME 1.9.1
204 (Caporaso et al., 2010), applying additional scripts that are available in the Microbiome Helper
205 Virtual Box (Comeau et al., 2017). OTUs were defined at the 97% similarity cut-off. The Protist
206 Ribosomal Reference database pr2_gb203_version_4.5 fasta (Guillou et al., 2013) was used as the
207 reference.

208 *2.5. Statistics and heatmaps*

209 The richness and diversity indices were calculated using the Catchall version 4.0 software (Bunge et
210 al., 2012), which included the species richness estimate (Chao1) and the Shannon diversity index as
211 measures of the number of different species detected and the relative abundances of the species
212 present in a sample, respectively. Heatmaps were generated using the R library ggplot2 graphics
213 (Kolde, 2015).

214

215 3. Results

216 3.1. Output of the metagenomic libraries

217 In this study, the viral, bacterial and protozoan communities present in the different irrigation water
218 sources were evaluated using a metagenomic approach. The total number of reads obtained as well as
219 the richness and diversity indices are summarized in Table 2. In total, 95.13% of the Illumina viral
220 reads were assembled into contigs. The highest viral richness and diversity were observed in raw
221 sewage samples, particularly those collected in winter. Drinking water samples presented the lowest
222 number of viral reads compared to the other water sources. Regarding bacterial populations, reservoir
223 water samples as well as river water samples collected in the summer and spring showed the highest
224 richness and diversity values. For the reservoir, groundwater and river water samples, the highest
225 Chao values and Shannon indices for bacteria were obtained in the samples collected during summer
226 (Table 2). After quality filtering, trimming and detection of PCR-chimeras, the highest number of raw
227 reads (110.766) obtained from the analysis of protozoa mock community was obtained for the pooled
228 samples of reclaimed water collected in the summer season. Due to the wide range of raw reads,
229 obtained from the 18S rRNA amplicon-based analysis, and the relatively low abundances of
230 pathogenic protozoa, it was not possible to calculate richness and diversity values.

231

232 3.2. Virome of the different irrigation water sources and raw sewage

233 The diversity of the viral species detected is graphically represented in Figure 1 according to the viral
234 family and water source. Human viral pathogens were detected in all the different types of samples
235 analyzed, except for drinking and reservoir water (Table 3). Over 67% of the viral metagenomic
236 sequences showed significant similarity to the viral sequences stored in one or more of the three
237 databases used (NCBI Viral Genomes database, GenBank and UniProt). Most of the known viral
238 species (> 77%) detected in the conventional irrigation water sources were plant viruses or
239 bacteriophages. Bacteriophages belonging to the *Siphoviridae* (dsDNA), *Myoviridae* (dsDNA),
240 *Podoviridae* (dsDNA) and *Microviridae* (ssDNA) families were always the most represented in all the
241 irrigation water sources.

242 The 95% of the viral reads from the drinking water samples corresponded to phages infecting
243 different bacteria, including *Arthrobacter* spp., *Propionibacterium* spp. and *Pseudomonas* spp. Up to
244 39 different species belonging to 6 phage families were detected in this type of water. NGS did not
245 detect any waterborne pathogens in drinking water (Table 3), with only a few contigs identified that
246 corresponded to the *Myoviridae* and *Microviridae* phage families and *Phycodnaviridae*.
247 *Phycodnaviridae* (dsDNA), which includes viruses that infect freshwater eukaryotic algae and
248 phytoplankton, was very prevalent in the reservoir, groundwater and river water samples. Viruses
249 belonging to *Virgaviridae* (ssRNA+) and *Tombusviridae* (ssRNA+) were highly abundant in the river
250 water samples collected in the autumn and winter, whereas a greater diversity of viruses from the
251 *Phycodnaviridae* family were detected in groundwater in summer and river water in spring. A large
252 number of viruses affecting insects, including members of the *Iridoviridae* (dsDNA) and *Parvoviridae*
253 (ssDNA) families that infect cockroaches, flies and mosquitoes, were mainly detected in the river and
254 reservoir water samples. The drinking water samples had higher proportions of known viral sequences
255 (77%) than river (68%) or reclaimed water (67%) samples; however, a large fraction of the viral reads
256 could not be assigned. Reclaimed water samples presented a mixture of bacteriophages, plant viruses
257 and viruses infecting invertebrates (e.g., aphids and flies), but no human pathogens. Less than 1% of
258 the viral reads corresponded to non-pathogenic viruses that infect humans. More than 30 different
259 viral families were detected in the urban sewage samples, with the majority of the contigs belonging
260 to the *Siphoviridae* and *Myoviridae* families, which showed the highest diversity in the samples
261 collected during the summer (167 and 103 different species, respectively) and winter (172 and 144
262 species, respectively). When analyzing the data obtained from the raw sewage samples pooled
263 according to season (Figure 1), pathogenic viruses belonging to the *Caliciviridae* and *Astroviridae*
264 families were present throughout the year, but the number of contigs showed seasonality, being higher
265 during the autumn and winter (Table 3). A high diversity of human astrovirus contigs were detected in
266 the raw sewage samples collected in the winter and autumn (maximum length, 2,483 bp). Human
267 adenovirus 41 was detected in raw sewage, as well as in groundwater samples collected during the
268 autumn. Norovirus GI and GII as well as sapovirus GI, GIV and GV, which are important human
269 pathogens belonging to the *Caliciviridae* family, were found throughout the year in raw sewage

