

Document downloaded from:

<http://hdl.handle.net/10251/159974>

This paper must be cited as:

Oliver-Tomás, B.; Hitzl, M.; Owsianiak, M.; Renz, M. (2019). Evaluation of hydrothermal carbonization in urban mining for the recovery of phosphorus from the organic fraction of municipal solid waste. *Resources Conservation and Recycling*. 147:111-118.
<https://doi.org/10.1016/j.resconrec.2019.04.023>



The final publication is available at

<https://doi.org/10.1016/j.resconrec.2019.04.023>

Copyright Elsevier

Additional Information

Evaluation of hydrothermal carbonization in urban mining for the recovery of phosphorus from the organic fraction of municipal solid waste

Borja Oliver-Tomas,^{a,c} Martin Hitzl,^a Mikołaj Owsianiak,^b Michael Renz*^c

^a Ingelia, S.L., C/Jaime Roig 19, 46010 Valencia, Spain

^b Division for Quantitative Sustainability Assessment, Department of Management Engineering, Technical University of Denmark, Bygningstorvet, Building 116B, DK-2800 Kgs. Lyngby, Denmark

^c Instituto de Tecnología Química (UPV-CSIC), Universitat Politècnica de València - Consejo Superior de Investigaciones Científicas, Avenida de los Naranjos s/n, 46022 Valencia, Spain; E-mail: mrenz@itq.upv.es

Abstract

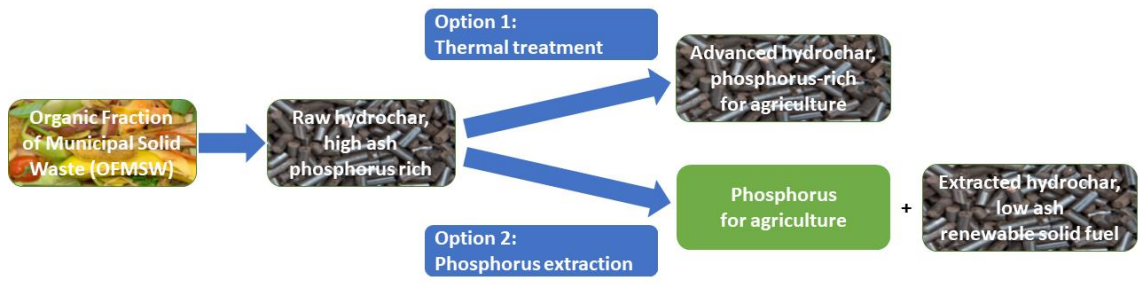
The organic fraction of municipal solid waste was identified as an alternative phosphorus resource: hydrothermal carbonization provided phosphorus-rich hydrochar. Two alternative valorization pathways can be considered for the latter: the use as a fertilizer or as solid fuel after phosphorus extraction. By means of life cycle assessment (LCA) the environmental impact of extracting phosphorus and using the hydrochar as solid fuel was evaluated. Therefore, in a first step, phosphorus extraction with nitric acid, hydrochloric acid and sulfuric acid was experimentally investigated on laboratory scale. Nitric acid proved to be the most suitable because it offered high extraction efficiency and improved solid fuel properties such as lower ash content and lower levels of chlorine and sulfur. In contrast, hydrochloric acid increased the chlorine content and sulfuric acid only replaced phosphate by sulfate, but did not reduce the ash content of hydrochar. Then phosphorus can be precipitated and used as fertilizer.

Although technically feasible, LCA points out that the separate use of hydrochar and phosphorus represents an overall environmental burden for wide range of impact categories, including climate change and resource depletion. Therefore, other applications for phosphorus-rich hydrochars, like agriculture and horticulture, should be considered.

Keywords

Advanced hydrochar, low-ash solid fuel, phosphorus extraction, phosphorus fertilizer,

Graphical Abstract



37

38

39 Introduction

40 Phosphorus is an essential element for all life on earth, and the European Commission (EC) has
41 identified phosphate rock as a critical raw material in 2017 with a nonexistent recycling
42 rate.(European Commission, 2017) Approximately 95% of the world's phosphate production is
43 used in the fertilizer industry.(Desmidt et al., 2015) This figure makes it clear that phosphate
44 consumption is strongly related to nutrition and population growth.

45 Phosphorus is a chemical element whose use does not make it disappear, but spreads it out on
46 the planet. Therefore, the recovery of phosphorus from alternative sources is urgently needed.
47 A promising approach is to perform a fixation step for the phosphorus, a type of
48 homogenization and concentration. For example, sewage sludge is a plant nutrient rich
49 resource(Zhang et al., 2018, 2017) and when sewage sludge is incinerated, phosphorus is
50 concentrated in the ash and the liquid phase is then absent.(Cieřlik and Konieczka, 2017; Melia
51 et al., 2017) The ashes can directly be used as fertilizers but phosphorus extraction with
52 mineral acid and precipitation increases its bioavailability.(Cieřlik and Konieczka, 2017)

53 The present study was initiated when elevated phosphorus concentrations were determined
54 for hydrochar originating from the organic fraction of municipal solid waste (OFMSW). It
55 became evident that hydrothermal carbonization (HTC) can be part of a novel process for
56 phosphorus concentration from this undervalorized resource. The HTC process is an emerging
57 technology and produces hydrochar, a solid carbonaceous material, as main product from wet
58 lignocellulosic biomass such as the OFMSW.(Funke and Ziegler, 2010; Hu et al., 2010; Titirici et
59 al., 2012; Maria M Titirici et al., 2007) The wet conditions during the process make the
60 previous drying superfluous, which is a significant advantage over alternative thermochemical
61 processes (torrefaction, pyrolysis, and gasification), considering the economics for wet
62 feedstocks.(Cha et al., 2016; Kambo and Dutta, 2015)

63 In terms of valorization and commercialization of phosphorous-containing hydrochar, two
64 main pathways can compete with each other: application of phosphorus-containing hydrochar
65 to soil(Lang et al., 2019) or phosphorus extraction for its use in agriculture and the use of the
66 depleted hydrochar as solid fuel.

67 For the use of hydrochar in agriculture, a post-treatment is required as pristine hydrochar may
68 involve phytotoxicity. Therefore, several options are known(Busch et al., 2013, 2012; Dalias et
69 al., 2018; Fornes and Belda, 2017) and a by-product-free treatment is to desorb the undesired
70 compounds and to use them directly for heat generation for the HTC process.(Hitzl et al., 2018)
71 In the latter case, the product is referred to as advanced hydrochar, which is suitable for
72 agriculture, to distinguish it clearly from pristine hydrochar. The use of pristine hydrochar in
73 agriculture, e.g. in Spain, involves climate change benefits,(Owsianiak et al., 2018) and a rough
74 estimation confirmed that this also applies to advanced hydrochars.(Hitzl et al., 2018)

75 When hydrochar is considered a solid fuel and phosphorus should be destined to agriculture,
76 then phosphorus should be extracted at the production site and, thereby, further
77 concentrated. The reason for this is that burning of hydrochar as solid fuel is foreseen in
78 smaller units such as domestic or district heating systems. This leads again to a territorial
79 spread of the phosphorus-rich ash, which would cause again logistic costs for its collection.

80 An additional economic benefit of this approach is the improvement in the fuel quality of
81 hydrochar.(Wang et al., 2018) Phosphorus forms part of the ashes of hydrochar, and its
82 removal reduces the ash content. This is relevant because carbonaceous materials are
83 classified as solid fuels in terms of their ash content.(International Organization for

84 Standardization, n.d.) Improving the product quality of hydrochar as a solid fuel increases its
85 value and contributes positively to the economic balance of the production process.

86 Here we show that the phosphorus extraction from hydrochar is generally possible with
87 mineral acid. By simply neutralizing the extract, phosphorus is recovered and can then be
88 marketed as a solid fertilizer from secondary phosphate sources. Based on these data, we
89 performed the Life Cycle Assessment (LCA) of the developed process, considering the
90 extracted hydrochar as solid fuel and the consequences of its combustion.

91 [Materials and Methods](#)

92 [Hydrothermal carbonization](#)

93 All residual biomass was processed at Ingelia's pilot plant.(Hitzl et al., 2015) The production of
94 hydrochar from the OFMSW has been described elsewhere.(Hitzl et al., 2018) The organic
95 fraction was obtained from households in Guipúzcoa (The Basque Country, Spain) by separate
96 collection which was operated by Gipuzkoako Hondakinen Kudeaketa SAU. It contained
97 biodegradable plastic bags, glass and plastic bottles as "impurities" in low amount and a water
98 content of 75%. For the trial, 9.24 t of this resource were transformed into 1.05 t of dry
99 hydrochar with an ash content of 12.9% (determined by calcination at 815 °C).(Hitzl et al.,
100 2018) Details on the processing of green waste can be found in literature.(Hitzl et al., 2015) as
101 well as for the transformation of orange peel waste (OPW).(Burguete et al., 2016) Hydrochar
102 from sewage sludge was also produced at Ingelia's pilot plant during the FP7 NEWAPP
103 project.("New technological applications for wet biomass waste stream products," n.d.)

104 [Phosphorus extraction](#)

105 In a round-bottom flask, 4.00 g of hydrochar was placed together with 15 mL of acid solution.
106 The latter was prepared from HTC process water obtained from Ingelia, adding the amount of
107 the corresponding concentrated acid to reach the desired concentration. The final
108 concentration of sulfuric acid was 5 wt%, of nitric acid 6 wt%, and of hydrochloric acid 4 – 5
109 wt%. During the acidification of the process water, a solid precipitated which was removed by
110 filtration and discarded. The hydrochar suspension was stirred for four hours at different
111 temperatures, namely at 40, 60 , 80, and 100 °C (see also , Figure S2, ESI). The solid was
112 collected by filtration, washed with 100 mL of deionized water and dried overnight at 100 °C.
113 Phosphorus content in hydrochar was calculated from ash content and ash composition, the
114 latter determined by inductively coupled plasma optical emission spectrometry (ICP-OES) on a
115 Varian 715 ES spectrometer. Phosphorus recovered was calculated from the amounts and
116 concentrations of the acid extract and the washing water. Their phosphorus content was
117 determined directly by ICP-OES. Higher heating values (HHV) were determined following the
118 analytical standard UNE-EN 14918:2011.

119 [Phosphorus precipitation](#)

120 For phosphorus precipitation the pH of the acid extract was adjusted to 8 to 9.5 adding calcium
121 hydroxide. The solid was collected by filtration and dried in an oven at 100 °C overnight.

