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Additional Information

Drought early warning based on optimal risk forecasts in
 regulated river systems: application to the Jucar River Basin
 (Spain)

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9

10 Abstract

11 Droughts are a major threat to water resources systems management. Timely anticipation 12 results crucial to defining strategies and measures to minimise their effects. Water managers 13 make use of monitoring systems in order to characterise and assess drought risk by means of 14 indices and indicators. However, there are few systems currently in operation that are capable 15 of providing early warning with regard to the occurrence of a drought episode. This paper 16 proposes a novel methodology to support and complement drought monitoring and early 17 warning in regulated water resources systems. It is based in the combined use of two models, 18 a water resources optimization model and a stochastic streamflow generation model, to 19 generate a series of results that allow evaluating the future state of the system. The results for 20 the period 1998-2009 in the Jucar River Basin (Spain) show that accounting for scenario 21 change risk can be beneficial for basin managers by providing them with information on the 22 current and future drought situation at any given moment. Our results show that the 23 combination of scenario change probabilities with the current drought monitoring system can 24 represent a major advance towards improved drought management in the future, and add a 25 significant value to the existing national State Index (SI) approach for early warning purposes.

26

Highlights 28

29	 Modelling the past to anticipate future drought is an ineffective and risky approach
30	• A new method for continuous drought monitoring and early warning in regulated
31	catchments is proposed
32	• Reservoir storage probability is a reliable indicator for drought status in regulated
33	catchments
34	 New approach adds value to existing monitoring and early warning methods
35	Keywords
36	Monitoring; Early Warning System; Optimisation Modelling; Water Resources Systems
37	Analysis; Aquatool
38	
39	1. Introduction

40 Droughts are a major threat to the sound operation and management of water resources 41 systems. Developing new approaches to anticipate them will help in defining strategies and 42 measures to minimise their effects. The use of monitoring systems to calculate drought indices 43 and indicators can help water managers characterize droughts and define risk scenarios. The 44 activation of a drought scenario in a system will trigger a number of measures addressed to 45 minimise the possibilities of developing into a worse scenario and minimizing the possible effects of the current situation. 46

47 The assessment of drought severity requires the use of an index which fulfils well-known criteria (Tsakiris et al. 2013): operational usefulness, physical meaning, sensitivity to a wide 48 49 range of drought conditions, applicability in all parts of the globe, guick response to changes 50 due to drought and high availability of required data. Commonly, such an index is a prime variable for assessing the effect of a drought and defining different drought parameters, which 51

52 include intensity, duration, severity and spatial extent as defined by Yevjevich (1967) in his 53 theory of runs. A time series of drought indices provides a framework for evaluating drought parameters of interest. Generally, drought indices are categorized as meteorological, 54 hydrological, agricultural or remote sensing-based (Rossi and Cancelliere 2013). Mishra and 55 56 Singh (2010) and Pedro-Monzonis et al. (2015) made an extensive review of existing 57 univariate drought indices both concluding that each index performance is region specific 58 mostly due to the characteristics of the variables used for their calculation and the purpose of 59 the analysis. In addition, in recent time some authors have also attempted to combine all the 60 variables (e.g. precipitation, soil, water content) that lead to different physical forms of drought 61 in so-called multivariate drought indices (Rajsekhar et al. 2015). In some cases, the index is 62 built as an aggregation of variables selected according to their relation each drought type 63 (Keyantash and Dracup 2004; Rajsekhar et al. 2015). Inother, the index is constructed using 64 copulas to derive the joint distribution of two or more variables (Kao and Govindaraju 2010; 65 Hao and AghaKouchak 2013).