270 samples. However, the number of species, as well as the number of detected reads, was higher during
271 the winter and autumn. Norovirus GII was also detected in the river water samples collected during
272 the autumn, corresponding with the peak excretion season. A 396-bp contig belonging to the beta
273 human papillomavirus 104 was found in the raw sewage sample collected during spring. Within the
274 *Picornaviridae* family, sequences belonging to human enteroviruses A, B and C were detected in the
275 raw sewage samples throughout the year. Remarkably, a 653- and an 867-bp contig were detected that
276 corresponded to EV-71, showing 86% and 94.8% identity over 99% of the sequence coverage,
277 respectively. Among the viruses belonging to the *Picornaviridae* family that could cause
278 gastroenteritis, the Aichi virus presented greater abundance during the summer, whereas salivirus was
279 more prevalent during the winter. Interestingly, hepatitis E virus (HEV) was detected in the raw
280 sewage and groundwater samples collected in the summer.

281

282 *3.3. Bacteriome of the different irrigation water sources and raw sewage*

283 The diversity of the bacteriome varied among the samples. Drinking water as well as samples of
284 groundwater, river water and reclaimed water collected in the winter and autumn presented the lowest
285 diversity values (Table 2). Raw sewage diversities did not vary across seasons. Figure 2 shows the
286 relative abundances of several bacteria that could be related with human infections in each water
287 sample.

288 The most abundant bacterial genus in drinking water, accounting for 33.2% of the bacterial reads, was
289 *Pseudomonas* (Figure 2). The genera *Cellvibrio* (27.0%), *Anoxybacillus* (5%) and *Caulobacter* (5%),
290 which are associated with plant rhizospheres, streams and fresh water, were also prevalent in drinking
291 water.

292 Most of the bacterial reads (18%) for reservoir water samples corresponded to genera representing
293 less than 1% abundance in the samples, thus demonstrating high bacterial diversity and richness. Only
294 the *Bacillus* (17%) and *Enterobacter* (6%) genera presented relatively high abundances in this type of
295 irrigation water source (Figure 2). The other taxa represented by 4 to 6% of the reads were *Raoultella*,
296 *Actinobacter* and non-culturable bacteria.

297 Regarding waterborne human pathogens, groundwater samples collected in the autumn presented the
298 highest abundance of *Enterobacter* (34.9%) (Figure 2), while the genera *Geobacillus* (20%) and
299 *Thermus* (18.1%) were abundant in the samples collected in the spring. The genus *Pantoea* showed a
300 relative abundance of 38.1% in the samples collected in winter, but less than 1% abundance in the
301 samples from the other seasons. Different bacterial genera with relative abundances ranging from 9%
302 to 14% were observed in the samples collected in summer, which showed the highest values for
303 bacterial richness and diversity compared to the samples collected in the other seasons (Table 2). The
304 percentage of non-culturable bacteria ranged from 0.65% in winter to 4.8% in spring.

305 River water samples also presented variations in bacterial richness and diversity between the seasons.
306 *Limnohabitans*, common planktonic bacteria occurring in river streams, was detected in low
307 percentages across the seasons (3.9%), but showed increased prevalence in the autumn, with over a
308 third of the total number of reads corresponding to this bacterial genus (32.5%). The human-
309 associated *Bacillus*, *Pseudomonas* and *Yersinia* genera showed low abundances throughout the year,
310 but increased prevalence in the winter (maximum abundances of 23.9%, 21.6% and 12.4%,
311 respectively) (Figure 2). Similarly, *Flavobacterium* presented increased relative abundance during the
312 autumn (21.1%). *Enterobacter* and *Serratia* were detected in all the river water samples, but at lower
313 percentages (Figure 2).