122 [LCA modelling part](#)

123 Environmental performance of phosphorus removal was assessed using environmental LCA. It
124 was conducted in accordance with the requirements of the ISO 14044 standard and the
125 guidelines of the EU Commission's International Reference Life Cycle Data System (ILCD)
126 Handbook.(*ILCD Handbook: General guide for Life Cycle Assessment - Detailed guidance*, 2010,

127 *ISO 14044:2006 Environmental management -- Life cycle assessment -- Requirements and*
128 *guidelines, 2006)*

129 *Goal and scope*

130 Because hydrothermal carbonization is a waste treatment option, the functional unit was
131 defined as “hydrothermal carbonization of 1 kg of wet biowaste with 100% content of biogenic
132 carbon of total carbon”. This definition allows for a fair comparison between biowaste with
133 varying phosphorus content and with hydrothermal carbonization of biowaste without
134 phosphorus recovery. The most likely use of hydrochar is as solid fuel for domestic heating.
135 Nitrogen and sulfur content of the hydrochar were taken into account for emissions, whereas
136 a complete elimination of the mineral acids during the washing process was assumed.

137 The current study is considered a micro-level decision support (type A) situation according to
138 ILCD guidelines, and the assessment applies an attributional approach where average Spanish
139 data are used. System boundaries specifying the processes included have been described
140 before.(Owsianiak et al., 2016) We included replaced the conventional waste management
141 system (composting), replaced combustion of fossil coal, construction of a HTC plant,
142 production of the hydrochar and HTC process water and transportation of the hydrochar. In
143 addition, production of acids and base for recovery of P was included. System expansion was
144 performed for processes with recovery of commodities; recycled steel substitutes the
145 production of virgin steel and the HTC process water (concentrated at the HTC plant using
146 reverse osmosis) substitutes the production of inorganic fertilizers. Likewise, impact offsets
147 (also known as credits) are given to recovered phosphorus (replacing phosphorus from
148 phosphate rock), to avoided conventional treatment of biowaste, and to energy recovery when
149 hydrochar is used for domestic heating (replacing combustion of fossil coal), in accordance
150 with the recommendations of the ILCD guidelines for this decision support type.

151 Product systems were modelled in SimaPro, version 8.3.0.0 (PRe Consultants bv, the
152 Netherlands). Parameters and data underlying the modeling of HTC plant are documented in
153 literature.(Owsianiak et al., 2016) Data for foreground processes in the HTC system, are based
154 on primary data measured at a HTC plant at Ingelia S.L. (Valencia, Spain). Data for generic
155 processes, such as electricity production and waste management processes, are based on
156 those available in the ecoinvent database, version 3.2.(Weidema et al., 2013)

157 ILCD’s method was used in the life cycle impact assessment phase (ILCD 2011 Midpoint+,
158 version 1.05),(Hauschild et al., 2013) as implemented in the LCA modeling software SimaPro,
159 version 8.3.0.0 (PRe Consultants bv, the Netherlands). All ILCD impact categories were used,
160 except ionizing radiation impacts on ecosystems which is considered not sufficiently
161 representative for this type of impact.

162 *Scenarios*

163 In total, twelve scenarios were considered (Table 1). Overall, three different acids (HNO₃,
164 H₂SO₄ and HCl) in two concentrations applied to phosphorus extraction from hydrochar made
165 from the OFMSW and a hypothetical one with the double phosphorus amount. In detail, for
166 the first three scenarios (1–3) the same acid (and base) concentrations and amounts were
167 used as in the experiments. Initial phosphorus content in the hydrochar for these scenarios
168 was the content measured for hydrochar made from the OFMSW (1.25%; 22.1% of the ashes),
169 which is in a higher range than values measured for hydrochars made from other biowaste
170 types. Assuming that amounts of acids and bases can be optimized during up-scaling, further
171 three scenarios were considered with amounts of acids and base reduced 10 times (scenarios

172 4–6). In addition, as starting material for the extractions, in six scenarios hydrochar was
 173 considered to involve the double amount of phosphorus, compared to the current one.
 174 Hereby, extraction efficiency was maintained. In the first group, acid amount was maintained
 175 as in the experimental procedures (scenarios 7–9) and in the second group (scenarios 10–12)
 176 reduced acid amounts and concentrations were used. Sensitivity of the results to the
 177 parameters presented in Table 1 was done by simply comparing impact scores between
 178 scenarios.

179

Scenario	Sensitivity parameter	Acid type	Acid concentration	P content
1	Baseline	HNO ₃	High (6%)	High (1.25%)
2, 3	Acid type	HCl; H ₂ SO ₄	High (4%; 6%)	High (1.25%)
4–6	Acid concentration	HNO ₃ ; HCl; H ₂ SO ₄	Optimized (1/10 of initial)	High (1.25%)
7–9	P content	HNO ₃ ; HCl; H ₂ SO ₄	High (4%; 6%)	Very high (2.50%)
10–12	All potential improvements	HNO ₃ ; HCl; H ₂ SO ₄	Optimized (1/10 of initial)	Very high (2.50%)

180

181 *Table 1. Overview of scenarios used in the scenario analysis.*

182

183 Results and discussion

184 Phosphorus extraction

185 As previously reported, the organic fraction of municipal solid waste has been converted into
186 hydrochar on a ton scale.(Hitzl et al., 2018) In a nutshell, 9.24 t of wet starting material gave
187 1.05 t of dry hydrochar, which corresponded to a mass yield of 45% (based on dry matter). The
188 total ash content of the hydrochar was 13% and the concentrations of its major components
189 are shown in Figure S1. From these data a total phosphorus content in the hydrochar of 1.25%
190 (as P) was calculated. This high concentration was relatively surprising as the phosphorus
191 content was generally neglected in the analysis of hydrochar. When it was reported, it was
192 significantly lower at 0.42% for food waste.(Idowu et al., 2017)

193 The second HTC product, the process water, generally contains a low phosphorus
194 concentration of less than 5 ppm. Therefore, it can be concluded that phosphorus is fixed in
195 hydrochar, i.e. all phosphorus present in the wet raw material is recovered in dry hydrochar
196 and, its concentration is increased by a factor of nearly nine in the present case.

197 Nevertheless, the phosphorus concentration in the solid hydrochar was still an order of
198 magnitude lower than in the phosphate rock, in which its concentration is typically between 13
199 and 17%. However, hydrochar offers advantages over the rock for the extraction process. In
200 particular, hydrochar already has a small particle size (<5 cm) and a high grindability index (the
201 particle size is further reduced easily by milling).(Kempegowda et al., 2017; Smith et al., 2018)
202 In addition, the chemical transformation during the HTC, that is, the dehydration, breaks up
203 the plant structure(Hitzl et al., 2015) and the hard biomass is “perforated by a spongy,
204 continuous system of nanopores”.(M M Titirici et al., 2007) Therefore, it should be easy for an
205 extracting liquid to completely penetrate the carbonaceous material and access all phosphorus
206 domains.

207 With the aim to proof this hypothesis, the hydrochar derived from OFMSW was extracted with
208 three different mineral acids, sulfuric acid, hydrochloric acid and nitric acid. In fact, Figure S2
209 shows that all three acids were able to remove the phosphorus almost quantitatively with an
210 efficiency of more than 96%. In addition, the extraction was independent of the reaction
211 temperature: In the temperature range from 40 ° C to 100 ° C, no significant change in the
212 efficiency was observed. This is relevant because in the industrial HTC process a hot hydrochar
213 water slurry exits the reactor and cools down during the downstream processing. The results
214 of the extraction indicated that the temperature of the carbonaceous material could be
215 neglected and the extraction could be carried out after excess process water had been
216 removed in the filter press (to a humidity of approximately 50%).

217 Phosphorus extraction worked well with all three mineral acids, but the effect on ash content
218 was not the same in the three cases. For sulfuric acid, the ash content did not vary
219 significantly, whereas for nitric acid and hydrochloric acid, the ash content after extraction
220 decreased from 13% to approximately 4% (Figure 1a). Analysis of the ash components showed
221 that calcium and phosphorus are significantly reduced in the latter cases, whereas with sulfuric
222 acid the calcium concentration was still similar to the initial one (Figure 1b). In view of the low
223 solubility of calcium sulfate, this was a predictable observation: calcium phosphate in the
224 hydrochar is replaced by calcium sulfate. The sulfate content is included in the “other” fraction
225 in Figure 1b. From these results it was concluded that sulfuric acid was suitable for phosphorus
226 extraction. However, with high calcium hydrochars, the ash content did not decrease when

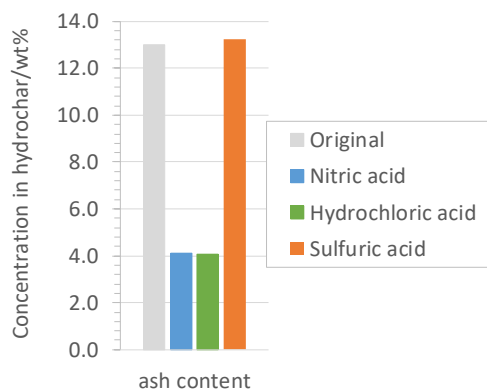
227 phosphorus was removed. If a lower ash content was desired, sulfuric acid was not a valuable
228 choice and hydrochloric acid or nitric acid should be used.

229 When mineral acids were used for phosphorus extraction from sewage sludge ashes, sulfuric
230 acid was the reagent of choice for economic reasons, although it increased the amounts of
231 inorganic residues.(Ottosen et al., 2013) In contrast, in the present case the organic nature of
232 the by-product was a game changer: sulfate deposits deteriorated the product properties.

233 When using nitric acid and hydrochloric acid ash contents were clearly reduced (Figure 1a),
234 and in particular calcium and phosphorus concentration in the acids (Figure 1b). Therefore, it
235 can be concluded that these two mineral acids were very efficient in dissolving of calcium
236 phosphate from hydrochar.