66 An indicator system is a drought monitoring system that allows the anticipation in the 67 application of mitigation measures for the reduction of socio-economic and environmental impacts of droughts (Estrela and Vargas 2012). Such systems can also be considered early 68 69 warning systems for their capacity to anticipate the effects that drought may have on the 70 system in order to trigger necessary mitigation measures (Rossi et al. 2008). In most cases, 71 these systems are normally formed by basic variables selected at different points in a river 72 basin that are capable of defining the current drought status. Their reliability will depend on 73 their capacity to represent, using real-time data: 1) the relationship between significant 74 reductions of water availability with deviations of meteorological and hydrological components 75 from their average; 2) detecting early stages of drought development; 3) provide results that 76 allow comparison between events both in time and space; and 4) assessing the severity of 77 the ongoing situation in order to support decision making for triggering drought mitigation 78 actions. Additionally, in the case of regulated water resources systems, it would be desirable

that the indicator is capable of showing the evolution of management and how this wouldchange the drought status of the system if new operation rules are envisaged.

81 Different drought early warning systems have been developed at different spatial scales, but 82 a very small number of such systems are actually in operation (Rossi and Cancelliere 2013). This is mainly due to the low density of meteorological and hydrological gauging networks, the 83 84 sharing of the data among different agencies with different objectives, and to the lack of 85 universal standards in computing drought indices (Rossi 2003). In addition, the development 86 of indicator systems based on observational frameworks cannot provide sufficient anticipation 87 with regard to the event in progress in order to activate the necessary measures to mitigate 88 its effects (Haro et al. 2014). Efforts have been made to correlate drought indices to impacts 89 (Stagge et al. 2015), but these relationships only provide insight after the event has finished 90 and the impacts reported. Mishra and Singh (2011) acknowledged that to develop suitable 91 techniques for forecasting the onset and termination of droughts is still a major research 92 challenge due to the inability to predict drought conditions accurately for months or years in 93 advance. Due to these inaccuracies and uncertainties, drought management relies nowadays 94 mainly on risk assessment. Risk assessment during the operation phase of a system is often 95 referred as conditioned risk assessment. With this procedure, the state of the system is usually 96 evaluated for the short-term to explore alternative mitigation measures and policies for an 97 ongoing drought episode. This same assessment approach can be adopted for early warning purposes (Cancelliere et al 2009). 98

Alecci et al. (1986) considered that the risk assessment of a water supply system is a problem that is better approached through a set of several indices and analysing the probability of suffering shortages of different entities. This is due to the many complexities existing within a water resources system such as the stochastic nature of inflows, the high interconnection that exists between different components of the system, the competition for water by conflicting demands, the definition of what elements are at risk, and the uncertain character of the impacts in different drought episodes. Traditionally, reliability, resiliency and vulnerability have

106 been the indices used to capture the different performance aspects of water supply systems 107 (Hashimoto et al. 1982). However, these indices are normally representative of just one 108 particular use, defining the state of the system with regard to the probability of a failure for 109 such index. Since all drought events are unique, so too are their effects both temporally and 110 spatially. Therefore, it is necessary to have an indicator that is capable of summarising the 111 state of the system for any given situation. In regulated systems, it will be the volume stored 112 in reservoirs since it provides an overview of the previous management of the system and is 113 the basis for future resources allocation.

114 This paper proposes a novel methodology to support drought monitoring and scenario 115 definition in regulated water resources systems. It is based on the results of two models, an 116 optimisation model and a stochastic streamflow generation model, both of which have been 117 calibrated and validated in previous research (Haro et al. 2012a, 2012b, and 2014b; Ochoa-118 Rivera 2002). Using storage in reservoirs as a summary indicator of the future system status, 119 we propose a combined use of the two models to generate a series of results that can support 120 and complement drought monitoring and early warning systems currently in place in a river 121 basin. The methodology is applied to the Jucar River Basin in Spain to evaluate the probability 122 of a scenario change several years in advance. The proposed method has the potential to 123 enhance decision making under highly uncertain hydrological situations, and provide water 124 resource planners and managers with new insights both regarding the behavior of the system 125 and the development of drought episodes.