314 The most prevalent bacteria in reclaimed water was *Actinobacteria* (10.8% in spring to 2.5% in
315 autumn), with bacteria belonging to the *Limnohabitans* genus being relatively abundant throughout
316 the year. The number of reads corresponding to Rickettsiales bacteria increased only in the winter, but
317 none of these were associated with pathogenic species. The most prevalent human pathogens were
318 *Flavobacterium* and *Serratia* bacteria, showing peak abundances during the winter (Figure 2).

319 The number of human-associated bacterial genera ranged from 10 in the winter to 6-7 in the other
320 seasons for raw sewage samples. Among these genera, *Arcobacter* was the most abundant, with
321 percentages ranging from 14.9% in the autumn to 34.9% in the spring. However, *Arcobacter* showed
322 a generally high prevalence throughout the year, being more prevalent in the spring and summer and
323 showing a small decrease in the colder seasons. *Pseudomonas* was more prevalent in the summer and
324 winter, decreasing to 1% abundance in the spring. The other less abundant genera were *Acinetobacter*,

325 *Aeromonas*, *Bacteroides* and *Streptococcus*, with some of these containing bacteria that are
326 pathogenic to humans. Differences within a genus were observed between the seasons. For example,
327 the genus *Yersinia* showed an abundance of 7.4% in the winter, but an abundance of less than 0.1% in
328 the other seasons (Figure 2).

329

330 3.4. Parasitome of the different irrigation water sources

331 The waterborne protozoan pathogens detected in raw sewage and the different irrigation water sources
332 are summarized in Table 4. *Blastocystis* sp. was the most prevalent protozoa, while different free-
333 living amoeba (FLA) such as *Acanthamoeba* spp, *Entamoeba coli*, *E. dispar*, *E. moshkovskii*,
334 *Naegleria fowleri*, *N. australiensis*, *N. clarki* and members of the *Hartmannellidae* family (including
335 *Vermamaoeba vermiformis*) were also detected. *Naegleria* spp. were most frequently detected in
336 reservoir water, groundwater, river water and reclaimed water samples collected during the summer.
337 *N. fowleri*, the causative agent of primary amoebic meningoencephalitis (PAM), was detected in
338 groundwater. *Cryptosporidium* spp. was detected in reclaimed water samples collected during the
339 summer at a low relative abundance (< 0.05%). *Giardia* was not detected in any of the samples
340 analyzed.

341

342 4. Discussion

343

344 The metagenomic approach enables the description of viral, bacterial and protozoan populations in a
345 water sample, providing an opportunity to detect emerging pathogens and develop improved tools for
346 monitoring the quality of irrigation water. However, it is not an exhaustive detection method, since
347 pathogens detected by other molecular tools (e.g., q(RT)-PCR) can be missed (Fernandez-Cassi et al.,
348 2018). Moreover, the detection of pathogens using metagenomics may be hampered by low target
349 abundances and/or by higher microbial concentrations in the sample. The quantification of bacterial,
350 protozoan or viral abundances based on the number of reads is in fact a relative number because the
351 resulting data only indicate values in relation to the total number of microorganisms in a specific
352 sample. Since metagenomics assays are not suitable for quantification, these results are intended to

353 indicate the global distribution of families/species and the relationship between diversity and
354 seasonality.

355 *4.1. Raw sewage, the main source of microbial contamination*

356 Season-pooled samples of raw sewage were analyzed as sewage is the most important source of
357 environmental contamination and has been demonstrated to be a powerful tool for epidemiological
358 surveillance and risk evaluation (Fernandez-Cassi et al., 2018; Newton et al., 2015). As previously
359 reported, the majority of viruses identified in the raw sewage virome were bacteriophages (Bibby and
360 Peccia, 2013; Cantalupo et al., 2011). The low abundance of known human pathogenic viral families
361 (less than 1%) in raw sewage has also been reported in the literature (Cantalupo et al., 2011;
362 Fernandez-Cassi et al., 2018).