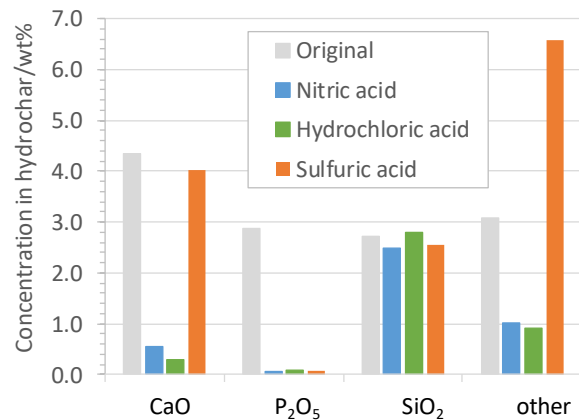
237

238 a)



239

b)



240 *Figure 1. Ash content (a) and concentration of different metal oxides (b) in hydrochar, in the original sample and*
241 *after extractions with different mineral acids.*

242

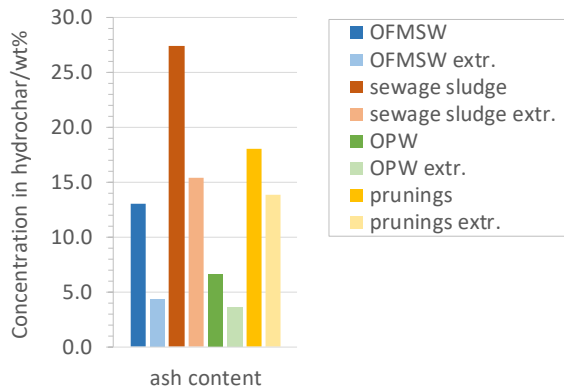
243 Phosphorus extraction of different hydrochar samples

244 Phosphorus extraction and ash reduction were successful with hydrochar from the OFMSW.
245 With the aim to evaluate the scope of the process, it was applied to several hydrochar samples
246 from the FP-7 NEWAPP project. In the latter, organic residual feedstocks were processed by
247 HTC on industrial scale, namely sewage sludge, orange peel waste (OPW) and green and
248 garden prunings. ("Industrial Scale Hydrothermal Carbonization: new applications for wet
249 biomass waste," 2016) The hydrochar samples had a different ash content, from high ash (27%
250 sewage sludge and 18% garden-pruning) to lower ash (OPW 7%, Figure 2a). In addition, ash
251 compositions were significantly different since hydrochar originating from sewage sludge
252 contained the expected increased phosphorus content, while for the other two samples the
253 concentration of this element was rather low (Figure 2b).

254

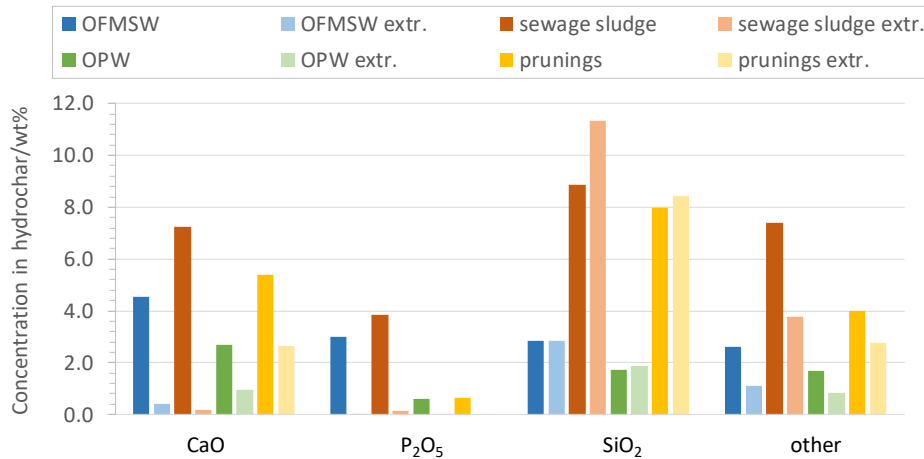
255

256 a)



257

258 b)



259

260 *Figure 2. Ash content (a) and concentration of different metal oxides (b) in hydrochar, measured for the original*
261 *sample and after extractions with hydrochloric acid (4%, 60 °C, 4 h). Hydrochars were prepared from the organic*
262 *fraction of municipal solid waste (OFMSW), sewage sludge, orange peel waste (OPW), and garden prunings.*

263

264 The hydrochloric acid extraction process (4%, 60 °C, 4 h) was able to reduce the phosphorus
265 content to almost zero in all cases (Figure 2b). When the phosphorus content was high (for
266 example, in the sample obtained from sewage sludge), the ash content was also significantly
267 reduced from 27% to 15%. At low phosphorus content (OPW, garden prunings), the gain in ash
268 reduction was rather low (Figure 2a). It was remarkable that calcium was almost completely
269 removed in the sewage sludge-derived hydrochar (initially 7%), as well as in the OFMSW
270 derived hydrochar, but not in the garden-pruning derived one (originally 5%). In the latter
271 case, the calcium content was reduced by only half. Therefore, these data showed that calcium
272 reduction was more efficient when phosphate was present at the same time. Therefore, it was
273 concluded that the present extraction method generally worked well for the phosphorus
274 extraction and was effective for reducing the ashes when both elements, calcium and
275 phosphorus, were present in significant amounts.

276 Performance of the extracted hydrochar as solid fuel

277 It has already been shown that phosphorus extraction can be carried out with the aim to
278 achieve two goals simultaneously: the recovery of phosphorus and the ash reduction. The
279 latter purpose is interesting when hydrochar is commercialized as solid fuel. Therefore,
280 relevant properties of the material after extraction were analyzed and evaluated. Only samples
281 produced with nitric acid and hydrochloric acid were considered since sulfuric acid was
282 previously discarded for improving fuel quality.

283 The first and most important feature of a solid fuel is its energy content, quantified as higher
284 heating value (HHV). Table 2 shows that the HHV for various hydrochar samples increased by
285 up to ten percent through the extraction process. This was due to two different facts: First, the
286 HHV calculated on dry and ash-free base was not significantly affected by the extraction
287 process, and, secondly, the ash content decreased. Since the latter did not contribute to total
288 HHV (on a dry basis), the intrinsic energy was less diluted in the solid when it was removed.

289 The removal of the ash also affects the mass yield, which was reduced (Table 2). This was the
290 penalty for the higher energy concentration in the extracted hydrochar. Therefore, the energy
291 concentration compensates the mass loss and the balance for the amount of energy involved
292 in a given amount of material was the same before and after the extraction process.

293 For environmental purposes, chlorine and sulfur are worrying elements in fuels. For the
294 present treatments, extraction with hydrochloric acid increased the chlorine content by more
295 than a factor of two (from 0.085% to 0.261%). This indicated that washing the hydrochar
296 sample (approx. 4 g) with 100 mL of water, after treatment with 15 mL of dilute acid was not
297 sufficient. However, higher water consumption was discarded as this would further worsen the
298 economic and environmental balance. Clearly, the better option was to change the extracting
299 agent to nitric acid. In this case, the chlorine content was halved (from 0.085% to 0.043%).

300 A positive effect was also detected for the sulfur content. The latter decreased for both cases,
301 hydrochloric acid and nitric acid treatment, by 20 and 65%, respectively. However, the initial
302 value was relatively high at 0.494%, which was reduced to 0.169%. After treatment, hydrochar
303 complied with the requirements of international standard ISO 17225-8: 2016. (International
304 Organization for Standardization, n.d.) For the TA2 specification, the sulfur content must be
305 less than 0.2%. With respect to the chlorine content, both the original and the nitric acid-
306 treated sample satisfied the requirement of 0.1% or less.

307 In summary, it can be stated that the phosphorus extraction with nitric acid improved the
308 hydrochar characteristics as a solid fuel. Not only was the ash content reduced below 5%, but
309 also element concentrations critical for combustion such as chlorine and sulfur were reduced.
310 By this, the requirements of ISO 17225-8 are fulfilled.

311 Recovery of phosphorus as a solid

312 In general, phosphorus precipitation is undergraduate chemistry and does not require much
313 attention. In the present case, however, HTC process water was employed for the preparation
314 of the extraction liquids. Since dissolved material can affect both the precipitation process and
315 the solids properties, these issues were also studied in detail, and related information is
316 included in the Electronic Supplementary Information (ESI).

317 X-ray powder diffraction (Figure S3) and FTIR spectroscopy (Figure S5) showed the formation
318 of calcium hydrogen phosphate (brushite) by neutralization of the extraction liquid. From this
319 solid, phosphorus is available to plants as it is dissolved by neutral water (Figure S8), while

320 calcination (or combustion) converted it into less soluble species. In summary, the results
321 confirmed that the OFMSW can provide a solid phosphorus fertilizer together with a high
322 quality solid fuel.

323 [Environmental performance – Impact scores for laboratory-scale phosphorus extraction](#)
324 The LCA was carried out considering twelve scenarios, starting from experimental conditions
325 and real phosphorus content and adding potential optimization events to the study (for more
326 details, see the Materials and Methods section).

327 Table 3 shows the environmental impact scores modelled for the baseline scenario (scenario 1
328 in Table 3; nitric acid) and two other scenarios with different acids (scenarios 2 and 3 in Table
329 3; hydrochloric acid and sulfuric acid) expressed in category-specific units. Overall, three major
330 observations are seen. First, impact scores are positive in most impact categories, suggesting
331 that the treatment of biowaste with phosphorus recovery through hydrothermal carbonization
332 is expected to result in environmental burden. This environmental burden is mainly due to the
333 need to produce relatively large amounts of acids and bases for P recovery. Secondly, negative
334 impact scores are seen for three impact categories, namely particulate matter formation,
335 acidification (in scenarios 2 and 3) and water resource depletion. This suggests that
336 environmental benefits are expected for these impact categories when biowaste is treated
337 hydrothermally with recovery of phosphorus. Our analysis shows that these benefits are
338 mainly due to the avoidance of fossil coal heat generation when using hydrochar as a solid fuel
339 (in terms of particulate matter and acidification impacts). Negative scores for water resource
340 depletion, however, are due to the avoided use of water in the upstream processes for the
341 production of base (potassium hydroxide as example of a base from the ecoinvent database).
342 The third finding is that phosphorus extraction with H₂SO₄ is environmentally friendly in twelve
343 out of 15 impact categories (excluding particulate matter, water resource depletion and, in
344 some cases, acidification) compared to extraction with HNO₃ or HCl. This is due to lower
345 amounts of sulfuric acid (on a mass basis) needed for extraction per treated waste unit.

346

347

348

entry	starting material	initial HHV ^a [MJ/kg]	final HHV ^b [MJ/kg]	HHV increase ^c [%DRY]	HHV change ^d [%DAF]	final ash content ^e [%]	mass yield ^f [%]	energy balance ^g [%]
1	OFMSW	27.9	30.6	9.6	-0.3	34	86	94.2
2	OPW	24.1	25.4	5.5	2.2	55	92	97.1
3	prunings	20.4	22.3	9.5	4.2	77	90	98.6

349 ^a HHV of hydrochar before extraction on dry base. ^b HHV of hydrochar after extraction on dry base. ^c Change in HHV caused
350 by the extraction procedure on dry base. ^d Change in HHV caused by the extraction procedure on dry and ash-free (DAF)
351 base. ^e Final ash content compared to the initial one. ^f Mass yield for hydrochar after the extraction procedure. ^g Energy
352 balance considering hydrochar as an energy carrier and comparing the final energy amount (mass x HHV) with the initial
353 one (Formula: energy balance = (final HHV x mass yield)/initial HHV).