126 **2. Case study description**

127 The Jucar River Basin is located in the eastern part of the Iberian Peninsula in Spain (Figure 128 1). This basin is the most important of the 9 water exploitation systems in the Jucar River 129 Basin Demarcation (Demarcacion Hidrografica del Jucar – DHJ in Spanish). In the Valencia 130 coastal plain, where the Jucar River has its mouth, there is a shallow lake called Albufera, with 131 an associated wetland. Both, the lake and the wetland depend on return flows from irrigated areas in the basin, and also on groundwater flows from the coastal aquifer beneath the plain
(Andreu et al. 2009). It is the largest system of the DHJ both in surface (22,261 km²) and in
volume of resources (1,548 hm³/year).

135 The river is an example of a typical Mediterranean river, characterized by a semi-arid climate 136 in most of the basin territory consisting of low precipitation rates (475mm/year) during the year 137 combined with exceptional convective storms that can lead to flooding and seasonal summer 138 scarcity that occurs when irrigation requirements are at their highest. Urban demand accounts 139 for circa 143.3 hm³/year and the water demand for irrigated agriculture reaches 1034.3 140 hm³/year. Water supply to small urban areas comes mainly from wells and springs, but large 141 metropolitan areas such as Albacete, Sagunto and Valencia rely on surface water (Andreu et 142 al. 2009). According to the White Book of Groundwater (CEDEX 1995), nearly three quarters 143 (73%) of the resources in the territory of the DHJ have subterranean origin. This highlights the 144 major importance that groundwater resources have in the management of these basins. The 145 total amount of available groundwater resources in the basin is 1,225 hm³/year. However, this 146 only represents the estimated volume in all the groundwater bodies without accounting for 147 their sharing between other basins or the relationship these bodies have with the surface water 148 system.

149 With regard to droughts, the Jucar River Basin can be considered to be one of the most 150 vulnerable areas in the western Mediterranean region, due to high water exploitation indexes, 151 and the environmental and water quality problems that arise when droughts occur. This 152 situation has triggered increased use of non-conventional resources in recent years, such as 153 reuse of wastewater and drought emergency wells. Also, conjunctive use of surface-ground 154 waters has historically been a very important option in the region to provide robustness against 155 droughts. The integrated use of these three resource options was considered a major success 156 in adapting to the latest drought episode between 2005 and 2008 (Ortega-Reig et al 2014).

The operation of the system is mainly multi-year. The Alarcon and Contreras reservoirs, at the headwaters of the system, are capable of storing the highly variable streamflow coming from their upstream sub-basins. The third most important reservoir in the system, the Tous, is operated on an annual basis. Before the summer season it stores incoming mid-basin streamflow and upstream reservoirs releases to supply the different demands within the Valencia Plain. By the end of the summer, the reservoir is emptied in order to prevent floods originated from often intense autumn rainfall events.

164 **3. Methodology**

165 In this section, we present the indicator system currently in use in the Jucar River basin as 166 well as in most of Spanish river basins. Despite being a useful methodology to evaluate the 167 actual drought conditions in the basin, it has low forecasting capacity; making preventive management of droughts inefficient and/or very difficult. To complement the information 168 169 provided by the indicator, we developed a methodology to derive the probability of drought 170 scenario change for a four year planning horizon. It is based on the Monte Carlo evaluation of 171 the results of multiple runs of an optimization model of the system. Based on this analysis, we 172 derive distribution functions on the future state of the basin and combine them with trigger 173 values for each drought scenario.

174 <u>3.1. Current drought indicator system for Spanish river basins</u>

175 One of the objectives of Spanish Drought Plans is providing means for anticipating drought 176 events. To do this, it is necessary to establish an early warning system that allows forecasting 177 drought characteristics and assessing their effects on the system. Spanish basin operators 178 have adopted a method of drought indicators based on the analysis of historic data that reflect 179 the availability of water in the system. This indicator is known as State Index (SI) and it is the 180 result of combining several hydro-meteorological variables obtained from a monitoring system. 181 The SI has a hydrologic character since its practical interest lays on its ability to serve as 182 decision-making instrument regarding water resources management in the basin. For each 183 catchment, managers select a set of variables that best represent the water resources for
184 different demand units in the basin using values of reservoirs storage, piezometric levels,
185 natural streamflow and areal precipitation. In the case of the Jucar River, the selected
186 variables are detailed in CHJ (2007)¹.