363 Our metagenomic approach identified important human pathogens in raw sewage that are transmitted
364 via the fecal-oral route. Human enteroviruses A, B and C, adenoviruses, astroviruses, caliciviruses
365 and picornaviruses, which mainly cause gastroenteritis, were detected in raw sewage throughout the
366 year. The higher abundance and diversity of astroviruses and caliciviruses during the colder seasons
367 are consistent with their well-documented seasonal peak that has been demonstrated by other
368 molecular tools (Bosch et al., 2014; Haramoto et al., 2005). HEV, an emerging zoonotic virus that
369 causes acute hepatitis, was detected in raw sewage samples collected during the summer. Although it
370 was occasionally detected, it is still important to report its presence as there is an increasing number
371 of foodborne HEV cases being reported in Europe (EFSA BIOHAZ Panel et al., 2017).

372

373 The *Arcobacter* and *Pseudomonas* genera were the most abundant bacterial populations observed in
374 raw sewage (Figure 2). Whereas *Pseudomonas* was more prevalent in the summer and winter,
375 *Arcobacter* showed a higher prevalence throughout the year, as has been stated by other studies (Tang
376 et al., 2016). Most of the bacterial reads corresponded to *A. cryaerophilus*, which is associated with
377 human and animal disease and has been previously reported to be the predominant *Arcobacter* species
378 occurring in raw sewage (Ferreira et al., 2015; Figueras et al., 2014; Fisher et al., 2014; Levican et al.,
379 2016; Salas-Massó et al., 2018).

380 Other pathogenic bacteria such as *Bacteroides* and *Aeromonas* also occurred at high percentages in all
381 the raw sewage samples analyzed, being more prevalent in those collected in the spring. Interestingly
382 *Aeromonas* and *Arcobacter* have been found predominating in metagenomics analysis of wastewater
383 and it have been indicated that this is due to the capacity of this bacteria to multiply in the sewerage
384 system (McLellan et al., 2010; McLellan and Roguet, 2019).

385 *Blastocystis* spp. were the most prevalent protozoa detected in raw sewage although it should be noted
386 that the rRNA copies of 18S used to identify protozoa vary in different types of waterborne protozoa
387 (Moreno et al., 2018). The effects of *Blastocystis* on human health remain controversial (Greige et al.,
388 2018). *Blastocystis* is a common protozoan parasite found in the human digestive tract and is the most
389 common eukaryotic parasite of the human intestine (Greige et al., 2018; Nieves-Ramírez et al., 2018;
390 Souppart et al., 2009). It has been linked to gastrointestinal symptoms, but has also been described as
391 an asymptomatic colonizer of the normal intestinal microbiota (Nieves-Ramírez et al., 2018).

392 FLA are ubiquitous protozoa that can cause opportunistic and non-opportunistic infections in humans
393 (Plutzer and Karanis, 2016; Serrano-Luna et al., 2013). FLA such as *Naegleria*, *Acanthamoeba* and
394 *Vermamoeba* species are ubiquitously distributed worldwide in various warm aquatic environments
395 (Panda et al., 2015). Waterborne transmission of pathogenic strains of *Acanthamoeba* spp. and
396 *Naegleria* spp. is of great relevance (Plutzer and Karanis, 2016). In our study, *N. clarki* was the most
397 frequently detected species in raw sewage.

398 Three *Entamoeba* spp. were identified in our raw sewage samples: *E. moshkovskii*, *E. coli* and *E.*
399 *dispar*. *E. moshkovskii* frequently occurs in regions where amebiasis is highly prevalent and has been
400 isolated from wastewater, rivers, lakes and brackish water (Serrano-Luna et al., 2013). *E. moshkovskii*
401 is primarily considered an FLA, but it can infect humans (Ali et al., 2003; Mahmoudi, 2015). *E.*
402 *dispar* has traditionally been considered non-pathogenic, but this view is being re-evaluated (Caliari et
403 al., 2015). Current evidence suggests that the risk factors for acquiring *E. moshkovskii* infections are
404 similar to those described for *E. histolytica* and *E. dispar* (Serrano-Luna et al., 2013).

405

406 4.2. Diversity of microorganisms in different irrigation water sources

407 4.2.1. *Drinking water*

408 Viruses infecting bacteria and algae were detected in drinking water. Previous studies with potable
409 water have described latent forms of bacteriophages to be the most abundant viruses in this type of
410 water source (Rosario et al., 2009). Although our results confirmed that bacteriophages were
411 detectable after chlorination, there were no human viral pathogens in drinking water.