354

355 *Table 2. Solid fuel properties of different hydrochars before and after applying an extraction procedure with HCl (4%, 60 °C, 4 h).*

356

357

358

359

Impact category	Unit	Impact score		
		HNO ₃	HCl	H ₂ SO ₄
Climate change (GWP)	kg CO ₂ eq	5.19E-01	2.91E-01	1.50E-01
Ozone depletion	kg CFC-11 eq	4.82E-08	8.85E-08	3.94E-08
Human toxicity, cancer effects	CTUh	1.88E-08	2.44E-08	1.61E-08
Human toxicity, non-cancer effects	CTUh	7.70E-08	1.30E-07	6.68E-08
Particulate matter	kg PM _{2.5} eq	-9.03E-05	-1.45E-04	-1.09E-04
Ionizing radiation HH	kBq U235 eq	9.33E-02	1.27E-01	8.67E-02
Photochemical ozone formation	kg NMVOC eq	1.70E-03	1.35E-03	1.09E-03
Acidification	molc H ⁺ eq	8.00E-04	-6.49E-05	-4.37E-04
Terrestrial eutrophication	molc N eq	6.97E-03	2.08E-03	1.14E-03
Freshwater eutrophication	kg P eq	8.56E-05	1.50E-04	5.88E-05
Marine eutrophication	kg N eq	7.61E-04	6.17E-04	5.14E-04
Freshwater ecotoxicity	CTUe	4.41E+00	5.48E+00	3.92E+00
Land use	kg C deficit	6.69E-01	6.46E-01	5.31E-01
Water resource depletion	m ³ water eq	-7.1E-04	-5.8E-04	-7.0E-04
Mineral, fossil and renewable resource depletion	kg Sb eq	3.01E-05	4.17E-05	4.47E-05

360

361 *Table 3. Impact scores in category-specific units for acid-based P extraction using HNO₃, HCl or H₂SO₄ (scenarios 1, 2*
 362 *and 3 in Table 1, respectively).*

363

364 [Can phosphorus extraction bring environmental benefits?](#)

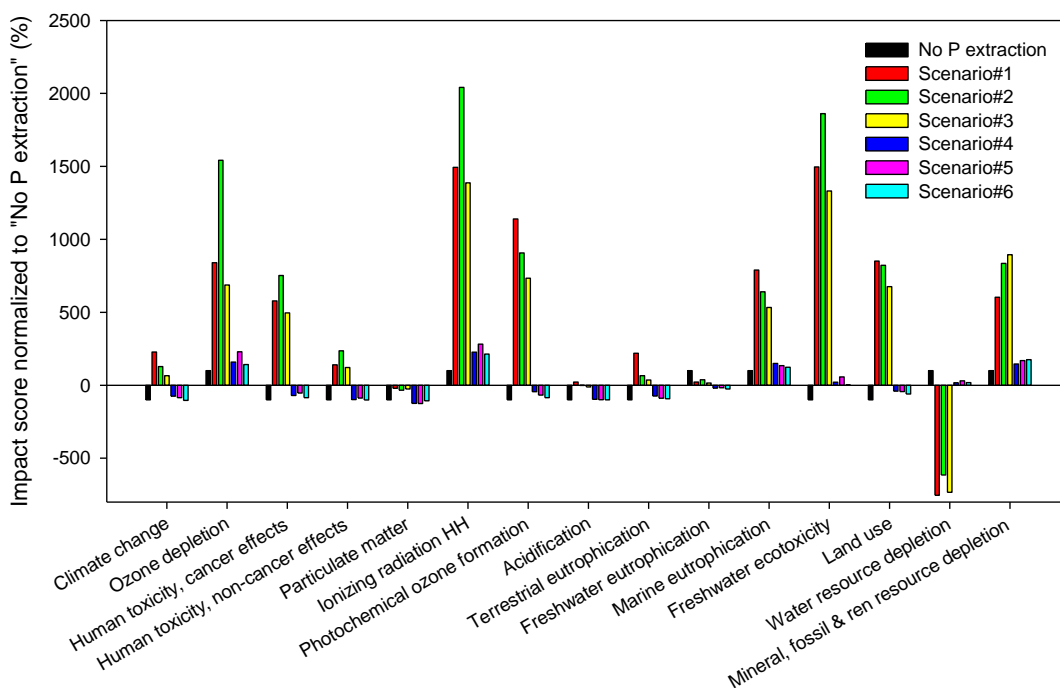
365 Using the experimental conditions from the laboratory scale extractions, only limited
 366 environmental benefits were noted. It should be further examined if benefits were reinforced
 367 in optimized extraction conditions (lower acid level) or for more appropriate cases (higher
 368 phosphorus content in hydrochar).

369 A relatively large contribution of acid and base used for extraction for all impact scores
 370 indicates that there is significant potential for environmental optimization (eco-design):
 371 reduction of acid and base levels during upscaling. In fact, the results presented in Figure 3

372 show that if the amounts of acids and bases are reduced by a factor of 10, which is not
 373 unrealistic when scaling from the bench to the pilot scale, the impact scores are reduced. The
 374 largest decrease was observed for the impact category freshwater ecotoxicity (where impact
 375 scores decrease by a factor of 100). For other impact categories, a reduction by a factor of up
 376 to 10 applies. In some cases impact scores become negative when amount of acid and base is
 377 reduced. This suggests that the environmental benefits of avoiding the burning of fossil coal in
 378 combination with the environmental benefits of avoiding the composting of biowaste
 379 outweigh the burden of hydrothermal treatment and phosphorus recovery. This is the case for
 380 a total of 9 out of 15 impact categories, including climate change. In 11 out of 15 impact
 381 categories, impact scores are comparable to the system without phosphorus extraction. They
 382 are higher for two impact categories (ionizing radiation and freshwater ecotoxicity) and lower
 383 for two impact categories (freshwater eutrophication and water resource depletion) than for
 384 systems without phosphorus extraction.

385 A comparison with systems in which hydrochar contained larger amounts of phosphorus
 386 (scenarios 7–9 in Table 3) showed that impact scores were comparable to scenarios 1–3,
 387 suggesting that the benefits of obtaining relatively large quantities of phosphorus were not
 388 sufficient to outweigh the environmental burden (Figure S9 and Figure S10). Even if the
 389 amounts of acids and bases were optimized, the environmental benefits of P extraction were
 390 relatively low compared to the system without phosphorus extraction.

391



392

393 *Figure 3. Impact scores for scenarios 1–6 in Table 3; expressed in % and normalized internally to the “No phosphorus*
 394 *extraction” scenario, for which impacts are equal either 100% (for positive impact scores) or –100% (for negative*
 395 *impact scores).*

396

397 Conclusions

398 The OFMSW has been proposed as an alternative phosphorus reservoir for phosphorus
399 recovery and solid fuel production. It has been shown that phosphorus can be extracted from
400 hydrochar with mineral acid with high efficiency. At high phosphorus content, the ash content
401 decreased accordingly when nitric acid or hydrochloric acid was used for the extraction. In
402 terms of fuel properties, nitric acid was the preferred mineral acid for extraction as the
403 chlorine and sulfur content of the hydrochar decreased. This further improved the fuel
404 properties of hydrochar and the latter even met the standard requirements for good quality
405 solid fuels.

406 When phosphate was precipitated from the extraction solution, a solid phosphor material was
407 obtained. The latter released phosphorus again in neutral water, so it was predicted that it was
408 readily available to the plants and the solid is suitable as a phosphorus fertilizer. It was
409 concluded that it was technically feasible to produce a solid fuel and a phosphorus fertilizer
410 from phosphorus-rich OFMSW. However, the environmental benefits were limited, as the LCA
411 of this concept showed. Phosphorus extraction was a burden in almost all fifteen-impact
412 categories, including climate change and resource depletion. It is therefore recommended to
413 study the application of phosphorus-containing advanced hydrochar in agriculture and
414 horticulture, without intermediate concentration by extraction. It is predicted that carbon
415 storage benefits are combined with fertilizing effects. (Hitzl et al., 2018; Owsianiak et al., 2018)

416 Acknowledgement

417 The authors are grateful for the financial support received from the Spanish Ministry of
418 Economy and Competiveness under the RTC-2015-4017-3 of the state programme "Research,
419 Development and Innovation Oriented to the Challenges of Society"

420 References

- 421 Boudia, S., Zuddas, P., Fernane, F., Fiallo, M., Sharrock, P., 2018. Mineralogical transformation
422 during hydroxyapatite dissolution in simple aqueous solutions. *Chem. Geol.* 477, 85–91.
423 <https://doi.org/https://doi.org/10.1016/j.chemgeo.2017.12.007>
- 424 Burguete, P., Corma, A., Hitzl, M., Modrego, R., Ponce, E., Renz, M., 2016. Fuel and chemicals
425 from wet lignocellulosic biomass waste streams by hydrothermal carbonization. *Green*
426 *Chem.* 18, 1051–1060. <https://doi.org/10.1039/c5gc02296g>
- 427 Busch, D., Kammann, C., Grünhage, L., Müller, C., 2012. Simple biotoxicity tests for evaluation
428 of carbonaceous soil additives: Establishment and reproducibility of four test procedures.
429 *J. Environ. Qual.* 41, 1023–1032. <https://doi.org/10.2134/jeq2011.0122>
- 430 Busch, D., Stark, A., Kammann, C.I., Glaser, B., 2013. Genotoxic and phytotoxic risk assessment
431 of fresh and treated hydrochar from hydrothermal carbonization compared to biochar
432 from pyrolysis. *Ecotoxicol. Environ. Saf.* 97, 59–66.
433 <https://doi.org/10.1016/j.ecoenv.2013.07.003>
- 434 Cha, J.S., Park, S.H., Jung, S.-C., Ryu, C., Jeon, J.-K., Shin, M.-C., Park, Y.-K., 2016. Production and
435 utilization of biochar: A review. *J. Ind. Eng. Chem.* 40, 1–15.
436 <https://doi.org/10.1016/j.jiec.2016.06.002>
- 437 Cieřlik, B., Konieczka, P., 2017. A review of phosphorus recovery methods at various steps of
438 wastewater treatment and sewage sludge management. The concept of “no solid waste
439 generation” and analytical methods. *J. Clean. Prod.* 142, 1728–1740.
440 <https://doi.org/10.1016/j.jclepro.2016.11.116>

441 Dalias, P., Prasad, M., Mumme, J., Kern, J., Stylianou, M., Christou, A., 2018. Low-cost post-
442 treatments improve the efficacy of hydrochar as peat replacement in growing media. J.
443 Environ. Chem. Eng. 6, 6647–6652. <https://doi.org/10.1016/j.jece.2018.10.042>

444 Desmidt, E., Ghyselbrecht, K., Zhang, Y., Pinoy, L., Van der Bruggen, B., Verstraete, W., Rabaey,
445 K., Meesschaert, B., 2015. Global Phosphorus Scarcity and Full-Scale P-Recovery
446 Techniques: A Review. Crit. Rev. Environ. Sci. Technol. 45, 336–384.
447 <https://doi.org/10.1080/10643389.2013.866531>