187 For each selected variable, the value of the SI has the following expression (CHJ 2007):

$$If V_i \ge V_{av} \rightarrow SI = \frac{1}{2} \cdot \left[1 + \frac{V_i - V_{av}}{V_{max} - V_{av}} \right]$$
Eq. 1

$$If V_i < V_{av} \rightarrow SI = \frac{1}{2} \cdot \frac{V_i - V_{min}}{V_{av} - V_{min}}$$
Eq. 2

188

Where V_i is the value of the variable in month i; V_{av} is the average monthly value of the variable 189 190 in the historic series considered; and V_{max} and V_{min} are the maximum and minimum monthly 191 values of the variable in the historic series considered respectively. The main reason to follow 192 this calculation approach is that the arithmetic average is a robust statistic, as well as simple; 193 so a comparison of the current variable value with the average of the historic series considered 194 will adjust better to the real situation of the studied region. Additionally, taking into account the 195 maximum and the minimum historic values allows homogenising the different variables into a 196 dimensionless numeric value capable of quantifying the current situation with regard to the 197 historic. This also permits to quantitatively compare the different variables selected between 198 them. Finally, the overall SI of the basin and hence its drought level is defined as the weighted sum of the SI values of each of the selected hydro-meteorological variables. The weight 199 200 assigned to each variable depends on the level of demand served. For the Jucar River, the SI 201 consists of a combination of 12 different variables including precipitation, streamflow, 202 piezometric levels and storage in reservoirs at different strategic points within the basin (CHJ 203 2007).

¹ A partial translation of the contents in CHJ(2007) is provided in Acacio et al. (2013)

204 Spanish Drought Plans establish four different levels of drought, or scenarios, namely: 205 normality, pre-alert, alert and emergency (CHJ 2007). These levels are determined according 206 to the values of the SI with the following thresholds: Normality (SI≥0.5); Pre-alert (0.5>SI≥0.3); 207 Alert (0.3>SI≥0.15); and Emergency (0.15>SI). Figure 2 shows the evolution of the SI in the 208 Jucar River Basin between October 1998 and September 2010. Between the end of the XX 209 century and the beginning of the XXI century the basin experienced a short but intense period 210 of drought that made the SI oscillate between the pre-alert and the alert levels until 2002 when 211 the situation returned to normality after a period of intense precipitation. Between 2005 and 212 2008, the system suffered the worst drought event on record with SI reaching emergency 213 levels several times during that period. After that, the system gradually recovered to pre-alert 214 in 2009 to finally reach the normality level in 2010.

Haro et al. (2014) showed the possibility that an indicator such as the SI might be insufficient in order to set and trigger the most appropriate drought mitigation measures early enough to be efficient. This method is limited to determine the current drought situation based on the comparison of present variables values with the variables occurred in the past; making its forecasting capability low, or even non-existent. Moreover, drought episodes vary between one and another. Hence, it is very unlikely that the SI is capable of working as an early warning system for droughts, advancing the real consequences of an upcoming event.

In addition, as commented above, it is important that the effects of management decisions and mitigation measures are included in the monitoring process and that their modifications are reflected in order to advance their efficacy and to better support decision-making. For this reason, the use of risk assessment methodologies in combination with indicator systems provides an interesting and novel framework to support decision making during drought situations in regulated systems.