412 *Pseudomonas* (33.2%) and *Cellvibrio* (27.0%) bacteria were highly abundant in drinking water, as has
413 been previously reported (Bertelli et al., 2018; El-Chakhtoura et al., 2018). Despite the low abundance
414 of this species, it is still higher than that reported in other studies (Bertelli et al., 2018). Among the
415 protozoa, we detected *Acanthamoeba* at low abundances in drinking water. This free-living
416 opportunistic protozoan parasite has been detected in drinking water used for irrigation and has been
417 also reported to be present in bottled water (Khan, 2003). Pathogenic protozoa have been detected in
418 municipal drinking water systems, even in countries with highly regulated water infrastructures,
419 causing waterborne and food-borne outbreaks (Kulinkina et al., 2016). This might be particularly
420 alarming in countries where no residual disinfectant is applied to the water (Braeye et al., 2015). The
421 low abundances of the two pathogenic microorganisms detected in the drinking water samples need to
422 be analyzed further, given that NGS data present important limitations and that the detected
423 microorganisms might be inactive.

424 4.2.2. *Reservoir water and river water*

425 Reservoir water and river water showed very similar distributions of the viral families, sharing 97 of
426 the complete viral genomes. Although the two types of water sources were sampled from different
427 river basins, it is well known that microbial communities are similar in these two ecosystems (Vaz-
428 Moreira et al., 2014). The reduced runoff within the Mediterranean region that might be caused by
429 climate change will affect the circulation patterns of river water into reservoirs and consequently the
430 proportions of nutrients and organic carbon in these ecosystems (Imek et al., 2011). Thus, viral
431 diversity may vary between intense rainfall events and long dry periods, when the demand for
432 irrigation water increases and the limited water circulating into reservoirs increases anoxic conditions.

433 We observed higher viral diversity in the river water samples collected during autumn and winter.
434 Except in the pooled samples collected in the summer, the most abundant *Microviridae* phage family
435 detected was the recently described Eel River Basin Microviridae (ERBM), most frequently found in
436 sediments (Bryson et al., 2015). It has been reported that the microbial community structure and
437 associated metagenomes, especially for bacteriophage families, can change after rainfall events
438 (Tseng et al., 2013). Thus, the seasonality in viral diversity for river water samples may be due to the
439 fact that the river flow decreases during the summer months (from 8 m³/s to 1 m³/s) and
440 bacteriophages are not resuspended from river sediments.

441 Regarding plant viruses, it is not clear how these viruses found in irrigation water could infect the
442 plants being irrigated. Detailed studies are required to elucidate their potential impact on agriculture
443 (Balique et al., 2015; Colson et al., 2010).

444 While pathogenic *Bacillus* species were not detected in reservoir water, 30% of the *Enterobacter*
445 reads corresponded to *Enterobacter cloacae*, which is an opportunistic human pathogen and an
446 indicator of fecal contamination (Chen et al., 2014). Thus, the presence of this bacterial species could
447 indicate the existence of a source of human fecal contamination. Contigs corresponding to common
448 fecal indicators, as crAssphage (1 in autumn) and Pepper mild mottle virus (PMMoV) (23 in winter
449 and 29 in autumn), were detected in riverwater.

450 As stated above, *N. australiensis* and *N. clarki* were detected in reservoir water (0.13%) and river
451 water (0.008%). *Naegleria* isolation has been reported to be higher during the warmer seasons in
452 diverse aquatic habitats worldwide (De Jonckheere, 2004; Panda et al., 2015; Pernin et al., 1998). *N.*
453 *australiensis* can be pathogenic to mice and could affect human health (Kao et al., 2012).

454 It is interesting to note that *Pseudomonas* and *Bacillus* bacteria were predominant in the pooled
455 samples from winter and summer. Similar results of the relative abundances of the *Flavobacteriaceae*
456 and *Comamonadaceae* families were also obtained from the nearby river Ter (Lekunberri et al.,
457 2018).