448 Ettarh, C., Galwey, A.K., 1996. A kinetic and mechanistic study of the thermal decomposition of
449 calcium nitrate. Thermochim. Acta 288, 203–219.
450 [https://doi.org/https://doi.org/10.1016/S0040-6031\(96\)03052-3](https://doi.org/https://doi.org/10.1016/S0040-6031(96)03052-3)

451 European Commission, 2017. Communication from the Commission to the European
452 Parliament, the Council, the European Economic and Social Committee and the
453 Committee of the Regions on the 2017 list of Critical Raw Materials for the EU [WWW
454 Document]. URL [http://eur-lex.europa.eu/legal-](http://eur-lex.europa.eu/legal-content/EN/ALL/?uri=COM:2017:0490:FIN)
455 [content/EN/ALL/?uri=COM:2017:0490:FIN](http://eur-lex.europa.eu/legal-content/EN/ALL/?uri=COM:2017:0490:FIN)

456 Fornes, F., Belda, R.M., 2017. Acidification with nitric acid improves chemical characteristics
457 and reduces phytotoxicity of alkaline chars. J. Environ. Manage. 191, 237–243.
458 <https://doi.org/http://dx.doi.org/10.1016/j.jenvman.2017.01.026>

459 Funke, A., Ziegler, F., 2010. Hydrothermal carbonization of biomass: A summary and discussion
460 of chemical mechanisms for process engineering. Biofuels, Bioprod. Biorefining 4, 160–
461 177. <https://doi.org/10.1002/bbb.198>

462 Hauschild, M.Z., Goedkoop, M., Guinée, J., Heijungs, R., Huijbregts, M., Jolliet, O., Margni, M.,
463 De Schryver, A., Humbert, S., Laurent, A., Sala, S., Pant, R., 2013. Identifying best existing
464 practice for characterization modeling in life cycle impact assessment. Int. J. Life Cycle
465 Assess. 18, 683–697. <https://doi.org/10.1007/s11367-012-0489-5>

466 Hitzl, M., Corma, A., Pomares, F., Renz, M., 2015. The hydrothermal carbonization (HTC) plant
467 as a decentral biorefinery for wet biomass. Catal. Today 257, 154–159.
468 <https://doi.org/10.1016/j.cattod.2014.09.024>

469 Hitzl, M., Mendez, A., Owsianiak, M., Renz, M., 2018. Making hydrochar suitable for
470 agricultural soil: A thermal treatment to remove organic phytotoxic compounds. J.
471 Environ. Chem. Eng. 6, 7029–7034.
472 <https://doi.org/https://doi.org/10.1016/j.jece.2018.10.064>

473 Hu, B., Wang, K., Wu, L., Yu, S.H., Antonietti, M., Titirici, M.M., 2010. Engineering carbon
474 materials from the hydrothermal carbonization process of biomass. Adv. Mater. 22, 813–
475 828. <https://doi.org/10.1002/adma.200902812>

476 Idowu, I., Li, L., Flora, J.R. V, Pellechia, P.J., Darko, S.A., Ro, K.S., Berge, N.D., 2017.
477 Hydrothermal carbonization of food waste for nutrient recovery and reuse. Waste
478 Manag. 69, 480–491. <https://doi.org/10.1016/j.wasman.2017.08.051>

479 ILCD Handbook: General guide for Life Cycle Assessment - Detailed guidance, 2010. . European
480 Commission.

481 Industrial Scale Hydrothermal Carbonization: new applications for wet biomass waste [WWW
482 Document], 2016. URL [http://www.newapp-project.eu/en/public-library/send/2-public-](http://www.newapp-project.eu/en/public-library/send/2-public-library/3-industrial-scale-hydrothermal-carbonization-new-applications-for-wet-biomass-waste.html)
483 [library/3-industrial-scale-hydrothermal-carbonization-new-applications-for-wet-biomass-](http://www.newapp-project.eu/en/public-library/send/2-public-library/3-industrial-scale-hydrothermal-carbonization-new-applications-for-wet-biomass-waste.html)
484 [waste.html](http://www.newapp-project.eu/en/public-library/send/2-public-library/3-industrial-scale-hydrothermal-carbonization-new-applications-for-wet-biomass-waste.html) (accessed 4.26.18).

- 485 International Organization for Standardization, n.d. Solid biofuels - Fuel specifications and
486 classes - Part 8: Graded thermally treated and densified biomass fuels (ISO/TS 17225-
487 8:2016) [WWW Document]. 2016. URL <https://www.iso.org/standard/71915.html>
488 (accessed 4.27.18).
- 489 ISO 14044:2006 Environmental management -- Life cycle assessment -- Requirements and
490 guidelines, 2006. . International Standards Organization.
- 491 Kambo, H.S., Dutta, A., 2015. A comparative review of biochar and hydrochar in terms of
492 production, physico-chemical properties and applications. *Renew. Sustain. Energy Rev.*
493 45, 359–378. <https://doi.org/10.1016/j.rser.2015.01.050>
- 494 Kempegowda, R.S., Tran, K.-Q., Skreiberg, Ø., 2017. Techno-economic assessment of
495 integrated hydrochar and high-grade activated carbon production for electricity
496 generation and storage, in: N., F., A., R., V., S. (Eds.), . Elsevier Ltd, SINTEF Energy
497 Research, P.O. Box 4761 Sluppen, Trondheim, Norway, pp. 341–348.
498 <https://doi.org/10.1016/j.egypro.2017.07.223>
- 499 Lafon, J.P., Champion, E., Bernache-Assollant, D., 2008. Processing of AB-type carbonated
500 hydroxyapatite $\text{Ca}_{10-x}(\text{PO}_4)_6-x(\text{CO}_3)x(\text{OH})_{2-x-2y}(\text{CO}_3)_y$ ceramics with controlled
501 composition. *J. Eur. Ceram. Soc.* 28, 139–147.
502 <https://doi.org/10.1016/j.jeurceramsoc.2007.06.009>
- 503 Landi, E., Celotti, G., Logroscino, G., Tampieri, A., 2003. Carbonated hydroxyapatite as bone
504 substitute. *J. Eur. Ceram. Soc.* 23, 2931–2937.
505 [https://doi.org/https://doi.org/10.1016/S0955-2219\(03\)00304-2](https://doi.org/https://doi.org/10.1016/S0955-2219(03)00304-2)
- 506 Lang, Q., Zhang, B., Liu, Z., Jiao, W., Xia, Y., Chen, Z., Li, D., Ma, J., Gai, C., 2019. Properties of
507 hydrochars derived from swine manure by CaO assisted hydrothermal carbonization. *J.*
508 *Environ. Manage.* 233, 440–446. <https://doi.org/10.1016/j.jenvman.2018.12.072>
- 509 Lu, H., Reddy, E.P., Smirniotis, P.G., 2006. Calcium Oxide Based Sorbents for Capture of Carbon
510 Dioxide at High Temperatures. *Ind. Eng. Chem. Res.* 45, 3944–3949.
511 <https://doi.org/10.1021/ie051325x>
- 512 Lysikov, A.I., Salanov, A.N., Okunev, A.G., 2007. Change of CO₂ Carrying Capacity of CaO in
513 Isothermal Recarbonation–Decomposition Cycles. *Ind. Eng. Chem. Res.* 46, 4633–4638.
514 <https://doi.org/10.1021/ie0702328>
- 515 Melia, P.M., Cundy, A.B., Sohi, S.P., Hooda, P.S., Busquets, R., 2017. Trends in the recovery of
516 phosphorus in bioavailable forms from wastewater. *Chemosphere* 186, 381–395.
517 <https://doi.org/10.1016/j.chemosphere.2017.07.089>
- 518 New technological applications for wet biomass waste stream products [WWW Document],
519 n.d. URL <https://cordis.europa.eu/project/rcn/110741/factsheet/en>
- 520 Ottosen, L.M., Kirkelund, G.M., Jensen, P.E., 2013. Extracting phosphorous from incinerated
521 sewage sludge ash rich in iron or aluminum. *Chemosphere* 91, 963–969.
522 <https://doi.org/https://doi.org/10.1016/j.chemosphere.2013.01.101>
- 523 Owsianiak, M., Brooks, J., Renz, M., Laurent, A., 2018. Evaluating climate change mitigation
524 potential of hydrochars: compounding insights from three different indicators. *GCB*
525 *Bioenergy* 10, 230–245. <https://doi.org/10.1111/gcbb.12484>
- 526 Owsianiak, M., Ryberg, M.W., Renz, M., Hitzl, M., Hauschild, M.Z., 2016. Environmental
527 Performance of Hydrothermal Carbonization of Four Wet Biomass Waste Streams at
528 Industry-Relevant Scales. *ACS Sustain. Chem. Eng.* 4, 6783–6791.

529 <https://doi.org/10.1021/acssuschemeng.6b01732>

530 Poinern, G.J.E., Brundavanam, R., Le, X.T., Djordjevic, S., Prokic, M., Fawcett, D., 2011. Thermal
531 and ultrasonic influence in the formation of nanometer scale hydroxyapatite bio-ceramic.
532 *Int. J. Nanomedicine* 6, 2083–2095. <https://doi.org/10.2147/IJN.S24790>

533 Singh, A., Purohit, K.M., 2011. Chemical Synthesis, Characterization and Bioactivity Evaluation
534 of Hydroxyapatite Prepared from Garden snail (*Helix aspersa*). *J. Bioprocess. Biotech.*
535 <https://doi.org/10.4172/2155-9821.1000104>

536 Smith, A.M., Whittaker, C., Shield, I., Ross, A.B., 2018. The potential for production of high
537 quality bio-coal from early harvested *Miscanthus* by hydrothermal carbonisation. *Fuel*
538 220, 546–557. <https://doi.org/10.1016/j.fuel.2018.01.143>

539 Straub, D.A., 2007. Calcium Supplementation in Clinical Practice: A Review of Forms, Doses,
540 and Indications. *Nutr. Clin. Pract.* 22, 286–296.
541 <https://doi.org/10.1177/0115426507022003286>

542 Tamaki, M., Kannan, S., Jaidev, T., Raj, S., 2009. Physical characterization of dibasic calcium
543 phosphate dihydrate and anhydrate. *J. Pharm. Sci.* 98, 905–916.
544 <https://doi.org/10.1002/jps.21443>

545 Tamimi, F., Sheikh, Z., Barralet, J., 2012. Dicalcium phosphate cements: Brushite and monetite.
546 *Acta Biomater.* 8, 474–487. <https://doi.org/https://doi.org/10.1016/j.actbio.2011.08.005>