228 <u>3.2. Drought scenario definition based on the risk assessment of the system's optimal</u>
 229 <u>operation</u>

230 The methodology developed is based on previous research by Sanchez-Quispe (1999), 231 Andreu and Solera (2006), Andreu et al (2007, and 2013) and Cancelliere et al (2009). Their 232 findings were successfully used in the management of previous drought episodes of the Jucar 233 River Basin. Here we present a further development of existing approaches by introducing an 234 optimisation approach that allows one to obtain the best results achievable in the system and better rules for the application of mitigation and prevention measures. This work further 235 develops that presented by Haro et al. (2014a) by extending its application to a multi-year 236 237 regulated basin. In addition, we show how the risk assessment methodology presented here 238 is applicable to forecast drought scenarios. Figure 3 provides a schematic summary of the 239 methodology, which is briefly described below.

240 We applied a monthly Monte Carlo optimisation process to a catchment management model 241 of the Jucar River Basin previously developed in the GUI of Aquatool DSS (Andreu et al 1996) 242 for the implementation of the European Water Framework Directive (CHJ 2004) and the 243 development of is latest basin plan (CHJ 2015), and shown in Figure 4. The model includes 244 the main surface storage facilities ('Alarcon', 'Contreras', and 'Tous' reservoirs) as well as the 245 main aguifers in the basin that have a crucial role in the management of the system ('Mancha 246 Oriental' and 'Plana de Valencia'). The most important demands are also represented, namely: 247 traditional irrigation in 'Plana de Valencia'; groundwater irrigation from 'La Mancha Oriental' 248 aquifer; conjunctive irrigation from the newer developments along the 'Jucar-Turia' canal; and 249 the urban demands of Valencia, Sagunto and Albacete, which is minor in quantity but more 250 sensitive to failures in the supply. Haro et al. (2012a and 2012b) and Haro Monteagudo (2014) 251 provide a detailed description of the optimization technique, equations and constraints utilised 252 by the model, as well as the input data it needs. A previous application can also be found in 253 Haro et al (2014b). The model runs on a monthly time step fed by synthetic streamflow series 254 generated stochastically from historically observed monthly values between 1980 to 2012. 255 There are 16 streamflow input nodes along the model network, represented as thick red 256 arrows in Figure 4. The synthetic series were generated with the stochastic analysis and

modelling module in Aquatool (Ochoa-Rivera 2002). The 16 observed streamflow time series
were normalised and standardised to calibrate the autoregressive model, AR(1), shown in
equation 3:

260 where X_t and X_{t-1} are *n* variables vectors; φ_1 is an *n* x *n* autocorrelation matrix; θ_0 is an *n* x *n* 261 matrix of coefficients that multiplies the random N(0,1) values vector represented by ε . For 262 this case, n has a value of 16. For the stochastic generation of synthetic streamflow series 263 from observed values, the last monthly observed value is used as a seed after normalisation 264 and standardisation. The generated time series of standardised values are converted to streamflow values following the inverse path. The validation of the model against the long term 265 266 characteristics of the historic series (average, standard deviation, number of dry years), makes it suitable to explore a large range of events. 267

The results of each optimisation run in the Monte Carlo process are the time series of reservoirs storage and releases, surface and groundwater supply to the different demands, aquifers relative storage and recharge, and flows in river streams. The statistical analysis of all runs yields a number of indicators to assess risk.

272 When confronting an ongoing drought situation from a risk minimisation approach and a high 273 level of uncertainty, it is more useful to rely on an index that summarizes the status of the 274 basin considering all the possible events. In the case of regulated river basins, this index is 275 the state of the reservoirs. The evolution of storage in reservoirs clearly reflects the operation 276 of a system during previous periods of time, and their present status defines the future use 277 possibilities. Hence, reservoir level state probability and storage probability are useful 278 indicators with regard to drought in a regulated catchment and may support the decision 279 making process with information about what can be expected in the future.

Based on the previous consideration, we use the storage probability in the different reservoirsin the basin as the basis to determine the risk level and the change of scenario probability at

the end of a number of campaigns for each month. It must be noted that reservoirs levels is an important element in the Jucar River Basin drought indicator system, representing almost 50% of the indicators value. We transform the reservoir levels probability distribution into state index distributions following the calculation method above by comparing the results to the historic series of observed levels. Afterwards, we determine the probability of scenario change for each month by crossing each state index distribution by the threshold levels defined by the state index methodology.