458 4.2.3. Groundwater

459 It has been reported that viruses can survive for a long time in groundwater and can pass through the
460 aquifer pores that physically filter out larger pathogens (Hunt et al., 2010; Ogorzaly et al., 2010). In
461 our study, waterborne excreted pathogens were not frequently detected in groundwater samples,
462 whereas plant viruses and bacteriophages were widely prevalent. However, we did detect HEV in the
463 pooled sample of groundwater collected in the summer and human adenovirus (HAdV) in the sample
464 collected during autumn. HAdV, which indicates human fecal contamination, has been associated
465 with precipitation events and can persist in the environment for a long time (up to 1,300 days)
466 (Bradbury et al., 2013; Kauppinen et al., 2018). These results were confirmed by running specific RT-
467 qPCR (Jothikumar et al., 2006) and sequencing (data not shown). Interestingly, we detected porcine
468 adenovirus in the groundwater sample collected during the summer (at a concentration of 10E+03
469 GC/L), confirming porcine fecal pollution (data not shown). HEV, a zoonotic pathogen that can cause
470 self-limiting or fulminant hepatitis in humans, was detected in groundwater despite its low prevalence
471 in the studied area. Intensive pig farming occurs in the surrounding areas from where the groundwater
472 was extracted, suggesting that leakage from the slurry tanks could have contaminated the groundwater
473 with the zoonotic pathogen.

474 Pooled samples of groundwater showed different distributions of bacterial populations between the
475 seasons. *Geobacillus* (20.0% in the spring to 0.9% in the summer), followed by *Thermus* and
476 *Anoxybacillus*, all showed seasonal peaks and were frequently distributed in water springs. The high
477 abundance of proteobacteria in groundwater has been also reported previously (Hong et al., 2013).

478 It is interesting to point out the relative abundance of *Vermamoeba vermiformis* (2.63%) in the
479 summer groundwater sample. *V. vermiformis* colonizes water systems and is a reservoir of pathogenic
480 bacteria such as *Legionella pneumophila* (Fouque et al., 2014). The detection of *N. fowleri* in
481 groundwater is of particular concern for human health as it is the principal cause of PAM, a rare and
482 severe disease that causes inflammation and destruction of the brain. The risk of infection depends on
483 the relatively high concentrations of *N. fowleri* in water (Cabanés et al., 2001). The relative
484 abundance of *N. fowleri* in the summer groundwater sample was 0.63%, but we were not able to
485 determine the concentration of the amoeba with our method. *Naegleria* spp. present an additional

486 threat to human health because they can act as vehicles for the multiplication and dispersal of
487 pathogenic bacteria (Huang and Hsu, 2010).

488 4.2.4. Reclaimed water

489 Plant viruses, including *Virgaviridae*, *Tombusviridae* and *Phycodnaviridae*, and bacteriophages, such
490 as *Siphoviridae*, *Myoviridae*, *Podoviridae* and *Microviridae*, accounted for more than 90% of the
491 known viral reads obtained with our reclaimed water samples, representing over 80% of the known
492 viral species detected (Figure 1). The wetland from where the reclaimed water originated appeared to
493 restore the naturally occurring microbial communities, as the distribution of the viral families detected
494 in the reclaimed water samples resembled that of freshwater samples (river and reservoir water). We
495 identified different circoviruses and parvoviruses in reclaimed water, which agrees with the findings
496 of other metagenomic studies performed on raw sewage and reclaimed water produced after the
497 chlorination of treated secondary effluents (Blinkova et al., 2009; Rosario et al., 2009). Viruses
498 belonging to these two families are highly stable in the environment and have a broad host range that
499 includes mammals and invertebrates. Given the huge biodiversity in complex ecosystems such as
500 wetlands, these viruses might not be of concern for human health. CrAssphage (1 in summer) and
501 PMMoV contigs (5, 21 and 23 in summer, winter and autumn respectively), were detected at the
502 wetland effluent samples.

503 According to the minimum quality requirements for water reuse in agricultural irrigation, wetland
504 effluents may be optimal for crop irrigation as class B recycled water, given that it is rich in
505 macronutrients like nitrogen and phosphorus (reported mean values over the year of sampling: 63
506 mg/l for N and 10.7 mg/l for P (data not shown)) and no human pathogens were detected.

507 The *Actinobacteria* and *Limnohabitans* genera were predominant in all the reclaimed water samples.
508 These bacteria inhabit a broad range of freshwater habitats and can constitute up to 30% of the free-
509 living bacteria in freshwater systems (Zwart et al., 2002). As previously stated for viruses, the
510 bacteriome of reclaimed water was similar to that of river water, demonstrating that the treatment
511 used to produce the reclaimed water restores the bacterial populations observed in river water.