547 Titirici, M.-M., White, R.J., Falco, C., Sevilla, M., 2012. Black perspectives for a green future:
548 Hydrothermal carbons for environment protection and energy storage. *Energy Environ.*
549 *Sci.* 5, 6796–6822. <https://doi.org/10.1039/c2ee21166a>

550 Titirici, M.M., Thomas, A., Yu, S.-H., Mueller, J.-O., Antonietti, M., 2007. A Direct Synthesis of
551 Mesoporous Carbons with Bicontinuous Pore Morphology from Crude Plant Material by
552 Hydrothermal Carbonization. *Chem. Mater.* 19, 4205–4212.
553 <https://doi.org/10.1021/cm0707408>

554 Titirici, M.M., Thomas, A., Yu, S.H., Müller, J.O., Antonietti, M., 2007. A direct synthesis of
555 mesoporous carbons with bicontinuous pore morphology from crude plant material by
556 hydrothermal carbonization. *Chem. Mater.* 19, 4205–4212.
557 <https://doi.org/10.1021/cm0707408>

558 Ucar, S., Bjørnøy, S.H., Bassett, D.C., Strand, B.L., Sikorski, P., Andreassen, J.-P., 2017.
559 Transformation of brushite to hydroxyapatite and effects of alginate additives. *J. Cryst.*
560 *Growth* 468, 774–780. <https://doi.org/https://doi.org/10.1016/j.jcrysgro.2016.11.019>

561 Wang, T., Zhai, Y., Zhu, Y., Gan, X., Zheng, L., Peng, C., Wang, B., Li, C., Zeng, G., 2018.
562 Evaluation of the clean characteristics and combustion behavior of hydrochar derived
563 from food waste towards solid biofuel production. *Bioresour. Technol.* 266, 275–283.
564 <https://doi.org/10.1016/j.biortech.2018.06.093>

565 Weidema, B.P., Bauer, C., Hischer, R., Mutel, C., Nemecek, T., Reinhard, J., Vadenbo, C.O.,
566 Wernet, G., 2013. Overview and methodology - Data quality guideline for theecoinvent
567 database version 3.

568 Yadav, R.S., Meena, S.C., Patel, S.I., Patel, K.I., Akhtar, M.S., Yadav, B.K., Panwar, J., 2012.
569 Bioavailability of Soil P for Plant Nutrition BT - Farming for Food and Water Security, in:
570 Lichtfouse, E. (Ed.), . Springer Netherlands, Dordrecht, pp. 177–200.
571 https://doi.org/10.1007/978-94-007-4500-1_8

- 572 Zhang, L., Zeng, G., Dong, H., Chen, Y., Zhang, J., Yan, M., Zhu, Y., Yuan, Y., Xie, Y., Huang, Z.,
573 2017. The impact of silver nanoparticles on the co-composting of sewage sludge and
574 agricultural waste: Evolutions of organic matter and nitrogen. *Bioresour. Technol.* 230,
575 132–139. <https://doi.org/https://doi.org/10.1016/j.biortech.2017.01.032>
- 576 Zhang, L., Zhang, J., Zeng, G., Dong, H., Chen, Y., Huang, C., Zhu, Y., Xu, R., Cheng, Y., Hou, K.,
577 Cao, W., Fang, W., 2018. Multivariate relationships between microbial communities and
578 environmental variables during co-composting of sewage sludge and agricultural waste in
579 the presence of PVP-AgNPs. *Bioresour. Technol.* 261, 10–18.
580 <https://doi.org/https://doi.org/10.1016/j.biortech.2018.03.089>
- 581
- 582

583 [Electronic supplementary information \(ESI\)](#)

584

585 [Evaluation of hydrothermal carbonization in urban mining for the](#)
586 [recovery of phosphorus from the organic fraction of municipal](#)
587 [solid waste](#)

588

589 Borja Oliver-Tomas,^{a,c} Martin Hitzl,^a Mikołaj Owsianiak,^b Michael Renz*^c

590

591 ^a Ingelia, S.L., Spain

592 ^b Technical University of Denmark, Denmark

593 ^c Instituto de Tecnología Química (UPV-CSIC), Spain; E-mail: mrenz@itq.upv.es

594

595 [Content](#)

596 Experimental details on the recovery of phosphorus as a solid

597 Figure S1. Ash composition of hydrochar obtained from ICP-OES analysis. Amounts are
598 expressed as weight% of the oxides of the elements. Initial ash content was 13 wt% on dry
599 base.

600 Figure S2. Phosphorus extraction efficiency with different mineral acids at different
601 temperatures.

602 Figure S3. XRD pattern of the precipitate (a) and of the precipitate after calcination (b).

603 Figure S4. XRD pattern of the hydrochar before extraction.

604 Figure S5. FTIR spectra a) of the precipitate and the precipitate after calcination and b) of the
605 precipitate in comparison with brushite.

606 Figure S6. Thermogravimetric (TG) analysis of the precipitate under nitrogen atmosphere.

607 Figure S7. Composition of a) precipitate and b) precipitate after calcination, measured by ICP-
608 OES.

609 Figure S8. Solubility of the two main components of the precipitate.

610 Figure S9. Impact scores for scenarios 1–3 and 7–9 in Table 3, expressed in %, normalized
611 internally to “No phosphorus extraction” scenario, for which impacts are equal either 100%
612 (for positive impact scores) or –100% (for negative impact scores).

613 Figure S10. Impact scores for scenarios 1–3 and 10–12 in Table 3, expressed in %, normalized
614 internally to “No phosphorus extraction” scenario, for which impacts are equal either 100%
615 (for positive impact scores) or –100% (for negative impact scores).

616

617 Recovery of phosphorus as a solid

618 For transportation means, solid phosphorus is preferred over dissolved and diluted
619 phosphorus. Therefore, the precipitation of the phosphorus from the extraction solution was
620 studied. By addition of lime ($\text{Ca}(\text{OH})_2$) pH was shifted to slightly alkaline (8–9.5) and up to
621 quantitative precipitation (99%) was achieved and the phosphor precipitate deeply analyzed.
622 Precipitation process and solid product characteristics might be influenced by components of
623 the HTC process water that was employed for the extraction (after adding concentrated
624 mineral acid) with the aim to simulate real world conditions.

625 By X-ray powder diffraction (XRD) calcium hydrogen phosphate (brushite) was identified in the
626 solid, contaminated by calcium carbonate (Figure S3a). As it was confirmed that calcium
627 carbonate was not present in the starting material, i.e. the hydrochar employed in the
628 extractions (Figure S4; at least not in crystalline form), it had to be concluded that the
629 carbonate was formed by incorporation of ambient carbon dioxide, probably during the
630 filtration processes in the laboratory. Hence, this might be avoided by changing the
631 experimental setup. The XRD pattern of the hydrochar indicated the presence of sand and
632 hydroxyapatite (Figure S4). The latter also reveals a potential origin of the phosphorus: up to
633 70% by weight of (human) bone is a modified form of hydroxyapatite, known as bone
634 mineral.(Straub, 2007)

635 The presence of both components in the precipitate, calcium hydrogen phosphate and
636 carbonate, was confirmed by IR spectroscopy (Figure S5). Calcium carbonate is evidenced by
637 bands at wavenumbers of 1430 cm^{-1} , 890 cm^{-1} and 710 cm^{-1} (green dashed lines), being the
638 first one the strongest and the most characteristic one.(Landi et al., 2003) Comparison of the
639 spectrum with the one of brushite (Figure S5b; blue dashed lines) permits to detect
640 characteristic bands such as 1645 cm^{-1} associated with water molecules in the lattice (Ucar et
641 al., 2017) and associated with the presence of the HPO_4^{2-} ion at 1135, 1060, 985, 575 and 525
642 cm^{-1} .(Ucar et al., 2017) The band observed at 1385 cm^{-1} (red dashed line) indicated
643 contamination with a small amount of nitrate from the extraction solution.(Lafon et al., 2008)

644 The presence of carbonate and nitrate species was further corroborated by their
645 decomposition by calcination at $815\text{ }^\circ\text{C}$. Hence, after calcination, the XRD pattern indicated the
646 presence of calcium hydroxyapatite and calcium phosphate (Figure S3b). In the FTIR spectrum,
647 the characteristic bands for carbonate and nitrate at 1430 and 1385 cm^{-1} were clearly
648 diminished (Figure S5a) whereas the characteristic phosphate band at 1030 cm^{-1} was clearly
649 evidenced.(Poinern et al., 2011; Singh and Purohit, 2011) Therefore, it can be stated that both
650 techniques confirmed a transformation of brushite and calcite during calcination.

651 When the decomposition process was elucidated in detail by thermogravimetric (TG) analysis,
652 it was seen that two major weight losses occurred, namely up to $200\text{ }^\circ\text{C}$ and between 600 and
653 $900\text{ }^\circ\text{C}$ (Figure S6). Both losses were in accordance with literature reports for the involved
654 materials. Differential scanning calorimetry has determined 190 to $200\text{ }^\circ\text{C}$ as desorption
655 temperature of the crystal water molecules of dibasic calcium phosphate dihydrate.(Tamaki et
656 al., 2009) In addition, calcium carbonate is transformed into calcium oxide at temperatures of
657 $700\text{ }^\circ\text{C}$ to $850\text{ }^\circ\text{C}$ (Lu et al., 2006; Lysikov et al., 2007) and nitrate breakdown occurred almost in
658 the same interval ($680\text{ }^\circ\text{C}$) under dynamic rising temperature conditions.(Ettarh and Galwey,
659 1996) The second weight loss was also in accordance with the behavior of mixed materials.
660 Hence, thermal decomposition of carbonated apatite started at temperatures higher than 600
661 $^\circ\text{C}$ in inert atmosphere:(Lafon et al., 2008) by IR measurements it has been confirmed that

662 after calcination in air at 600 °C decarbonatation had not occurred, whereas after the same
663 procedure at 800 °C only one third of the initial amount remained.

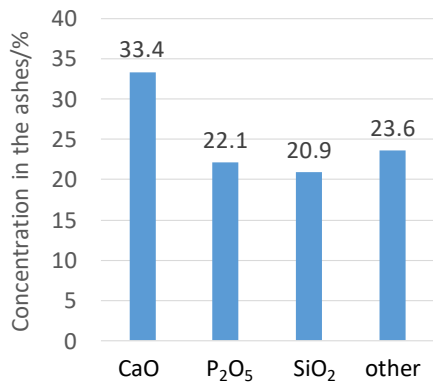
664 Chemical composition of the precipitate was further analyzed and quantified, namely
665 phosphorus by inductively coupled plasma – optical emission spectroscopy (ICP-OES) and
666 calcium by X ray fluorescence (XRF) spectroscopy. The combined results indicated a 41 wt%
667 content of $\text{CaHPO}_4 \cdot 2 \text{H}_2\text{O}$ and a 47% content of CaCO_3 in the precipitate (Figure S7a). After
668 calcination, 64 wt% of the obtained solid consisted of $\text{Ca}_3(\text{PO}_4)_2$ and 30 wt% of CaO . There is a
669 small incoherency with the DRX pattern as CaO was not detected. However, this might be due
670 to an amorphous or nanocrystalline nature of the material.