We applied this methodology in the Jucar River Basin for the period between hydrologic years 1998-1999 and 2008-2009. During these 10 years, two of the most important drought episodes for the Jucar River Basin in history took place (CHJ 2007; van Lanen et al. 2013): the short but intense drought of 1999-2000 and the long drought episode between 2005 and 2008.

The optimisation process tends to empty the reservoirs by the end of the optimisation period. Thus, setting the multiple risk assessment runs for just one year would not provide adequate results since we want to make use of the perfect forecast principle of optimisation. Therefore, optimisation periods of four years were used for each run extracting the results of the first year. Three hundred series of 48 months generated with the autoregressive model from equation 3 proved sufficient to yield representative results in the Monte Carlo optimisation process for each monthly run.

300 **4. Results**

301 <u>4.1. State Index complementation with scenario change probability</u>

Figure 5 shows the result of applying the proposed methodology together with the evolution of the Jucar River observed state index for the three first years of the optimisation period considered in each run. The fourth year is disregarded because it coincides with the end of the optimisation period, when the algorithm uses all the available water. For each month, we have the actual drought scenario as defined by the thresholds and the probability of each 307 scenario occurring one to three years later corresponding to Figures 5a to 5c, respectively. In 308 Figure 5a, the probability of a scenario change in the next year is low, with a general tendency 309 to remain at the same level. In Figures 5b and 5c, the probabilities of a scenario change 310 increase after two and three years and how this provides a better insight of what can be 311 expected in the system. With these results, the methodology proposed adds value to the actual 312 State Index by showing the probability that the current situation might change in the future. hence providing additional support for decision makers in terms of activating mitigation 313 314 measures, which normally require some time to start operating appropriately.

315 The probability of scenario change with one year anticipation (Figure 5a) is useful for the 316 middle and end of drought episodes as well as for annually operated systems. For example, 317 soft preventive measures could have been maintained in February 2001 despite the entrance 318 in the normality scenario in order to prevent the posterior quick fall to almost emergency one 319 year later. Conversely, the two and three year anticipation probabilities (Figures 5b and 5c) 320 are useful in detecting the possible start of a drought situation, especially in multi-year 321 systems. Between 2004 and 2008, the State Index dropped from the normality scenario to 322 emergency in about one year (June 2004 to June 2005) and then remained in that situation for two years. This situation is captured in Figures 5b and 5c, where the probabilities of being 323 324 in a scenario worse than normality two and three years after June 2004 exceeded 50%.

325 4.2. Approximation of SI values with risk results

Previous stakeholder participation experiences in the Jucar River with risk assessment tools have shown that, in general, risk results obtained for an 80% probability of exceedance level and one year in advance are trusted as good approximations of the future state of the system. These results can be easily extracted from the tools used to perform the proposed methodology, as well as any other risk level results. Hence, we explored the ability of the proposed methodology to approximate SI from a probabilistic perspective. 332 Figure 6 shows the evolution of SI approximated as the 80% risk level one year in advance 333 versus the actually observed SI in the Jucar River for the period October 1998 through 334 September 2009. Both indices reflect accurately the drought events occurred in the Jucar 335 River basin for the period of study. However, while the risk based SI follows the observed one 336 during the first part of the period, there is a six months delay disconnection right before the 337 beginning of the 2004-2008 drought episode. This is due to the operation of the optimisation 338 process. The objective function in the optimisation model works tries to maximise the stored 339 volume in reservoirs while meeting all the demands and environmental flows, minimising water 340 loses from the system. First, during the wet period prior to the 2004-2008 event, the 341 optimisation model achieves better storage levels before the episode starts because all the 342 demands are met and there is water that would be lost instead at a high cost for the objective 343 function. Since the optimisation process implies perfect forecast, the model is capable of 344 storing that water. Second, when reservoirs are near to empty, like during the drought period, 345 the objective function benefits more from supplying the demands than from storing water. 346 Hence, despite the risk based SI drops below the observed one, the demands still have a 347 better level of supply than in the real situation. Therefore, the risk based results offer an 348 envelope of the actual situation, providing managers with an idea of how the system can be expected to respond at different levels of risk. 349