512 It should be noted that the prevalence of *Cryptosporidium* spp. and *Giardia intestinalis* might have
513 been underestimated by our metagenomic approach. These parasites are the two most commonly
514 occurring protozoa associated with waterborne disease outbreaks (Plutzer and Karanis, 2016) and are
515 highly prevalent in sewage and reclaimed water. We detected *Cryptosporidium* spp. in the samples
516 collected during summer. Although the SMF method has demonstrated to be useful to recover
517 *Giardia* (Gonzales-Gustavson et al., 2017), this protozoan pathogen was not found in any of the
518 samples analyzed. The high microbial concentrations could have masked the presence of *Giardia*. It is
519 also interesting to note that *Entamoeba* was detected in reclaimed water. Although the pathogenicity
520 of the detected *E. moshkovskii* remains unknown, the presence of this protozoa confirms previous
521 claims that its prevalence might be underestimated (Shimokawa et al., 2012).

522 4.3. Metagenomic approach as a public health tool

523 Our results give a global picture of the microbial communities present in different types of irrigation
524 water sources produced by using a single concentration method and a sequencing platform. This
525 approach has huge potential given that NGS techniques are expected to become more affordable and
526 automatized. However, the use of metagenomics to assess water quality will not replace current water
527 quality tests in the near future. Increasing the sampling locations and the number of samples
528 sequenced will contribute to the monitoring of microbial trends and diversity shifts. Furthermore, the
529 metagenomics approach can be used to detect viable but non-culturable bacteria/protozoa as well as
530 emerging and known viral pathogens. As these new tools are still in development, several
531 improvements are needed to avoid possible biases and uncertainties associated with the pre-
532 amplification step and the loss of information during the filtering and processing of the huge amounts
533 of data. Further improvements in NGS technologies might enable the study of all 16S/18S rRNA gene
534 sequences to provide greater resolution for better taxonomical classification. The earlier we identify
535 the origin of fecal contamination, the better our chances are to find and eliminate the source of
536 microbial pathogens. Thus, it is critical to have a comprehensive understanding of all the viruses,
537 bacteria and protozoa (including potential pathogens) circulating in different irrigation water sources,
538 as this information helps assess water quality.

539 We detected only a few human pathogens in both conventional irrigation water sources (drinking,
540 reservoir, groundwater and river water) and reclaimed water. Microbial abundances and diversities in
541 the reclaimed water samples were very similar to those in the river water and reservoir water samples.
542 In the Mediterranean region, increased temperatures and decreased precipitation will cause a general
543 decrease in water availability. This may result in pathogens occurring at higher concentrations in
544 rivers as the river flow rate will not be sufficient to dilute the effluents from WWTPs (Rusiñol et al.,
545 2015). Seasonal effects have to be carefully evaluated, as one pooled sample per season might result
546 in overestimation of the seasonal effects. As sequencing costs of the high-throughput platforms are
547 being reduced, researchers may develop studies with large number of samples per season, which we
548 consider as one caveat of this study.

549 The Urban Waste Water Treatment Directive (91/271/CEE, 1991) does not require disinfection prior
550 to discharge despite the fact that WWTP effluents contain elevated levels of pathogens (Naughton and
551 Rousselot, 2017; Rusiñol and Girones, 2017). Thus, sustainable and green water treatments to produce
552 reclaimed water are not only a way of producing new irrigation water sources, but also a way of
553 reducing pathogen levels in the receiving water bodies. Water reuse for irrigation purposes is
554 consistent with population growth plans and is a chance to close the human water cycle.

555

556 5. Conclusions

- 557 • Viral diversity in river water varied between seasons, with bacteriophages becoming more
558 prevalent during autumn and winter, when norovirus GII strains also become detectable.
- 559 • Reservoir water contained *Enterobacter cloacae*, an opportunistic human pathogen and indicator
560 of fecal contamination, *Naegleria australiensis* and *Naegleria clarki*.
- 561 • The metagenomic approach revealed the presence of the human pathogens HEV, HAdV and
562 *Naegleria fowleri* in groundwater samples collected during the summer.
- 563 • A constructed wetland used as a sustainable system to treat secondary effluents from an urban
564 WWTP seemed to restore naturally-occurring microbial communities as the virome and
565 bacteriome of the reclaimed water resembled those of freshwater (river and reservoir water).

566

567 **Data availability statement**

568 Raw data can be found in zenodo under the DOI number 10.5281/zenodo.2620427.

569

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571

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577

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579

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