671 For the suitability of a phosphorus-containing material as solid fertilizer, water solubility is
672 crucial. Plants absorb the inorganic phosphate either in the form of HPO_4^{2-} or of H_2PO_4^- from
673 the soil solution.(Yadav et al., 2012) However, brushite has been employed as a dicalcium
674 phosphate cement, suitable for bioceramics.(Tamimi et al., 2012) Therefore, it can be expected
675 that it possessed a low water solubility. In addition, hydroxyapatite is treated as a stable
676 mineral under surface Earth conditions and adopted as a proxy for paleo-environmental
677 reconstructions and in choice of remediation strategies.(Boudia et al., 2018) As a consequence,
678 it was mandatory to evaluate the solubility of the precipitate, so that it could be demonstrated
679 that the extracted and precipitated phosphorus fulfilled the minimum criteria as fertilizer.

680 From Figure S8 it can be seen that the phosphorus component dissolved well in neutral water,
681 whereas the calcite was only sparsely soluble. For the calcined precipitate solubility was the
682 contrary (not shown): calcium oxide dissolved readily whereas the phosphorus did not dissolve
683 at all. From this it can be concluded that calcination of the precipitate, or combustion of
684 hydrochar, is detrimental for phosphorus availability for plant growth. However, phosphorus
685 as precipitate should be readily available for the plants.

686

687 *Figure S1. Ash composition of hydrochar obtained from ICP-OES analysis. Amounts are expressed as weight% of the*
688 *oxides of the elements. Initial ash content was 13 wt% on dry base.*

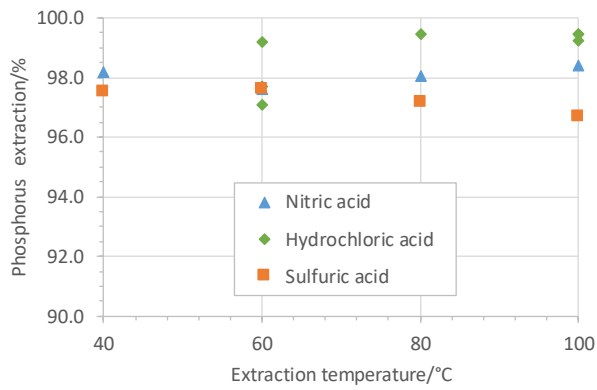


689

690

691

692 *Figure S2. Phosphorus extraction efficiency with different mineral acids at different temperatures.*



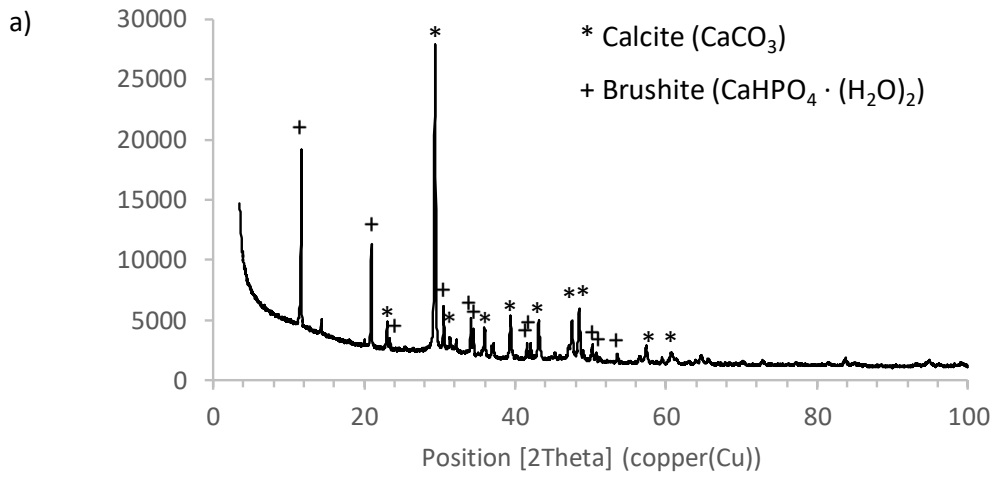
693

694

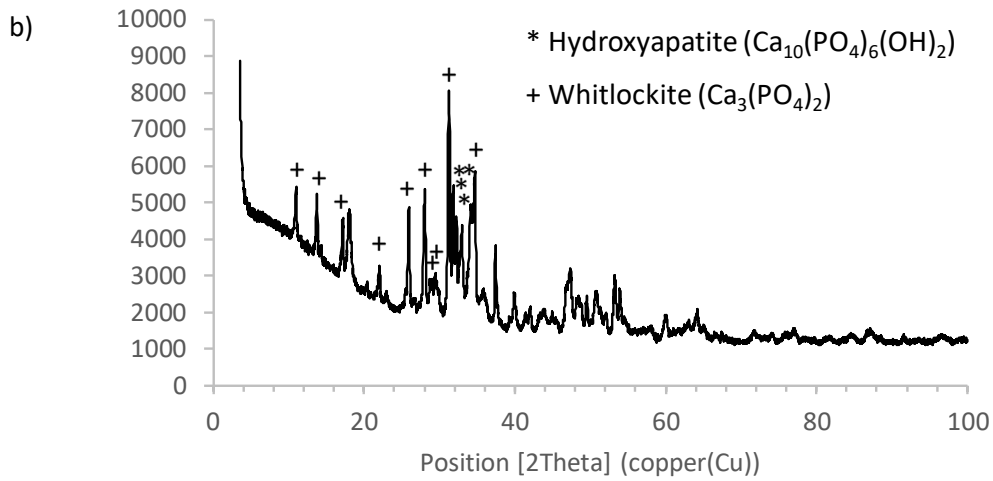
695

696

697 *Figure S3. XRD pattern of the precipitate (a) and of the precipitate after calcination (b).*



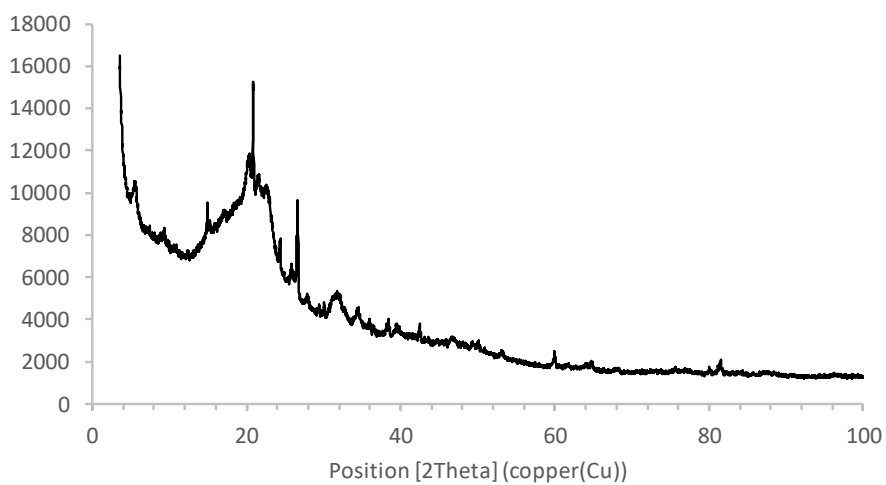
698



699

700

701 *Figure S4. XRD pattern of the hydrochar before extraction.*



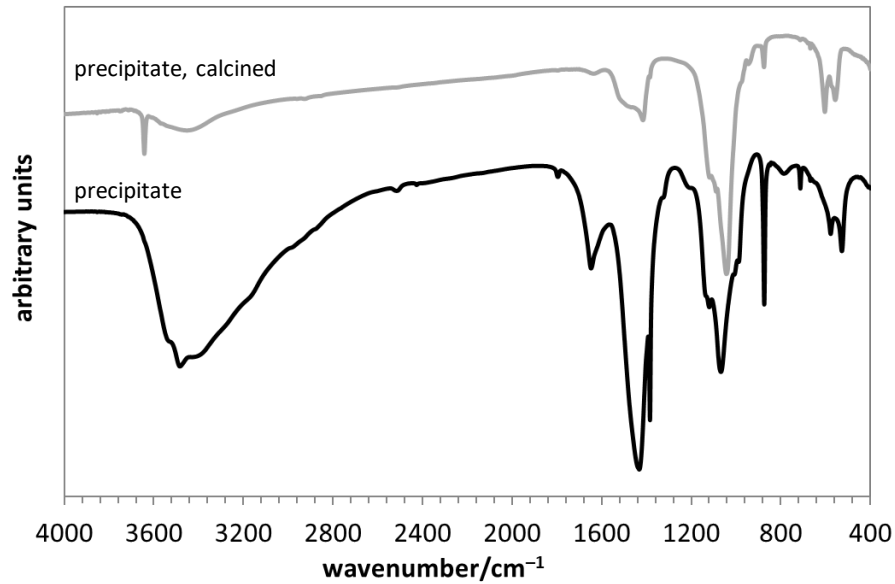
702

703

704

705 *Figure S5. FTIR spectra a) of the precipitate and the precipitate after calcination and b) of the precipitate in*
706 *comparison with brushite.*

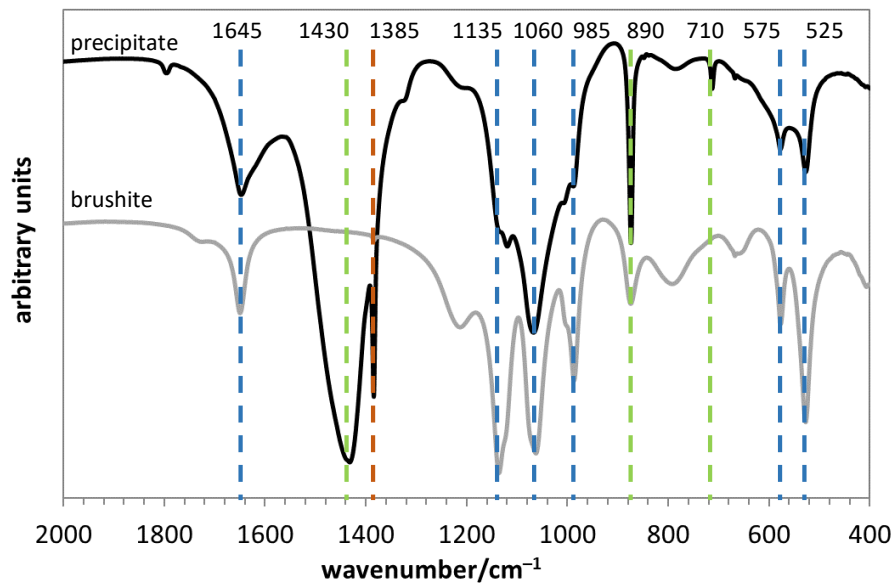
707 a)