350 **5. Discussion**

The predictions of the methodology presented improve with respect to the combined use of storage, streamflow and precipitation to define a drought state index because they include both previous precipitation and storage data, as well as information regarding the physical system what allows obtaining its best management options. It also includes up to date information of the human influence on the system by means of water demands for the different sectors, and allows considering the environmental needs of the riverine ecosystems in the form of environmental flows definition. In addition, the presented methodology can be used afterwards to assess the risk level with the existing management rules to evaluate the changes introduced by the mitigation measures. Since the methodology is meant to be used every month to monitor the state of the system, any new measures could be implemented in the model in real time. In this way, it is possible to select the best measures for each case and their optimal application.

363 <u>5.1. Methodological limitations</u>

364 The methodology has a number of inherent limitations. Firstly, it was limited by the quality of 365 the stochastic streamflow series used to drive the whole process. The definition of a good 366 stochastic model requires an amount of previously observed data that is not always going to 367 be available. In addition, depending on the stochastic model used, the generated streamflow 368 series will have a different capacity of capturing the dynamics of hydrology in the system. This, 369 together with the tendency of stochastic series to reach values around the historic average 370 after a number of generations, will limit the risk forecasting ability of the method. In this paper, 371 an autoregressive AR(1) stochastic model was used. Despite being capable of capturing the 372 basic statistical parameters of the observed series, Ochoa-Rivera et al. (2007) showed that 373 the approach to streamflow modelling has a significant influence in the final results. Hence, 374 different modelling methodologies should be explored before implementing the proposed 375 methodology.

376 Secondly, optimisation is a highly resources consuming process. This means that complex 377 models of the system under study will require longer calculation periods than more simple 378 ones. The creation of models capable of representing the reality of the system while 379 maintaining a low degree of computational complexity requires a high level of knowledge and 380 understanding about the system. The Jucar River Basin has been extensively studied by 381 researchers for many years, and the methodology presented here was relatively easily applicable. However, it will not be of immediate use in river basins where water level is scarce 382 383 and/or the relationships between the individual hydrological processes are not clear.

384 Finally, in order to be effective, the methodology and its results must be trusted, but also 385 understood, by those that will be later affected by the decisions derived from its use. The 386 model used in this study was developed conjunctively with the managers and water users of 387 the basin within a participatory process that required reaching agreements for everyone. In 388 the same way, the triggers that define each drought situation and the corresponding measures 389 are the results of negotiations between the different actors in the system. This trust building 390 process is achieved over time and thus, methods such as the one presented here are unlikely 391 to be successful at the beginning of participative management processes. Anyway, as 392 observed in Andreu et al. (2009) and Andreu et al. (2013), the very process of implementing 393 similar methodologies finally resulted in better knowledge of the system and understanding of 394 stakeholders needs with an overall improvement of management.

395 <u>5.2. Implications for drought management</u>

396 Existing drought monitoring systems are normally limited to measure a series of climatic and 397 hydrologic variables and calculating various indices that allow determining what is the state of 398 the system compared to the past. Such is the case of the state index used in Spanish drought 399 management plans shown above. This approach may be useful, if not the only one possible 400 in some cases, but has been revealed insufficient for its use in some systems, especially 401 regulated water resources systems (Haro et al. 2014). Using indicators based on observation 402 of hydrologic variables, and comparison with past data in systems where human activities take 403 place, are unable to represent the changes occurring in the system along time. Anthropogenic 404 actions influence not only river flows themselves with extractions and returns but also runoff 405 production and groundwater recharge, delaying or preventing water from reaching the 406 streams. Accounting for all of this and translating observed flows in one point to natural regime 407 is often an arduous task that is not always rewarded with appropriate results. In addition, the 408 parameters used for drought indices calculation are variable with time. This causes that new 409 maximum and minimum observed values have the chance to change dramatically the shape 410 of the indicator evolution. For example, if an exceptionally wet, or dry, period occurred, several 411 hydrological variables (precipitation, streamflow, reservoir storage levels, etc.) could reach
412 unprecedented levels that might change the values of the state index resulting in completely
413 erroneous impressions regarding past drought events, as well as influencing the perception of
414 future ones.