708

709

710 b)



711

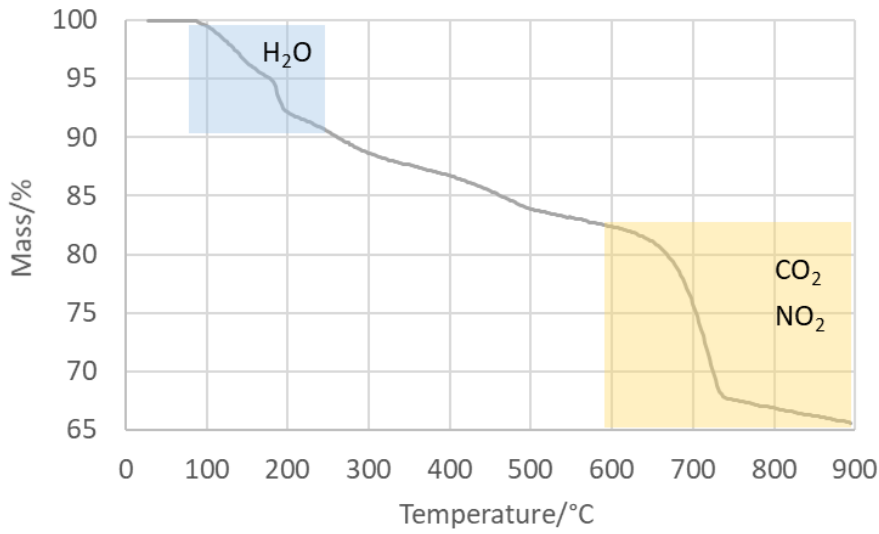
712

713

714

715

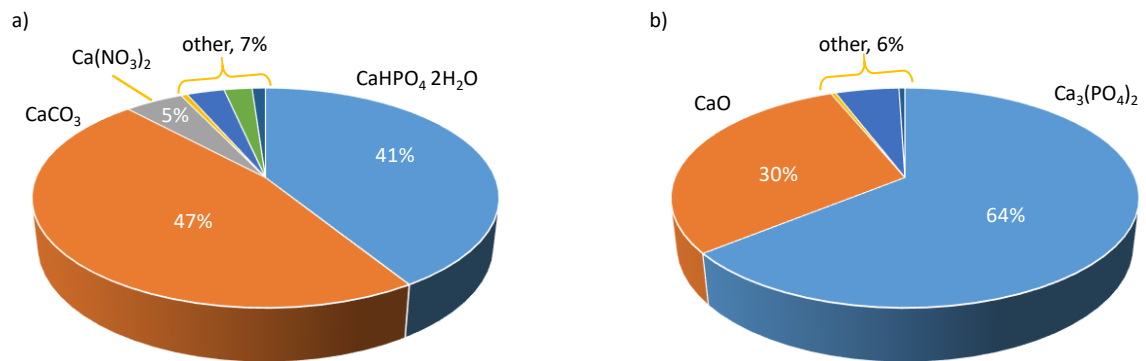
716 *Figure S6. Thermogravimetric (TG) analysis of the precipitate under nitrogen atmosphere.*



717

718

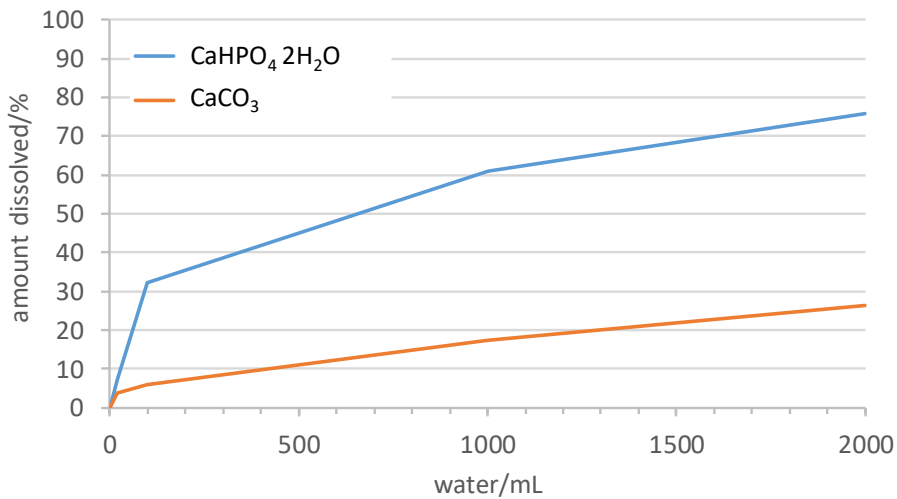
719 *Figure S7. Composition of a) precipitate and b) precipitate after calcination, measured by ICP-OES.*



720

721

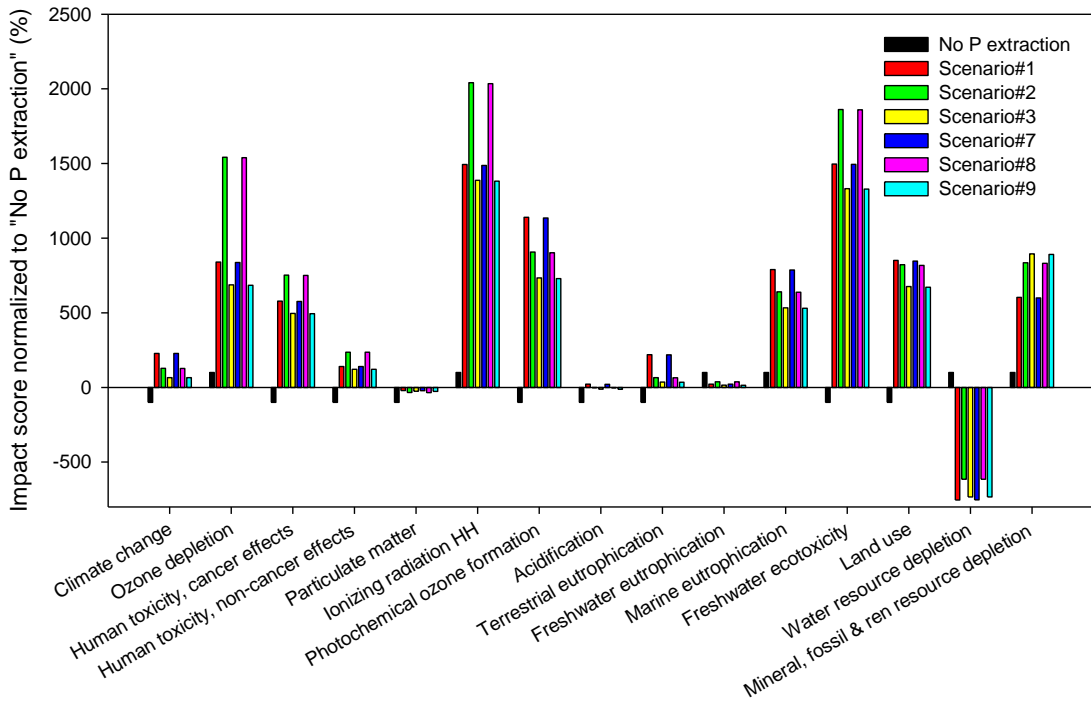
722 *Figure S8. Solubility of the two main components of the precipitate.*



723

724 Additional LCA results

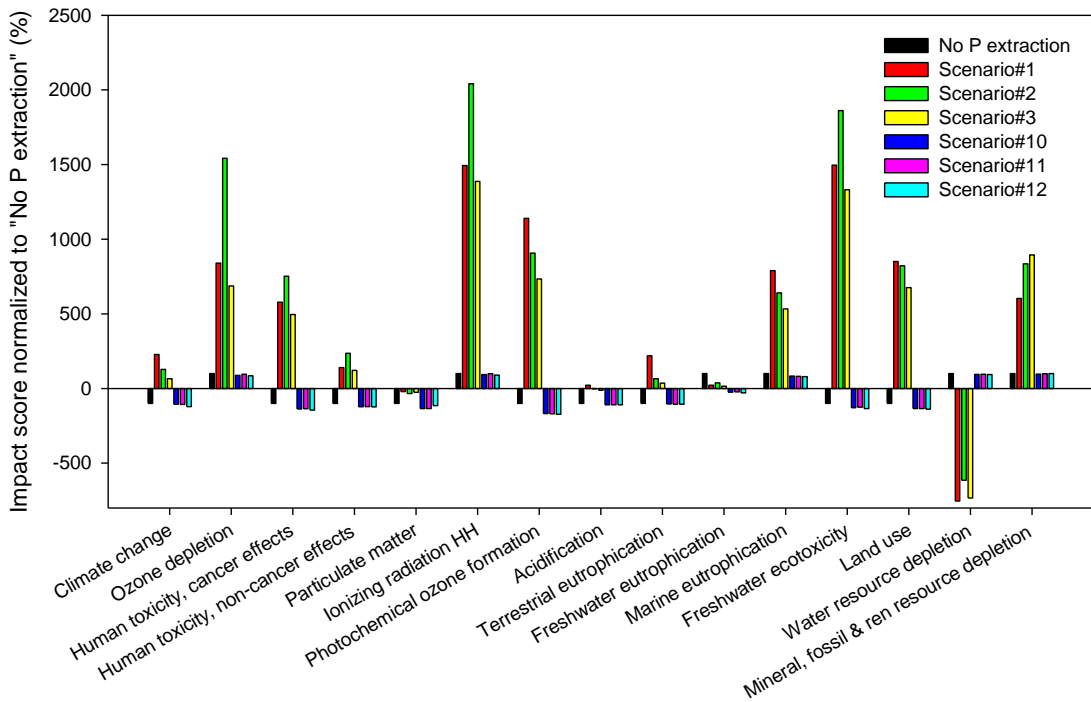
725 Figure S9. Impact scores for scenarios 1–3 and 7–9 in Table 3, expressed in %, normalized internally to “No
 726 phosphorus extraction” scenario, for which impacts are equal either 100% (for positive impact scores) or –100% (for
 727 negative impact scores).



728

729

730 Figure S10. Impact scores for scenarios 1–3 and 10–12 in Table 3, expressed in %, normalized internally to “No
 731 phosphorus extraction” scenario, for which impacts are equal either 100% (for positive impact scores) or –100% (for
 732 negative impact scores).



733