415 In regulated systems, the volume stored in the different reservoirs of the system, especially 416 the regulation reservoirs, is normally regarded as a good approximation of the actual status of 417 the whole system. Moreover, the comparison between the storage levels at the beginning and 418 the end of the hydrologic year are commonly accepted as a summary of how the management 419 of the system has been. However, the volumes stored nowadays are not comparable with the 420 volumes stored, for example, ten years ago since water uses in the system change over time. 421 This makes that the behaviour of the system, and thus the storage in reservoirs is different 422 should the new demands were considered and indicators such as the one used by river basin 423 districts in Spain cannot reflect that. In addition, the existence of high risk levels of developing 424 drought scenarios during normality situations raise concern about the need for a more 425 appropriate definition of what is considered to be normality in a water resources system. For 426 this, it is undoubtedly necessary to have a deep knowledge about the system. The use of both 427 simulation and optimisation models allow enhancing the knowledge that managers and users 428 have of the system as well as building common understanding on the needs and concerns of 429 the different actors involved.

430 Finally, following a drought preventive strategy in a water resources system needs maintaining 431 a continuous state of vigilance. Hence, drought monitoring systems should warn of the risk 432 that a certain situation, that is considered to involve risk, develops into a worse scenario 433 instead of just informing about the current state of the system. In this way, the measures 434 addressed to minimise the risk or mitigating the effects of a fully developed drought episode 435 would have enough time to operate and be efficient, and they could even be less severe than 436 when applied with urgency. Water resources systems management involves some 437 bureaucracy and it is necessary to take into account that the activation of measures normally

will take some time after the declaration of a new drought scenario. Thus, being able to
anticipate the state of the system in a way like the one presented in this work can definitely
help improving the performance of drought plans.

441 6. Conclusions

442 This paper has proposed a new methodology to support drought monitoring and scenario 443 definition in regulated water resources systems. It allows approaching droughts risk 444 assessment and early warning from a new perspective with regard to previous approaches, 445 adding value to the existing monitoring methods currently in use. The use of optimisation 446 modelling to obtain the best management of the system during uncertain hydrologic periods 447 such as droughts permits anticipating the possible outcomes of these situations without the 448 need of considering the operation rules in place that might result ineffective in these cases. An important advantage of the method developed is its capacity for dealing with complex 449 450 systems, providing a general picture of the situation in the basin while most of the previously 451 developed indices are applicable only to a demand or to a group of demands. Thus, the 452 proposed method constitutes a step forward in the definition of drought early warning systems 453 in regulated basins. The application of the methodology in the Jucar River shows its potential 454 for supporting the definition of drought scenarios and hence improving the overall drought 455 management process in the basin. Furthermore, the methodology proposed is easily 456 exportable to other cases of study since it makes use of generalized modelling tools freely 457 available online, although it is important to keep in mind that it is necessary a good knowledge 458 of the system in order it to be effective.

Since no drought is identical to another, especially given a changing climate, modelling the past to anticipate future drought is an ineffective and risky approach. Including future changes in climate and hydrology is essential, but also future water demands and operation policies must be considered in order to attain useful and reliable results for an efficient anticipation to future drought events. Different operation policies may also require different approaches with

regard to drought management, both in the definition of scenario thresholds for measuresactivation and the variables monitored, and the tools necessary to support decision making.

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